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**VINÍCIUS RIBAMAR ALENCAR MACEDO**

**ATRIBUTOS DO SOLO QUE DETERMINAM A PRODUTIVIDADE DAS  
CULTURAS PARA PREVENIR A DEGRADAÇÃO DE SOLOS  
ESTRUTURALMENTE FRÁGEIS**

**São Luís - MA**  
**2018**

VINÍCIUS RIBAMAR ALENCAR MACEDO  
Engenheiro Agrônomo

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Tese apresentada ao Programa de Pós-Graduação em Agroecologia da Universidade Estadual do Maranhão, para obtenção do título de Doutor em Agroecologia.

Orientador: Prof. Dr. Emanuel Gomes de Moura

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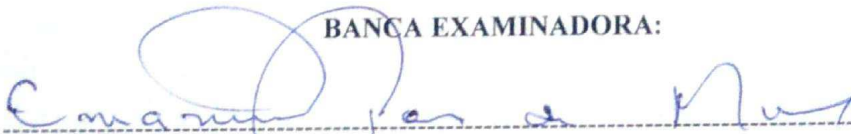
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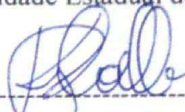
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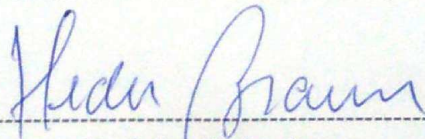
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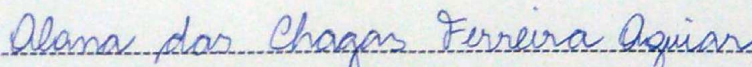
**Prof. Dr. Emanuel Gomes de Moura** (Orientador)  
Universidade Estadual do Maranhão (UEMA)



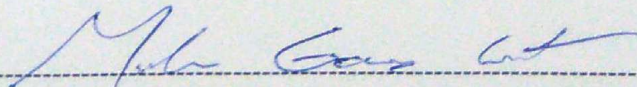
**Dra. Kátia Pereira Coelho**  
Universidade Estadual do Maranhão (UEMA)



**Prof. Dr. Heder Braun**  
Universidade Estadual do Maranhão (UEMA)



**Prof.ª Dra. Alana das Chagas Ferreira Aguiar**  
Universidade Federal do Maranhão (UFMA)



**Prof. Dr. Marlon Gomes da Costa**  
Instituto Federal do Maranhão (IFMA)

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## RESUMO

Na região centro norte maranhense, os solos têm maior tendência a degradação em função de algumas de suas características, tais como, alto nível de intemperização, decomposição acelerada da matéria orgânica do solo e predisposição a coesão. A coesão impossibilita ou dificulta o preparo do solo com aração e gradagem, uma vez que a aplicação dessas técnicas pode aumentar a compactação destes solos. Estas características dificultam o desenvolvimento da agricultura. Além disso, é predominante o uso do sistema de corte e queima para plantio de culturas agrícolas, um sistema ultrapassado que já não se sustenta, pela ausência de áreas para desbravar e contribuição com a degradação dos recursos naturais. Dada essas circunstâncias e ao aumento da preocupação da sociedade com a forma de produzir alimentos, é necessário a adoção de práticas agrícolas que aumentem a produtividade das culturas, considerando as peculiaridades de cada região, permitindo atender às necessidades do presente sem comprometer a produção alimentícia de futuras gerações. Portanto, a eficiência do uso de nutrientes é o principal fator que influencia o manejo dos agroecossistemas no trópico úmido devido à baixa capacidade de enraizamento das plantas no solo e altas taxas de perda de nutrientes. Alguns estudos desenvolvidos na região, tem sugerido o uso do sistema do plantio direto na palha de leguminosas arbóreas (cultivo em aleias) como prática de manejo. Esse sistema tem se destacado na melhoria de propriedades físicas, químicas e biológicas do solo. Tais vantagens poderiam permitir intensificar a produção pelo aumento da eficiência de uso dos nutrientes e da água com a consequente diminuição da degradação do solo na periferia amazônica. Essas vantagens precisam ser devidamente mensuradas por indicadores de qualidade do solo que antes de tudo devem ser sensíveis às variações do manejo, bem correlacionado com as funções desempenhadas pelo solo, capaz de elucidar os processos do ecossistema, ser compreensível, útil para o agricultor e preferencialmente, de fácil e barata mensuração. Baseado nestas experiências, buscou-se aprofundar o entendimento sobre os indicadores que afetam a produção do milho em solo tropical coeso.

**Palavras chaves:** Cultivo em aleias, Sustentabilidade, Intensificação



## ABSTRACT

In the north central region of Maranhao, soils are more prone to degradation due to some of their characteristics, such as high level of weathering, accelerated decomposition of soil organic matter and predisposition to cohesion. The cohesion makes it difficult or difficult to prepare the soil with plowing and harvesting, since the application of this technique can increase the compaction of these soils. These characteristics hamper the development of agriculture. In addition, it is predominant to use the slash and burning system for planting agricultural crops, an outdated system that is no longer supported by the absence of areas to clear and contribution to the degradation of natural resources. Given these circumstances and the increasing concern of society with the way of producing food, it is necessary to adopt agricultural practices that increase crop productivity, taking into account the peculiarities of each region, allowing to meet the needs of the present without compromising the food production of future generations. Therefore, the efficiency of nutrient use is the main factor that influences the management of agro-systems in the humid tropics due to the low capacity of planting in the soil and high rates of nutrient loss. Some studies developed in the region have suggested the use of no-tillage system in tree legume straw (alley cropping) as a management practice. This system has excelled in improving the physical, chemical and biological properties of the soil. Such advantages could allow to intensify the production by increasing the efficiency of use of nutrients and water with the consequent reduction of soil degradation in the Amazonian periphery. These advantages need to be properly measured by soil quality indicators that must first be sensitive to management variations, well correlated with soil functions, able to elucidate ecosystem processes, be understandable, useful to the farmer and preferably, easy and inexpensive measurement. Based on these experiments, we sought to deepen the understanding of the indicators that affect maize production in cohesive tropical soil.

**Key words:** Alley cropping, sustainability, Intensification

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## **REFERENCIAL TEÓRICO**

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### **CAPÍTULO I**

## 1 Introdução Geral

O aumento da pressão demográfica e a constante necessidade de produzir mais alimentos, tem gerado uma crescente conscientização e busca da sociedade por sistemas agrícolas que sejam sustentáveis. Há certa urgência em intensificar a agricultura e, ao mesmo tempo, conservar os recursos naturais como água e solo, bens essenciais à sobrevivência global.

Na região do trópico úmido, especificamente no centro norte maranhense, os solos têm maior tendência a degradação em função de algumas de suas características, tais como alto nível de intemperização, decomposição acelerada da matéria orgânica do solo, baixa capacidade de reter cátions e predisposição a coesão (MOURA *et al.*, 2008a). Essas características dificultam o desenvolvimento da agricultura na região.

O processo de degradação do solo é acelerado pelo sistema de corte e queima, que é baseado no uso do fogo para derrubada de florestas para implantação de pastos e culturas alimentícias. O fogo pode influenciar diretamente as propriedades do solo através de processos de aquecimento e combustão e, indiretamente, através das mudanças na sua cobertura vegetal e redistribuição do solo através da erosão acelerada pós-queimada (SANTIN e DOERR 2016).

A grande adesão ao sistema de corte e queima pela maioria dos pequenos produtores agrícolas se explicava pela rápida limpeza da área e pelo baixo custo, ou ainda porque no ano da implantação, por haver uma grande quantidade de biomassa, ocorre a disponibilização de alguns nutrientes pela cinza (FERRAZ JR, 2004). Além disso, não há um suporte de políticas públicas para fornecer a estes pequenos produtores fertilizantes químicos e implementos agrícolas.

A ausência de grandes extensões de terra, diminuição da produtividade das culturas e o tempo de pousio das áreas agrícolas insuficiente para permitir a recuperação das áreas (VARMA, 2003) gerou a necessidade de se buscar sistemas de manejo do solo mais sustentáveis e determinar quais destes sistemas e quais culturas são mais rentáveis e ocasionam menos degradação aos solos, considerando as características de cada região.

É difícil a recomendação de tecnologias no centro norte do Maranhão que permitam a sustentabilidade dos agroecossistemas em solos originados de rochas sedimentares clásticas, intemperizados e de fertilidade natural baixa (AGUIAR *et al.*,

2010, MOURA *et al.*, 2008 a). Estes solos possuem características de coesão, ou seja, estruturalmente frágeis e possuem uma tendência a perder a maioria dos poros estruturais durante o umedecimento e endurecer durante a secagem, impedindo o desenvolvimento radicular e dificultando os cultivos agrícolas (LEY *et al.*, 1995).

Além disso, a presença de altas quantidades de ferro livre e areia fina dificulta a utilização de práticas comuns em outras regiões do Brasil, a exemplo da aração e gradagem para preparo do solo, uma vez que, solos coesos, podem resultar em áreas degradadas onde a camada arável pode ser recompactada pela ação das chuvas intensas, comum no trópico úmido (MOURA *et al.*, 2009 e BUSSCHER *et al.*, 2002).

Nestas condições, a eficiência do uso de nutrientes, que é o rendimento por unidade de fertilizante aplicado é o principal fator que influencia o manejo dos agrossistemas no trópico úmido, devido à baixa capacidade de enraizamento das plantas no solo e altas taxas de perda de nutrientes (MOURA *et al.*, 2008b).

Alguns estudos desenvolvidos na região, tem sugerido o uso do sistema do plantio direto na palha de leguminosas arbóreas (cultivo em aleias) como prática de manejo. Esse sistema pode contribuir com sustentabilidade das áreas de plantio e aumento da produtividade das culturas, pois possibilita a reciclagem de nutrientes, melhora a qualidade dos indicadores físicos do solo, aumenta a capacidade de aeração, reduz o impacto causado pelas gotas de chuva e fornece moderadas quantidades de nutrientes (ADEKALU *et al.*, 2006; MOURA *et al.*, 2013; MOURA *et al.*, 2016). Tais vantagens poderiam permitir intensificar a produção e amenizar a degradação do solo na periferia amazônica.

Baseado nestas experiências, buscou-se aprofundar o entendimento sobre os indicadores que afetam a produção do milho em solo tropical estruturalmente frágil.

## 2 Referencial Teórico

### 2.1 A Agricultura na Região do Trópico Úmido

A sustentabilidade da agricultura em ambientes tropicais perpassa por inúmeras restrições decorrentes do baixo teor de nutrientes e da rápida mineralização da matéria orgânica do solo (TIESEN *et al.*, 1994; ZECH *et al.*, 1997). Ademais, a maioria dos solos tropicais atingiu um estágio avançado de intemperismo, o que evidencia que o

desgaste químico nos trópicos é bastante acelerado em comparação com a zona temperada (PASRICHA e FOX, 1993). Assim, muitos solos tropicais dependem da reciclagem de nutrientes da matéria orgânica do solo para manter a fertilidade.

No trópico úmido maranhense, região centro norte, a maioria dos solos é predominantemente arenoso na textura da superfície e predisposto a coesão, o que o torna estruturalmente instável logo após a remoção da floresta ou da pastagem para implantação dos cultivos agrícolas (MOURA *et al.*, 2008).

Entende-se como coesão uma forma particular de comportamento de solo em que este se torna muito duro quando seco, a ponto de dificultar ou mesmo impossibilitar o desenvolvimento dos vegetais, e friável quando úmido diminuindo a coesão (NORTHCOTE, 1975; EMBRAPA, 2013).

Os solos coesos têm estrutura frágil porque são constituídos basicamente por areia fina e silte (BUSSCHER *et al.*, 2002). Mullins *et al* (1987) relataram uma explicação física para o termo coeso e estabeleceram uma série de problemas agrônômicos associados, incluindo o tempo restrito de cultivos e o aumento da resistência física para o crescimento das raízes. Logo, a coesão impossibilita ou dificulta o preparo do solo com aração e gradagem, uma vez que a aplicação dessa técnica pode aumentar ainda mais a compactação destes solos.

Nessas circunstâncias, no centro norte maranhense, o método mais frequente usado na conversão dos ecossistemas naturais em áreas agrícolas é o corte seguido de queima, o qual ocasiona quebra dos ciclos biogeoquímicos, com liberação de nutrientes imobilizados na biomassa florestal e emissão de partículas e gases para a atmosfera (BONILLA, 2005). O uso demasiado desse sistema ocorre em função de custos reduzidos para preparo das áreas de plantio e pelo incremento temporário de nutrientes oriundos das cinzas, que não se mantem nos anos seguintes obrigando o agricultor a mudar de área.

Para aumentar a produtividade das culturas alimentícias e intensificar o uso agrícola dos solos sem abrir novas áreas, é preciso focar na melhoria da qualidade do solo e no aumento da eficiência do uso de nutrientes. Desse modo, a eficiência do uso dos nutrientes é o principal fator que afeta a produtividade no manejo de agroecossistemas do trópico úmido devido a menor absorção e altas perdas de nutrientes (MOURA *et al.*, 2012).

Vale ressaltar que as altas temperaturas e as precipitações intensas dificultam ainda mais o manejo do solo tropical coeso, uma vez que contribuem com os processos de intemperização dos solos tropicais.

Portanto, diferentes estratégias de manejo do solo devem ser adotadas com a finalidade de aumentar a produtividade das culturas por unidade de área e diminuir o impacto da produção de alimentos ao meio ambiente. Isso se explica pela insuficiência de terras aptas para expandir as fronteiras agrícolas e pela necessidade de redução do uso de fontes não renováveis. Espera-se também manter a fertilidade e biodiversidade dos solos, reduzir efeitos da erosão e aumentar a eficiência do uso de nutrientes (HOCHMAN *et al.*, 2011, GEHRING *et al.*, 2013).

A partir disso, alguns trabalhos desenvolvidos no trópico úmido têm demonstrado que o uso das leguminosas arbóreas, como por exemplo: *Leucaena leucocephala*, *Gliricidia sepium*, *Acacia mangium* e *Clitoria farchildiana* como cobertura do solo podem proporcionar significativas melhorias nas propriedades físicas, químicas e biológica do solo (FERRAZ, 2004; MOURA, 2013; AGUIAR *et al.*, 2010).

Moura *et al* (2015) estudaram a relação entre as exigências ecológicas da macrofauna do solo e seus impactos sobre os principais indicadores físicos e químicos na região do trópico úmido. Este estudo relatou que a aplicação de resíduos de leguminosas com alta relação C/N (*Gliricidia sepium* e *Acacia mangium*) aumentaram as funções ecossistêmicas em um sistema de plantio direto, pois favoreceu a abundância de grupos funcionais do solo, predadores e transformadores de liteira. A liteira transformada associada com o efeito da cobertura melhora vários atributos, tais como infiltração da água, porosidade do solo, densidade do solo, estoque de carbono da liteira, fração leve livre da matéria orgânica do solo (MOS) e carbono orgânico total.

Somente o uso de cobertura com biomassa de leguminosas arbóreas não garante altos níveis de nutrientes na zona radicular, o que sugere a adoção de estratégias que incluam fertilizantes minerais para aumentar a quantidade de nutrientes durante os estágios críticos da cultura (AGUIAR *et al.*, 2010). Há que se identificar, reduzir o uso ineficiente dos nutrientes e simultaneamente intensificar a produção, manter a fertilidade do solo, a biodiversidade e reduzir os impactos ambientais da agricultura (AGUIAR *et al.*, 2014).

Moura *et al.*, (2016) sugerem o uso da gessagem associada ao uso de leguminosas para solos coesos, pois melhora o enraizamento das culturas, aumenta a porosidade do solo e diminui a resistência a penetração. Este estudo avaliou a aplicação de 6 e 12 toneladas/ha de gesso, no qual se observou pequenas diferenças no conteúdo de cálcio quando se aplicou a menor dose de gesso combinado com leguminosas em comparação ao tratamento que recebeu apenas gesso, o que permite a recomendação dessa dosagem para solos tropicais arenosos com baixa capacidade de retenção de cátions.

Deste modo, é imprescindível estimular o desenvolvimento de práticas e sistemas sustentáveis de uso do solo e conservação da água por causa da degradação generalizada dos recursos ambientais (MOTAVALLI *et al.*, 2013).

Portanto, há que se considerar o impacto das mudanças de uso do solo, para poder enfim garantir que determinada prática seja viável, eficiente e sustentável. Essa avaliação é feita por meio de indicadores de qualidade do solo, que antes de tudo deve ser responsivo às variações do manejo, bem correlacionado com as funções desempenhadas pelo solo, capaz de elucidar os processos do ecossistema, ser compreensível e útil para o agricultor e, preferencialmente, de fácil e barata mensuração (DORAN e ZEISS, 2000).

## 2.2 Qualidade do Solo

A qualidade do solo é cada vez mais recomendada como integradora e avaliadora da qualidade ambiental (MONREAL *et al.*, 1998), segurança alimentar (LAL, 1999) e viabilidade econômica (HILLEL, 1991). As avaliações de qualidade do solo têm o potencial para refletir o status do solo como um recurso essencial (DORAN e ZEISS, 2000). Muitas práticas agrícolas aumentam a vulnerabilidade do solo a processos de degradação, como erosão, acidificação, salinização e declínio estrutural do solo (CARTER *et al.*, 1997; SLAVICH, 2001).

É crescente a conscientização da sociedade sobre o impacto das práticas agrícolas na qualidade do solo, o que levou a um interesse na mensuração dos efeitos de tais práticas para auxiliar no manejo e garantir o uso de práticas sustentáveis. O conceito de qualidade e saúde do solo se disseminou na década de 1990 e foi definido por vários pesquisadores (LARSON e PIERCE 1991; DORAN e PARKIN 1994; KARLEN *et al.*, 1997).



Doran e Parkin (1994) definiram a qualidade do solo como: a capacidade de um solo funcionar dentro dos limites do ecossistema, sustentar a produtividade biológica, manter a qualidade ambiental e promover a saúde de plantas e animais, enquanto outros definiram como a capacidade do solo ou aptidão para sustentar o crescimento das culturas sem resultar em degradação do solo ou prejudicar o meio ambiente (GREGORICH e ACTON, 1995).

A fim de avaliar e monitorar se as práticas agrícolas têm impacto na qualidade do solo, é necessário mensurar propriedades do solo ao longo do tempo a partir daquelas escolhidas como indicadoras e responsivas da função do solo avaliado. Logo, uma avaliação adequada da qualidade do solo deve incluir propriedades biológicas, químicas, físicas e suas interações (KARLEN *et al.*, 2003).

Há que se considerar substancialmente o papel do carbono orgânico (CO), dada a sua importância para uma variedade de propriedades do solo e processos químicos, físicos e biológicos, tais como estrutura do solo, reciclagem de nutrientes e seu papel como substrato para microorganismos do solo.

Carter *et al.*, (1997) delineou uma sequência para avaliar a qualidade do solo compreendendo funções, processos, propriedades, indicadores e metodologia. Andrews *et al.* (2004) afirmou que os indicadores apropriados para a avaliação indireta da função do solo são determinados por suas funções fundamentais em atingir os objetivos do manejo.

Essa abordagem inicialmente requer a definição de metas de manejo. Para a maioria das atividades agrícolas, a produtividade, geralmente definida pelo rendimento é um dos objetivos estabelecidos. Larson e Pierce (1991) argumentam que a qualidade do solo não deve ser limitada à produtividade, mas também ter foco na sustentabilidade, no entanto, para a continuidade das atividades agrícolas é importante que a produtividade seja um componente.

#### *Escolha de Indicadores de Qualidade do Solo*

As propriedades do solo são classificadas como inertes ou dinâmicas. As propriedades inertes, como tipo de solo e textura, pouco mudam com uso do solo ou práticas de manejo. Portanto, tem valor para a caracterização inicial do solo, mas não para

monitorar mudanças ao longo do tempo. As propriedades dinâmicas, no entanto, como o pH do solo ou o teor de C no solo, mudam em resposta às práticas de manejo.

MacEwan (2007) definiu que um indicador do solo que pode representar a condição do sistema ou sua capacidade de executar as funções do sistema. Ele definiu os atributos de um bom indicador como sensibilidade para mudança, facilidade de medição e interpretação, metodologia repetível e reversibilidade, para que tanto a melhoria como a decadência possa ser monitorada.

Tal abordagem depende de um banco de dados adequado e/ou base de conhecimento (opinião de especialistas) da faixa ideal ou limite dos indicadores escolhidos. O valor crítico de um indicador é definido como o valor acima ou abaixo do qual um determinado sistema de manejo do solo não seja mais sustentável (SYERS *et al.*, 1994). Para certas propriedades do solo os valores críticos podem estar bem definidos, como pH do solo, salinidade e toxicidade de metais, mas para outras propriedades do solo, como carbono orgânico do solo, números de minhoca e outros indicadores biológicos do solo, podem não ser bem conhecidos ou quantificados. O limite dos indicadores para o solo selecionado pode variar regionalmente e com a cultura adotada, também espacialmente e temporalmente dentro de um mesmo campo agrícola.

### 2.3 Indicadores Físicos

As modificações dos atributos físicos são mais pronunciadas quando ocorre a transição do ambiente natural para um sistema agrícola ou surge durante a implantação de sistemas conservacionistas em substituição ao sistema de preparo convencional. (ROSSETTI, *et al.*, 2013)

Um solo que fornece um meio ideal para as necessidades de produção agrícola tem forma estrutural bem desenvolvida, estabilidade a estresses de água e mecânicos e capacidade de recuperar a sua estrutura após perturbação (KAY, 1990; DEXTER, 2002). O conceito de estrutura do solo envolve a distribuição, faixa de tamanho e conectividade dos espaços porosos entre partículas e agregados no solo (OADES, 1993), que afetam o armazenamento e fornecimento de água, nutrientes e oxigênio (HAMBLIN, 1985; CASS *et al.* 2002).

CASS *et al* (2002) listaram seis propriedades que são consideradas como substitutas a condição estrutural do solo e resistência à degradação estrutural,

nomeadamente taxa de infiltração de água, capacidade de água disponível, capacidade de aeração, resistência à penetração nas mesmas condições e estabilidade de agregados. Ademais, pode ser adicionada textura, consistência do solo, densidade aparente e porosidade (LARSON e PIERCE 1991; CARTER *et al.*, 1997, GLOVER *et al.* 2000, SPARLING *et al.*, 2004; PATTISON *et al.*, 2008).

Existem alguns problemas com o uso das propriedades físicas como indicadores da qualidade do solo. Estes incluem o fato de que a medida de algumas propriedades físicas do solo seja intrinsecamente mais custosa do que a medida de atributos químicos, pois não é realizada rotineiramente por laboratórios, e geralmente requer equipamentos especializados e treinamento que tornaria difícil para os produtores realizarem as medições.

Além disso, a limitação ao crescimento das raízes pode ser devido às restrições do subsolo que podem levar considerável tempo de mudança, exigindo monitoramento durante muitos anos antes de serem detectadas as alterações. No entanto, para a medição das propriedades físicas e químicas do solo, é necessário minimizar o efeito da variabilidade espacial e temporal na ordem que os dados são interpretáveis. Assim, é necessário conduzir a medida sempre no mesmo local e sob a mesmas condições ou mais próximo possível. Apesar dessas limitações, consideramos a medição das propriedades físicas do solo listadas anteriormente em termos de adequação como indicadores da qualidade do solo.

### *Textura do Solo*

A determinação da proporção mineral de partículas de diferentes categorias de tamanho em um solo (análise do tamanho da partícula) é convencionalmente usada para quantificar a propriedade textural do solo (FAO, 1999)

A textura foi escolhida como indicador em vários estudos (LARSON e PIERCE 1991; CARTER *et al.*, 1997; GRACE e WEIER 2007; PATTISON *et al.*, 2008) e é uma importante propriedade física para caracterizar solos. Não é uma propriedade dinâmica, pois é improvável que as práticas de manejo mudem a textura do solo ao longo do tempo.

### *Estabilidade de Agregados*

Uma estrutura de superfície de solo estável é importante para promover infiltração da água e resistência à erosão (WHITE, 2010). Para que um agregado seja estável, as forças entre as partículas devem ser suficientemente fortes para evitar que a partícula se separe como resultado de forças disruptivas, como: impacto de gotas de chuva, maquinário pesado ou expansão da argila (WHITE, 2003).

A estabilidade de agregados é determinada pela reação do solo quando imerso em água e pode ser medido por peneiração úmida em laboratório ou por um teste de dispersão em campo (HAMBLIN, 1985; EMERSON, 2002). A estabilidade dos agregados é uma propriedade que é sensível às mudanças de manejo e afeta a porosidade, infiltração de água e capacidade de água disponível.

### *Consistência do solo*

A consistência do solo ou a força do solo relaciona-se com a capacidade do solo para resistir à perda de estrutura por meio da compactação, induzida por chuvas ou irrigação, etc., e para resistir à penetração por raízes de plantas e fauna de solo. Um solo com boa qualidade física deve ser forte o suficiente para manter sua estrutura e manter as plantas verticais, mas também fraco o suficiente para permitir a penetração extensiva por raízes de plantas, flora e fauna do solo (TOPP *et al.*, 1997).

A consistência do solo ou a resistência do solo através do perfil do solo é dependente do teor de umidade no momento da medição, e para uma interpretação significativa e comparação temporal de dados, deve ser medido com o mesmo teor de umidade. A força do solo pode ser medida qualitativamente pela força necessária para acrescentar agregados secos ao ar tipicamente em laboratório ou quantitativamente por um penetrômetro no campo (OLIVER *et al.*, 2013).

### *Densidade Global e Porosidade*

A densidade aparente tem sido sugerida como medida da qualidade do solo em vários estudos (LARSON e PIERCE 1991; CARTER *et al.*, 1997; GLOVER *et al.*, 2000, SPARLING *et al.*, 2004, PATTISON *et al.*, 2008). A densidade aparente, no entanto, é

uma medida insensível de como o crescimento da raiz e o movimento do ar e da água responde à estrutura do solo (KIRCHHOF e DANIELLS 2001).

A porosidade total (Pt) é a porção do solo em volume, não ocupada por sólidos. Embora a Pt determine o volume do solo que pode ser ocupado por raízes e água (PAGLIAI e VIGNOZZI, 2002), não dá uma medida do número, tamanho ou forma dos poros dentro do solo. Geralmente, os solos com maior porosidade têm maior capacidade para armazenar água, solutos, gases e calor (TOPP *et al.*, 1997, DEXTER, 2002).

Em um solo com alto conteúdo de argilas expansivas, como a montmorilonita, o volume total do solo e a porosidade mudam consideravelmente com as mudanças no teor de água (WHITE, 2003). A porosidade cheia de ar na capacidade de campo é o volume mínimo de ar disponível no estado drenado mais úmido, é sugerido um valor crítico de 10-15% do volume do solo (DEXTER, 1988; CASS *et al.*, 2002).

#### *Infiltração e Água disponível*

A Infiltração de água é um bom indicador do grau de penetração de água através de um perfil de solo. A infiltração, no entanto, é afetada por quão rapidamente a água (por chuva ou irrigação) é aplicada para o solo e o teor de água do solo na época em que a água é aplicada. A sucção matricial é o desenho da força dominante da água no solo quando estiver seca, mas a gravidade torna-se a força dominante quando o solo está molhado. A infiltração de água em solos argilosos, que se quebram quando seco, geralmente ocorre preferencialmente pelas rachaduras para que o solo se alimente dos lados dos agregados bem como da superfície (WHITE, 2009).

A distribuição da água no solo tem uma forte influência sobre distribuição das raízes no solo (LANYON *et al.*, 2004). A capacidade de água disponível (CAD) é o volume de água por unidade de profundidade do solo que está nominalmente disponível para absorção pelas plantas. Seu limite superior é definido pela capacidade de campo, e o limite inferior é o ponto de murcha permanente.

Embora a água disponível tenha sido usada por alguns autores como um indicador de qualidade do solo (LARSON e PIERCE 1991; CARTER *et al.* 1997; GUGINO *et al.* 2009; WHERRET e MURPHY 2012), um dos problemas com o uso da CAD é que a quantidade de água disponível mudará com a profundidade de enraizamento (WHITE, 2009).

Em resumo, a estabilidade de agregados e a consistência do solo (ou força) são consideradas propriedades físicas do solo potencialmente úteis a serem monitoradas ao longo do tempo como indicadoras do efeito de práticas de manejo na qualidade do solo.

## 2.4 Indicadores Químicos

A principal função do solo em relação à sua qualidade química para a produção agrícola é fornecer nutrientes para o crescimento das culturas (HEIL e SPOSITO 1997; WHITE, 2010). As principais propriedades químicas utilizadas como indicadores em sistemas agrícolas são pH, capacidade de tamponamento de pH, matéria orgânica do solo, soma de bases (SB), capacidade de troca catiônica (CTC), nitrogênio total, fósforo, potássio e vários micronutrientes.

### *pH do Solo*

O pH do solo é um dos determinantes mais importantes da fertilidade do solo através da sua influência na solubilidade de íons metálicos, como Al, Mn, Fe, Cu, Zn e Mo, seu efeito no fornecimento de nutrientes cátions e ânions, e sua influência sobre a presença e atividade de microrganismos no solo. Os solos ácidos são frequentemente associados a deficiências de nutrientes, cátions, como Ca, Mg e K, e deficiências de P (HEIL e SPOSITO, 1997). À medida que o pH do solo aumenta, os cátions  $Fe^{3+}$ ,  $Mn^{2+}$ ,  $Zn^{2+}$  e  $Cu^{2+}$ , formam hidróxidos insolúveis e ficam menos disponíveis para as plantas (WHITE, 2003).

### *Matéria Orgânica do Solo (MOS)*

A matéria orgânica do solo (MOS) não é adquirida pelas plantas como um nutriente do solo, mas o seu ciclo é importante por causa da associação com nutrientes (N, P e S) e as interações benéficas com as propriedades químicas, físicas e biológicas do solo (HOYLE *et al.*, 2011). Os nutrientes podem ser liberados no solo em formas disponíveis às plantas após a decomposição dos compostos orgânicos por microrganismos.

A proporção de cada elemento que é liberado em forma mineral varia de acordo com a composição dos compostos orgânicos decompostos (tais como C:N, C:P e os índices C:S) e as demandas da população microbiana para cada elemento (HEIL e SPOSITO 1997; WHITE, 2010). A MOS também contribui para a capacidade de troca catiônica de um solo devido à presença de grupos carboxílicos e outros grupos funcionais. A magnitude desta contribuição dependerá do pH do solo e aumentará conforme o pH do solo aumenta e dos grupos funcionais ácidos presentes.

Embora Hazelton e Murphy (2007) recomendem um mínimo 2% de MOS para manter a estrutura adequada do solo, a MOS executa uma gama de funções nos solos e o grau de importância da MOS para funções específicas difere a cada tipo de solo.

O conteúdo de MOS pode variar amplamente dentro e entre os tipos de solo devido a numerosos fatores, incluindo manejo, clima, composição mineral do solo, biota e topografia (BALDOCK e SKJEMSTAD 2001). Apesar dos problemas relacionados a variabilidade espacial e temporal das medidas e ausência de limites críticos para MOS, consideramos ser um componente essencial da avaliação da qualidade do solo por sua importância para uma ampla gama de funções do solo e medições recomendadas ao longo do tempo para monitorar o impacto das práticas de manejo na MOS.

#### *Capacidade de Troca Catiônica (CTC)*

A CTC indica a carga negativa por unidade de massa do solo e é medido como o número total de moles de carga (como  $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ,  $\text{Na}^+$  e  $\text{K}^+$ ) que podem ser deslocados por unidade de massa de solo por um solução de extração (HEIL e SPOSITO 1997; RENGASAMY e CHURCHMAN 2001; WHITE, 2003). Em solos ácidos, é necessário incluir também  $\text{Al}^{3+}$  e as vezes  $\text{Mn}^{2+}$  (RENGASAMY e CHURCHMAN 2001).

A CTC é um importante agente de controle da estabilidade da estrutura do solo, disponibilidade de nutrientes para o crescimento da planta, pH do solo e reação a fertilizantes e outros melhoradores (HAZELTON e MURPHY 2007).

## 2.5 Relação Cálcio/Matéria Orgânica do Solo

É largamente reconhecido que a MOS é o componente chave para sustentar a fertilidade do solo e conservar a qualidade ambiental (SCHNITZER, 1991; TRUMBORE, 1997). As associações de MOS com minerais, óxidos e agregados do solo foram determinadas como os principais mecanismos pelos quais a MOS está protegida da decomposição microbiana na maioria dos solos do mundo (WIESMEIER *et al.*, 2014; SIX e PAUSTIAN 2014).

Como a interação com os minerais do solo via cátions polivalentes é um dos principais meios para a estabilidade do carbono orgânico do solo (COS) e os principais cátions polivalentes são  $\text{Ca}^{2+}$  e  $\text{Mg}^{2+}$ , acredita-se que o Ca e Mg trocáveis exercem um forte controle sobre o acúmulo de COS (YANG *et al.*, 2016).

No entanto, pouca atenção tem sido dada à dinâmicas do acúmulo da MOS em relação as quantidades de cálcio (Ca). O cálcio é um nutriente essencial para as plantas, como é um cátion divalente ( $\text{Ca}^{2+}$ ) que participa de papéis estruturais na parede celular e membranas (MARSCHNER, 1995).

A deficiência de cálcio é rara na natureza, mas o excesso de cálcio restringe as comunidades vegetais em solos corrigidos com calcário (calagem). A calagem do solo e as doses de fertilizantes à base de cal são consideradas medidas tomadas para ajustar a acidez do solo e atender às necessidades de cálcio das plantas. (KOLAR *et al.*, 2007).

A interação entre as mudanças na carga de matéria orgânica com alterações no pH e aumento da ligação de cátions polivalentes a altas concentrações de Ca, pode possibilitar as menores taxas de processamento de MOS em solos manejados a altas concentrações de Ca e pH (WHITTINGHILL e HOBBIE 2012).

As variações do pH influenciam a atividade microbiana do solo e o acúmulo de MOS, e variação no Ca trocável entre as idades da paisagem pode explicar melhor os padrões locais da atividade microbiana do que a variação do pH. Desse modo espera-se menos atividade microbiana em solos manejados com maiores concentrações de Ca devido à estabilização da matéria orgânica por ponte de cátions polivalentes, reduzindo o acesso à matéria orgânica por microorganismos (WHITTINGHILL e HOBBIE 2012).



## 2.6 Degradação do Solo

O solo está sob crescente ameaça de uma ampla gama de atividades antrópicas que contribuem com a degradação e prejuízo a longo prazo à sua disponibilidade e viabilidade de uso. Durante as últimas cinco décadas, as atividades agrícolas foram diretamente afetadas por crescimento populacional e distribuição espacial das pessoas. Mudanças no uso da terra e atividades agrícolas intensivas têm efeitos negativos sobre o valor da produção e capacidade produtiva dos solos agrícolas (ZINCK *et al.* 2004; EMADODIN *et al.* 2009; EMADODIN e BORK 2011).

A agricultura intensiva tende a acentuar a erosão do solo e a perda de matéria orgânica do solo, o que representa uma ameaça para a sustentabilidade da agricultura a longo prazo, sobretudo em eventos climáticos extremos, como a seca. É imperativo a urgência em melhorar a eficiência geral dos sistemas agrícolas e reduzir a pressão exercida sobre o meio ambiente e sobre o solo (GOMIERO, 2013).

No passado, o aumento da demanda mundial de alimentos foi atingido pelo aumento das áreas agrícolas, melhoria genética das plantas e intensificação do uso de insumos (por exemplo, fertilizantes e pesticidas), porém no futuro, o aumento da intensificação agrícola pode resultar em mais degradação ambiental, caso práticas agrícolas mais sustentáveis não forem amplamente adotadas (TILMAN *et al.*, 2002).

As práticas agrícolas tradicionais diminuíram a produtividade do solo, na medida em que muitos solos agrícolas são esgotados de nutrientes e incapazes de sustentar naturalmente a produtividade das culturas. Nas próximas décadas, um desafio crucial para a agricultura será atender às demandas de alimentos sem prejudicar ainda mais o meio ambiente. Aumentar a produtividade e os retornos econômicos para a agricultura de forma sustentável é um desafio central para atingir os objetivos globais de redução da pobreza e manejo ambiental (FAO, 2012).

Um terço dos solos agrícolas do mundo, ou aproximadamente 2 bilhões de hectares de terra são afetados pela degradação do solo. No sentido geral, a degradação do solo pode ser descrita como a deterioração da qualidade do solo, ou seja, perda parcial ou total de uma ou mais funções do solo (RITSEMA *et al.*, 2005).

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**Identifying indicators in a structurally fragile tropical soil to recommend management practices and prevent land degradation in the Amazonian periphery**

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**CAPÍTULO II**

1 Identifying indicators in a structurally fragile tropical soil to recommend  
2 management practices and prevent land degradation in the Amazonian  
3 periphery

4

5 Vinicius Ribamar Alencar Macedo<sup>a</sup>, Emanuel Gomes de Moura<sup>a</sup>

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7 <sup>a</sup>Agroecology Program of Maranhão State University, Cidade Universitária  
8 Paulo VI, Lourenço Vieira da Silva, Avenue, 65025-970 São Luís – Maranhão,  
9 Brazil.

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11 ABSTRACT

12 In the Amazonian periphery, farmers need to intensify their productivity and be  
13 assured of the sustainability of the areas designated for crops, to prevent illegal  
14 deforestation of new areas. The aim of this study was to identify the main  
15 indicators that affect crop productivity in a structurally fragile Amazonian soil,  
16 and to determine the importance of maize yield indicator and their relationships  
17 with promising management practices to intensify crop production in  
18 agricultural areas in a sustainable way. The experiment was established with  
19 three leguminous species, *Clitoria fairchildiana*, *Acacia mangium* and *Leucaena*  
20 *leucocephala*, and an area without leguminous plants. Maize was sown between  
21 the rows of leguminous plants. Chemical and physical soil properties and maize  
22 grain yield were determined. Multiple linear regression, hierarchical partitioning

23 analysis and spatial variability were used. Our results suggest the possibility of  
24 intensifying crop production in tropical soil prone to hardsetting through soil  
25 management practices, because Calcium (Ca) and soil organic matter (SOM)  
26 contents were highly correlated and exerted strong influence on maize grain  
27 yield (accounted for 30 up to 50 % of maize yield in two years). Other indicators  
28 like macroporosity, soil density and resistance to penetration, which also  
29 influenced the yield, were correlated with Ca and SOM, both of which can be  
30 increased or improved by agronomic practices. It is worth highlighting that due  
31 to negative influence of pH on total independent contributions to increased  
32 organic matter, higher calcium contents must be achieved by gypsum application  
33 rather than using liming material.

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35 **KEY WORDS:** calcium; organic matter; grain yield; alley cropping; gypsum

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44 INTRODUCTION

45 Land degradation implies long-term decline in soil productivity and reduction  
46 in soil capacity to perform specific environmental functions of values to human  
47 welfare (Aweke *et al.*, 2015). The growing public interest in environmental  
48 sustainability has been identified as one of the most important drivers for  
49 modern soil science. Consequently, there is an increased demand to determine  
50 effects of soil management practices on land degradation and for the  
51 sustainability of agroecosystems (Schoenholtz *et al.*, 2000). This demand has led  
52 to a greater need to understand and evaluate soil quality indicators that can  
53 support the prevention, recognition, monitoring and rehabilitation of degraded  
54 terrestrial environments (Viana *et al.*, 2014). However, as environments differ, as  
55 well as soil functions of interest, there is no methodology to characterize soil  
56 quality based on a universal set of indicators (Bouma, 2002), as there is also no  
57 universal list of indicator properties suitable for all regions and ecosystem  
58 functions (Oliver *et al.*, 2013). However, a comprehensive analysis of soil will not  
59 accurately describe soil quality unless the indicators are chosen with a specific  
60 soil function in mind, within a defined ecosystem (Viana *et al.*, 2014).

61 In tropical regions, soil derived from clastic sedimentary rocks with low  
62 content of aggregator elements, like calcium, elemental iron and organic carbon,  
63 are widespread (Daniells, 2012). Unsuitable management of this soil, which is  
64 structurally fragile and prone to hardsetting results in unfavourable conditions

65 for sustainable agrosystems because it affects soil rootability, water and nutrient  
66 use efficiency (Moura *et al.*, 2016a). Together these factors decrease biomass  
67 productivity, soil organic matter, and increase land degradation risks. In  
68 addition, under tropical meteorological conditions, due to the high atmospheric  
69 evaporative demand, plants can lose turgidity, stop growing and decrease  
70 productivity of biomass even when soil moisture supply might be considered  
71 suitable for other regions (Denmead & Shaw, 1962). In these circumstances, due  
72 to negative interactions between the reduced soil rootability and atmospheric  
73 conditions the depth and fertility of the rootable layer, as well as, the soil physical  
74 indicators are crucial for plants to reach their full productivity potential, as well  
75 as for avoiding land degradation and achieving sustainable agrosystems (Moura  
76 *et al.*, 2013). Given the environmental condition of the humid tropics some  
77 authors (Aguiar *et al.*, 2010; Moura *et al.*, 2016b) have recommended taking  
78 advantage of the rapid growth of leguminous trees in alley cropping systems. In  
79 the selection of hedgerow species for alley cropping must be consider nitrogen-  
80 fixing capacity and biomass production, in order to ensure maximum crop  
81 benefit of the nutrients added through the tree prunings. These qualities are  
82 linked to relative potential of three leguminous shrubs to changes on some soil  
83 properties, through increased soil organic matter and nutrients availability.

84 In the Amazon, including periphery regions, it is estimated that there is  
85 around 170 thousand km<sup>2</sup> of deforested and underutilized land, often in a state

86 of degradation (Galvão *et al.*, 2016). In these circumstances, unsuitable soil use is  
87 one of the greatest threats to natural reserve areas and to maintaining a healthy  
88 ecosystem, mainly because rainforest areas are almost always considered by  
89 farmers to be favourable to support future pasture or cropping, when the  
90 depleted area in use can no longer do so (Moura *et al.*, 2016b). Yet, this process is  
91 now increasingly recognized as a fundamental cause of deforestation, land  
92 degradation and increased greenhouse gas emissions in the Amazon region and  
93 its surroundings (Aguiar *et al.*, 2011). Thus, is now becoming increasingly  
94 understood that linkages exist between land degradation, biodiversity and  
95 climate change. This leads to justifying land degradation control through local  
96 suitable practices as a legitimate global change topic rather than a solely local or  
97 domestic topic (Gisladottir & Stocking, 2005).

98       Therefore, to preserve the rainforest and maintain a healthy Amazonian  
99 ecosystem it is essential to provide information for understanding and managing  
100 land according to soil resilience and performance through indicators that reflect  
101 ecosystem processes and integrate physical, chemical, and biological properties.

102       Taking this into account, the aim of study was to identify the main  
103 indicators that affect maize productivity in a structurally fragile tropical soil from  
104 the Amazonian periphery. After this, we seek to determine the importance of  
105 each indicator and their relationships with promising management practices to  
106 maintain crop production and avoid degradation in agricultural areas.

## 107 MATERIALS AND METHODS

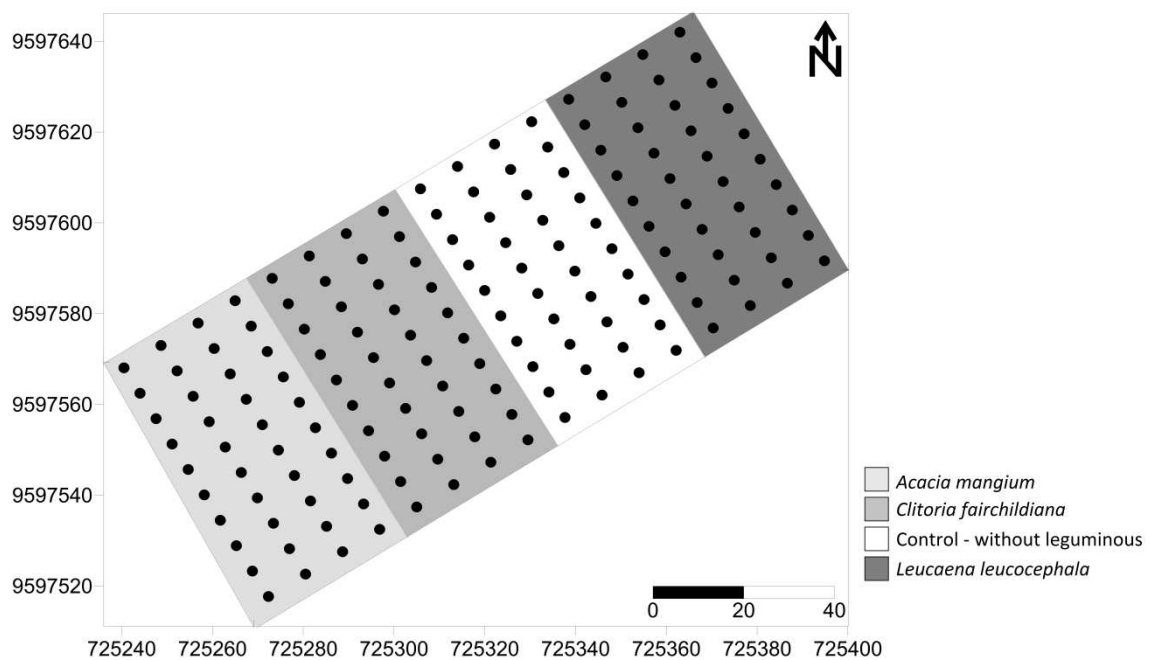
### 108 *Area description*

109           The experiment was conducted in Maranhão state, Brazil (3° 38' S, 42°  
110 58'W), in a plain with an almost flat landform, with great homogeneity and  
111 preferred by the farmers for cultivating maize and soybean. Climatologically, the  
112 region has a hot and semi-humid equatorial climate with two well-defined  
113 seasons: a rainy season that extends from January to June and a dry season with  
114 a marked water deficit from July to December. The annual average temperature  
115 is approximately 27 °C; the maximum temperature is 37 °C, and the minimum  
116 temperature is 23 °C. The annual mean rainfall during the experimental period  
117 was 1,439 mm year<sup>-1</sup>.

118           The local soils display hardsetting characteristics determined by  
119 relationship between penetration strength and volumetric water contents (Moura  
120 et al., 2012) and were classified as Arenic Hapludults (Soil Survey Staff, 2010).  
121 Before establishing the experimental area in December, 2010, chemical and  
122 physical properties of soil were determined, and are given as follows: pH 4.0 (in  
123 CaCl<sub>2</sub>); 20 g kg<sup>-1</sup> of organic-C; 15 mg dm<sup>-3</sup> of P; 25 mmol<sub>c</sub> dm<sup>-3</sup> of (Al + H); 15  
124 mmol<sub>c</sub> dm<sup>-3</sup> of Ca; 9 mmol<sub>c</sub> dm<sup>-3</sup> of Mg; 1 mmol<sub>c</sub> dm<sup>-3</sup> of K; 50 mmol<sub>c</sub> dm<sup>-3</sup> of CEC;  
125 50.0% of percentage base saturation; 300 g kg<sup>-1</sup> of coarse sand, 545 g kg<sup>-1</sup> of fine  
126 sand, 61 g kg<sup>-1</sup> of silt; 90 g kg<sup>-1</sup> of clay. In January 2011, the area of 11,900 m<sup>2</sup>, (which  
127 was fallow since 2008 after rice cultivation) was limed with 2 Mg ha<sup>-1</sup> of surface-



128 applied lime, corresponding to 780 and 260 kg ha<sup>-1</sup> of Ca and Mg, respectively.  
129 The experiment was established in 2011 using an alley cropping system. The *area*  
130 *was divided* into four equal lots of 42.5 x 70 m and leguminous trees were planted  
131 at a spacing of 0.5 m between plants and 2.5 m between rows. Three species were  
132 used: sombrero (*Clitoria fairchildiana*), acacia (*Acacia mangium*), and leucaena  
133 (*Leucaena leucocephala*). One-fourth of the area was maintained without  
134 leguminous plants. In order to measure soil physical parameters and maize  
135 productivity, 160 sampling points were delimited in a uniform grid of 10 x 7 m.  
136 The coordinates of each sample point were recorded using a GPS device (Figure  
137 1).



138  
139 Figure 1. Experimental field design.

140 In 2011 and 2012, maize (*Zea mays* L.) was grown during the rainy season  
141 between the rows of leguminous plants and in the control area with a spacing of

142 90 cm between the rows and 33 cm between the plants. Aiming at this  
 143 experiment, in February of 2013 and 2014, at the beginning of the rainy season,  
 144 three rows of maize (*Zea mays L.*) cultivar AG 7088 spaced 80 cm apart were sown  
 145 between the rows of leguminous. In all the years the soil was fertilized with 50  
 146 kg ha<sup>-1</sup> N, 120 kg ha<sup>-1</sup> P<sub>2</sub>O<sub>5</sub>, 100 kg ha<sup>-1</sup> K<sub>2</sub>O and 5 kg ha<sup>-1</sup> Zn. A total of kg ha<sup>-1</sup> of  
 147 N, supplied as urea, was *applied at the V3 maize growth stage, in all years*. The maize  
 148 was fertilized using 300 kg ha<sup>-1</sup> of formula 10-25-15 + 0.05 Zn. A total of 50 kg ha<sup>-1</sup>  
 149 of N.

150 The first pruning of leguminous plants was performed in 2012 and  
 151 continued in the following years. The annual pruning was 0.5 m above ground  
 152 level and the green biomass of the leguminous was evenly distributed between  
 153 the maize rows, just after planting. The amount of residues applied after  
 154 implementing the alley cropping system is shown in Table I.

155 Table I. Applied dry matter of legumes in the experimental field.

Legumes	Dry matter (Mg ha <sup>-1</sup> )		
	2012	2013	2014
<i>Leucaena leucocephala</i>	0.73	0.85	3.15
<i>Clitoria fairchildiana</i>	3.81	5.23	7.94
<i>Acacia mangium</i>	2.90	3.69	5.21

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159 *Soil physical analyses*

160           The soil was sampled using 100 cm<sup>3</sup> rings in May 2013 in order to  
161 determine, soil dry bulk density, total porosity ( $\phi_t$ ), macro-porosity (non  
162 capillary porous) and microporosity (capillary porous). Two replicates were  
163 collected per point from a depth of 8-12 cm (in the middle of the 0 -20 cm layer).  
164 Samples were saturated, weighed, placed on a tension table and equilibrated at -  
165 6 kPa. After weighing, each sample was oven dried at 105 °C. Soil dry bulk  
166 density ( $\rho_b$ ) was calculated as  $m/v$ , where  $m$  is the dry collection soil mass at 105  
167 °C and  $v$  is the ring volume. Total porosity ( $\phi_t$ ) was calculated as  $[1 - (\rho_b/\rho_p)]$ ,  
168 where  $\rho_b$  is the soil density (Mg m<sup>3</sup>) and  $\rho_p$  is the particle density (Mg m<sup>3</sup>), which  
169 was determined according to EMBRAPA (1997). We calculated the  
170 macroporosity corresponding to the volume of pores with equivalent diameter  
171 larger than 50  $\mu\text{m}$  as the mass difference between saturated sample and  
172 equilibrated sample at -6 kPa (Thomasson, 1978). Microporosity was determined  
173 by the difference between the macroporosity and total porosity.

174           The soil penetration strength (SPS) was measured with 5-cm gradations,  
175 at depths of 0-5 cm, 6-10 cm, 11-15 cm, and 16-20 cm, using a digital penetrometer  
176 (Falker, Porto Alegre, Brazil) with three replicates per point. Furthermore, the  
177 soil moisture was measured at depths 0 to 20 cm, with two replicates per point  
178 using a TDR - TRIME-FM instrument (IMKO, Ettlingen, Germany). Soil particle

179 size distribution was measured in the same soil samples taken for chemical  
180 analyses.

181 *Soil chemical analyses and maize grain yield*

182 Soil samples were taken for chemical analysis in May 2013 at depths of 0-20cm;  
183 six replicates were collected using a Dutch auger. The samples from each point  
184 were passed through a 2-mm sieve and then air-dried prior to the analyses. Soil  
185 samples were analysed for C, pH, P and exchangeable  $K^+$ ,  $Ca^{2+}$ ,  $Mg^{2+}$  and ( $H^+$  +  
186  $Al^{3+}$ ). At physiological maturity, the maize yield was determined in two 10-m<sup>2</sup>  
187 areas, within each 70-m<sup>2</sup> grid. Table II shows the analytical method for each  
188 chemical and physical variable in this study.

189 Table II. Analytical method for each physical and chemical variable.

Variable	Unit	Method	References
SOM	$\text{g dm}^{-3}$	Dichromate oxidation	Walkley & Black, 1934
P, Ca, Mg, K	$\text{mmolc dm}^{-3}$	Ion-exchange resin procedure	Raij <i>et al.</i> , 1986
H+AL	$\text{mmolc dm}^{-3}$	Solution SMP	Quaggio <i>et al.</i> , 1985
pH	-	pH in $\text{CaCl}_2$	Peech, 1965
$\rho_s$	$\text{Mg m}^{-3}$	Rings of $100 \text{ cm}^3$	Thomasson, 1978
$\varphi_t$	$\text{m}^3 \text{ m}^{-3}$	From the values of dry bulk density	Aikins & Afuakwa, 2012
Clay	$\text{g kg}^{-1}$	Pipette method	Gee & Bauder, 1986
Sand	$\text{g kg}^{-1}$	Pipette method	Gee & Bauder, 1986
Silt	$\text{g kg}^{-1}$	Pipette method	Gee & Bauder, 1986
SPS	MPa	By manual action	ASAE, 1999
MPS	MPa	By manual action	ASAE, 1999
Mac	$\text{m}^3 \text{ m}^{-3}$	non capillary porous	Danielson & Sutherland, 1986
Mic	$\text{m}^3 \text{ m}^{-3}$	capillary porous	Danielson & Sutherland, 1986
Soil moisture	%	Time domain reflectometry (TDR)	Topp <i>et al.</i> , 1980

190 SOM – Soil organic matter,  $\rho_s$  – Bulk density;  $\varphi_t$  - Total porosity; SPS – Soil penetration strength; MPS - Maximum penetration

191 strength; Mac - Macroporosity; Mic – Microporosity

192 *Statistical Analysis*

193 The data were subjected to the homogeneity of variance (Levene's test) and  
194 normality (Shapiro-Wilk) (Table III). Multiple linear regression (MLR) was used  
195 to identify the best predictors (physical and chemical soil properties) for yield of  
196 maize ("MASS" package version 7.3-45, (Venables & Ripley, 2002), implemented  
197 in the R statistical program (R Development Core Team, 2009). The selection of  
198 predictors was carried out by a stepwise backward selection including all of the  
199 variables, except to those eliminated by test of multicollinearity. The stepping  
200 criteria employed for entry and removal was based on the Akaike information  
201 criterion, AIC (Akaike, 1973). A standard regression coefficient ( $\beta$ ) was used to  
202 represent the amount of the response variable changed (yield of maize) when the  
203 explanatory variable (physical and chemical soil properties) changes 1 unit. MLR  
204 was also used to identify the best predictors for the soil organic matter (SOM).

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214 Table III. Descriptive statistics of physical and chemical soil properties, and two  
 215 years of yield of maize.

Variables	Mean	Min	Max	SD	CV
SOM, g dm <sup>-3</sup>	26.91	16.45	37.10	4.39	16.31
Ca, mmol <sub>c</sub> dm <sup>-3</sup>	16.07	7.50	33.50	5.27	32.82
K, mmol <sub>c</sub> dm <sup>-3</sup>	0.88	0.53	1.33	0.16	17.64
pH (CaCl <sub>2</sub> )	4.43	4.04	4.90	0.19	4.26
ρ <sub>s</sub> , g cm <sup>-3</sup>	1.40	1.31	1.52	0.05	3.26
Clay, g 100g <sup>-1</sup>	17.38	11.50	23.50	2.60	14.96
Silt, g 100g <sup>-1</sup>	8.33	4.90	17.00	2.18	26.13
PS (6-10cm), Mpa	1.01	0.41	1.98	0.34	34.00
PS (16-20cm), MPa	1.54	0.90	2.40	0.31	19.85
MPS, Mpa	1.74	1.11	2.94	0.38	21.85
Mac, m <sup>3</sup> m <sup>-3</sup>	0.21	0.17	0.27	0.02	9.90
Mic, m <sup>3</sup> m <sup>-3</sup>	0.20	0.15	0.27	0.02	11.64
Yield <sup>1</sup> , Mg ha <sup>-1</sup>	4.35	2.40	6.49	0.71	16.40
Yield <sup>2</sup> , Mg ha <sup>-1</sup>	4.33	2.95	5.95	0.58	13.47

216 Min – Minimum value; Max – Maximum value; SD – Standard Deviation; CV –  
 217 Coefficient of Variation in %.

218 SOM – Soil organic matter; ρ<sub>s</sub> – Bulk Density; PS – Penetration Strength; MPS –  
 219 Maximum penetration strength; Mac – Macroporosity; Mic – Microporosity;  
 220 1Yield from 2013; 2Yield from 2014.

221 Next, a Hierarchical Partitioning analysis (HP) was performed (“hier.part”  
 222 packge version 1.0-4 (Walsh & Mac Nally, 2015), done in the R statistical program  
 223 (R Development Core Team, 2009), to estimate the percentage of total variation  
 224 in a specified variable, yield of maize or SOM, explained by a given independent  
 225 variable (Mac Nally, 2000; Mac Nally & Wash, 2004). Only the variables selected

226 by MLR were used in the HP. The HP considers all the possible models in an  
227 MLR to determine the independent contribution of each predictor variable to the  
228 total variability of the response variable (Mac Nally, 2000). Statistical significance  
229 of the independent contributions of each independent variable was determined  
230 using a randomization approach with 1,000 interactions (95% confidence limit)  
231 (Mac Nally, 2000).

232 The spatial variability was analyzed by constructing adjusted  
233 semivariograms in a mathematical model according to the following parameters:  
234 nugget effect ( $C_0$ ), sill ( $C_0 + C_1$ ), and range ( $a$ ). The fit of the semivariogram models  
235 to the experimental data was based on the Root Mean Square Error (RMSE)  
236 between estimated and experimental values. The ratio between nugget semi-  
237 variance and total semi-variance or sill was used to define different classes of  
238 spatial dependence. If the ratio was 25%, the variable was considered to be  
239 strongly spatially dependent, or strongly distributed in patches. The  
240 semivariograms and the maps were generated using kriging with a GS +  
241 geostatistical package Software (Gamma Design de Software, 2004).

## 242 RESULTS

243 The multiple regression model showed that there were differences as for the  
244 more significant indicators to changes in maize production, between 2013 and  
245 2014. However, there was a prevalence of indicators related to physical soil  
246 conditions in both years (Table IV). In 2013, SOM and soil density ( $p < 0.001$ ) was  
247 highly significant, while the clay content ( $p < 0.016$ ) and SPS in the 16 – 20 cm



248 layer ( $p < 0.046$ ) were less significant. In 2014, macroporosity ( $p < 0.001$ ), calcium  
 249 ( $p < 0.002$ ), SOM ( $p < 0.008$ ) and microporosity ( $p < 0.034$ ) were the most significant  
 250 indicators for grain yield. The proportion of variability accounted for by a  
 251 statistical model, or *coefficient of determination* ( $R^2$ ) was 0.30 in 2013 and 0.31 in  
 252 2014.

253 Table IV. Multiple regression with maize yield as dependent variable, in 2013  
 254 and 2014.

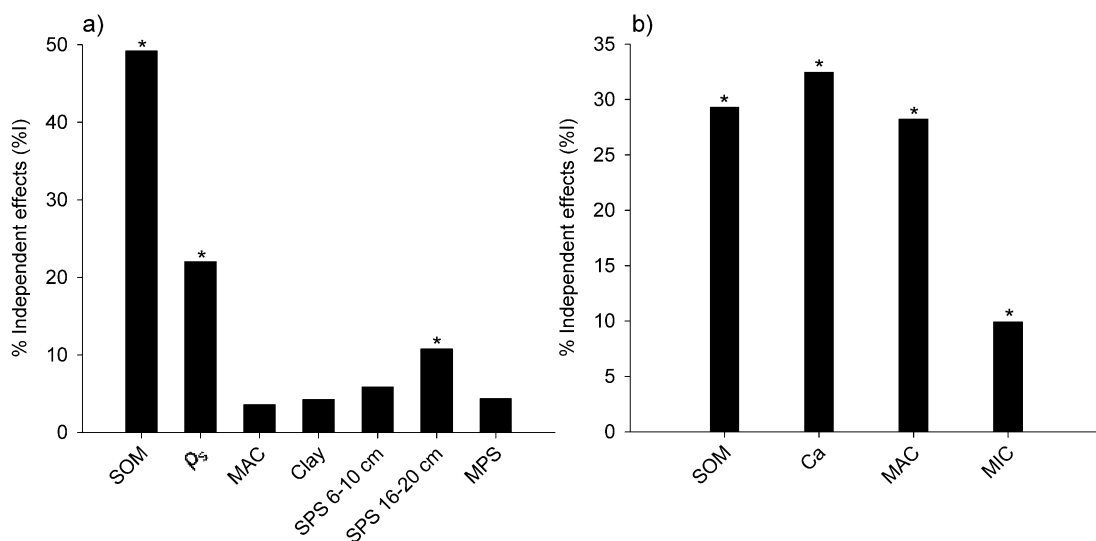
	B	SE	$\beta$	$p$
2013				
SOM	0.132	0.024	0.470	<0.001
qs	-5.362	1.268	-0.346	<0.001
Clay	-0.052	0.022	-0.191	0.016
SPS (16-20cm)	-0.734	0.333	-0.314	0.029
SPS (6-10cm)	0.405	0.201	0.195	0.046
MPS	0.466	0.266	0.248	0.082
Mac	-4.792	2.843	-0.140	0.094
2014				
Mac	-9.054	1.907	-0.324	<0.001
Ca	0.041	0.013	0.372	0.002
SOM	0.067	0.025	0.295	0.008
Mic	-5.427	2.539	-0.219	0.034

255 B – Regression coefficient; SE – Standard error;  $\beta$  – Standardised beta coefficient;  $p$  –  
 256 value.

257 SOM – Soil organic matter, qs – Bulk density;  $\phi_t$  - Total porosity; SPS – Soil penetration  
 258 strength; MPS - Maximum penetration strength; Mac - Macroporosity; Mic –  
 259 Microporosity.

260 For complementing multiple regression analysis and identifying the most  
 261 likely causal factors while alleviating multicollinearity problems, HP, an

262 analytical method of multiple regression was used. In 2013, estimates of total  
 263 contribution by HP showed that the highest total contributions to grain yield  
 264 were provided by SOM (50%), bulk density (22%) and SPS in the 16 – 20 cm layer  
 265 (11 %) (Figure 2a, 2b), while in 2014 calcium, SOM and macroporosity, all around  
 266 30%, were indicators with the highest total independent contribution.  
 267 Microporosity appears with 10% of total independent contribution to grain yield.



268

269 Figure 2. Distribution of the percentage of independent effect (%I) of each  
 270 variable (physical and chemical soil properties) on the maize grain yield in 2013  
 271 (a) and 2014 (b). Significant predictors are indicated by asterisks ( $p < 0.05$ ). SOM –  
 272 Soil organic matter,  $\rho_s$  – Bulk density;  $\rho_t$  - Total porosity; SPS – Soil penetration  
 273 strength; MPS - Maximum penetration strength; Mac - Macroporosity; Mic –  
 274 Microporosity

275

276 Results of both 2013 and 2014 showed that SOM contents might be even  
 277 more important to explain the variation in maize grain yields. In addition, due to  
 278 its relation with other physical soil properties, a multiple regression model was  
 279 used to select predictor variables for SOM. The model showed that calcium ( $p <$

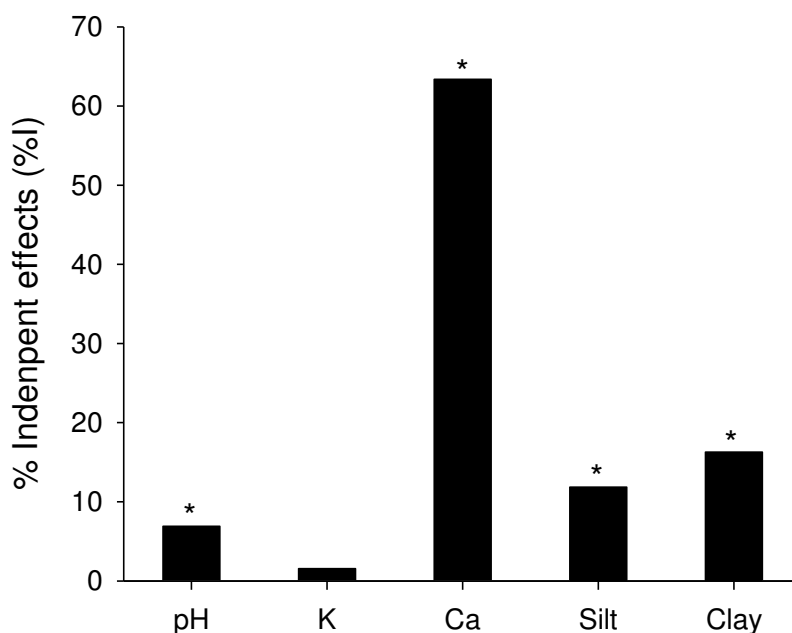
280 0.001), clay ( $p < 0.001$ ), silt ( $p < 0.002$ ), K ( $p < 0.008$ ) and pH ( $p < 0.014$ ) were  
 281 significant (Table V).

282 Table V. Multiple regression of physical and chemical soil properties with  
 283 soil organic matter.

	B	SE	$\beta$	$p$
Ca	0.521	0.048	0.626	<0.001
Clay	0.385	0.102	0.228	<0.001
Silt	0.383	0.121	0.190	0.002
K	3.488	1.307	0.123	0.008
pH	-2.973	1.191	-0.128	0.014

284 B – Regression coefficient; SE - Standard Error;  $\beta$  - Standardised beta  
 285 coefficient;  $p$  - value.  
 286

287 The proportion of variability accounted for by a statistical model, or  
 288 *coefficient of determination* ( $R^2$ ) was 0.69. Meanwhile, estimates of total  
 289 independent contribution by HP showed that the highest total independent  
 290 contributions to SOM variation were given by calcium (63 %), clay (16 %), silt (12  
 291 %), and pH (7 %) (Figure 3).



292

293 Figure 3. Distribution of the percentage of independent effect (%I) of each  
 294 variable (physical and chemical soil properties) on the soil organic matter.  
 295 Significant predictors are indicated by asterisks ( $p < 0.05$ ).

296

297 Spatial analysis of data showed that the fields had significant spatial  
 298 variability in their indicators, which exhibit some degree of positive spatial auto-  
 299 correlation. Geostatistical methods are often suitable for analysis of properties  
 300 that show spatially correlated behaviour. Semivariograms were computed for  
 301 calcium, SOM and yield, and parameters for the best fitting spherical models  
 302 were estimated (Table VI). The semivariogram for SOM contents showed a  
 303 higher nugget effect and sill than calcium contents and yields. The nugget  
 304 parameter of the semivariogram is a measure of unexplained variability. The  
 305 percentage of unexplained variability can be estimated from the ratio of the  
 306 nugget to the sill. Approximately 18% and 23% of the variation in measured  
 307 calcium and SOM, respectively, at the field site was unexplained. This

308 unexplained variability could be due to variability at scales of sampling less than  
 309 the 7 m interval used for this study or to variability resulting from measurement  
 310 errors. The range of influence for calcium and SOM measured was 70 m and 69  
 311 m, respectively. The range (A) is a measure of the maximum distance over which  
 312 properties remain spatially correlated. At distances shorter than the range the  
 313 pairwise sample variation depends upon the distance of separation between  
 314 them. The values of DSD indicate the degree of spatial dependence. DSD of 0 to  
 315 25% indicate a strong spatial dependence, 25 to 75% indicate a moderate spatial  
 316 dependence, and 75 to 100% corresponds to a weak spatial dependence  
 317 (Cambardella et al., 1994).

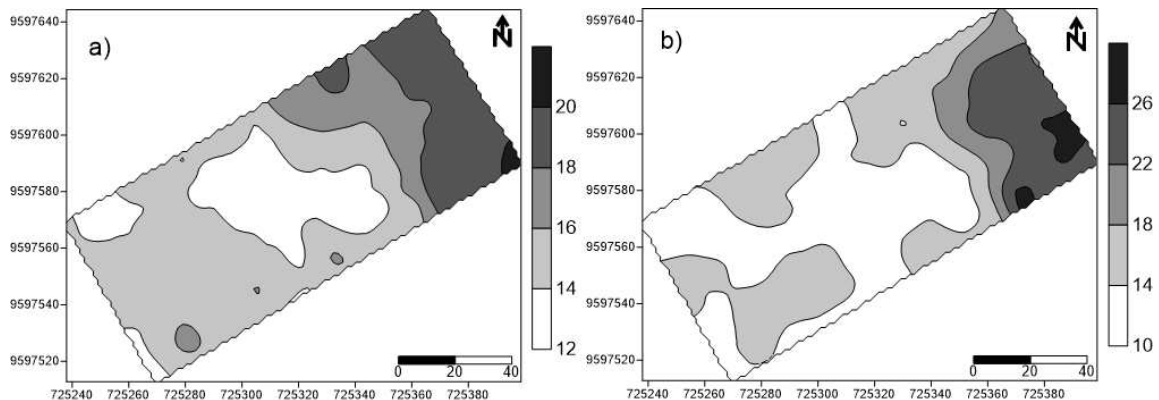
318 Table VI. Parameters of semivariogram adjusted for Ca, SOM, maize grain  
 319 yield in 2013 and 2014.

Variable	Model	C <sub>0</sub>	C <sub>0</sub> + C	a (m)	DSD(%)	r <sup>2</sup>	CV
Ca	Spherical	4.82	26.2	70	18	0.94	0.97
SOM	Spherical	1.54	6.70	69	23	0.93	0.95
Yield 2013	Spherical	0.25	0.65	67	38	0.89	0.95
Yield 2014	Spherical	0.12	0.40	48	31	0.81	0.87

320 C<sub>0</sub> – nugget effect; C<sub>0</sub>+ C – sill; a – range; DSD - Degree of spatial  
 321 dependence [ $C_0 * 100 / (C + C_0)$ ]; CV – Cross validation; SOM – Soil organic matter.  
 322

323 A positive spatial auto-correlation, as shown by semivariograms here,  
 324 indicates that similar attributes are grouped together spatially. Thus, if spatial  
 325 dependence is detected, the modelled semivariograms can then be used to map

326 the interested variable by kriging, an interpolation method that produces  
327 unbiased estimates with minimal estimation variance. The maps obtained  
328 through kriging showed significant predictor variables for yield that could be  
329 influenced by management practice as SOM and calcium are spatially associated  
330 and present low values in most of the area (Figure 4a, 4b). Thus, around 98% of  
331 the area showed low or very low calcium contents ( $<26 \text{ mmol} \cdot \text{dm}^{-3}$ ), while SOM  
332 was very low ( $< 20 \text{ g dm}^{-3}$ ) in almost 99% of the sampled area.



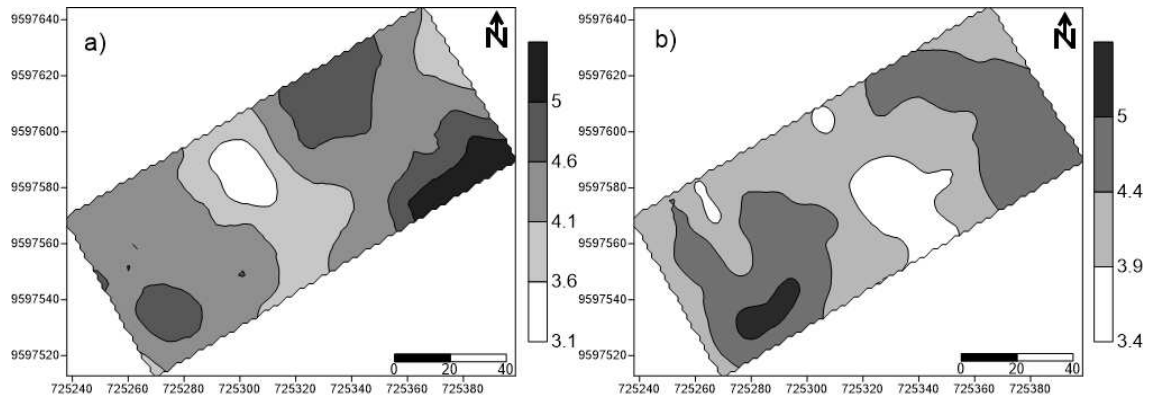
333 Figure 4. Spatial distribution of (a) Soil organic matter and (b) calcium in the  
334 experimental field.  
335

336 In the same way, the range of maize grain yield variation in 2013 (2.40 to  
337  $6.49 \text{ Mg ha}^{-1}$ ) and 2014 (2.95 to  $5.95 \text{ Mg ha}^{-1}$ ) also showed low productivity values  
338 (Figure 5a, 5b). Thus, in around 96% of the area, the productivity was lower than  
339  $5 \text{ Mg ha}^{-1}$  in 2013. In 2014 the area with productivity lower than  $5 \text{ Mg ha}^{-1}$   
340 increased to 98%.

341

342

343



344

345 Figure 5. Spatial distribution of maize grain yield 2013 (a) and 2014 (b) in the  
 346 experimental field.

347

348 DISCUSSION

349 There was a clear prevalence of the indicators related to physical soil conditions  
 350 to maize production. Indeed, in structurally fragile tropical soil, physical  
 351 properties exert a dominant influence on soil management and plant  
 352 productivity because hardsetting during soil drying has serious implications for  
 353 root growth due to increasing penetration resistance, and decreasing water and  
 354 nutrient uptake (Fabiola *et al.*, 2003). Furthermore, differences for significant  
 355 indicators between yields from 2013 and 2014 can be explained by the way in  
 356 which the physical properties affect plant growth. In tropical soils, physical  
 357 factors in the root zone that directly affects root growth and function are water  
 358 availability and soil strength (Letey, 1985). Each of these factors is closely related  
 359 to the soil water content, which unfortunately, changes constantly through the  
 360 growing season making it difficult to explain plant growth and yield responses  
 361 to physical soil conditions in a given season (Jayawardane & Chan, 1994).

362           The SOM substantially influenced the maize yield in 2013 and 2014. SOM  
363 increases resistance to structural breakdown by wetting, which is a characteristic  
364 behaviour of structurally fragile tropical soil. Therefore, only soils with less than  
365 a critical concentration of SOM are likely to breakdown (Nciizah & Wakindiki,  
366 2012). Mullins (1999) suggested a threshold of 2% SOM, below of which the soil  
367 would be vulnerable to hardsetting. The maps from the field showed that the  
368 area with SOM higher than the critical level of 2% was less than 1%. This finding  
369 suggests enormous risks for land degradation when SOM content is highly  
370 reliant on soil management practices. Indeed, environmental factors, such as high  
371 temperature, water, oxygen, and nutrient availability in humid tropical  
372 environments can accelerate residue decomposition and impair the possibility of  
373 increasing SOM contents (Christensen, 2000). Therefore, Moura *et al.* (2009) has  
374 suggested a constant addition of residues allowed by alley cropping in a no till  
375 system for maintaining the dynamic equilibrium between C inputs and high  
376 decomposition rates in humid tropical regions. Furthermore, in these  
377 circumstances, tillage can decrease SOM content increasing bulk density and  
378 water-suspendable solids, which can lead to a decline in structural and soil  
379 physical properties associated with an increase in penetration strength as  
380 observed by Ley *et al.* (1995). The negative correlation between grain yield and  
381 SPS in the 16 – 20 cm layer in this experiment may be explained by the generally  
382 decreasing SOM in soil depths.



383 Influence of calcium over yield can be accounted for by physical  
384 improvements caused by increasing flocculation and aggregation in the subsoil  
385 enhancing root activity, which leads to greater soil aggregation (Sumner, 2009).  
386 The positive effect of calcium on soil physical and chemical properties due to  
387 flocculation of soil particles, thereby creating supportive conditions in degraded  
388 tropical soil, was also reported by Anikwe *et al.* (2016). Less than 2% of the  
389 mapped area had a calcium content level higher than 26 mmol<sub>c</sub> dm<sup>-3</sup>, which  
390 shows great potential to improve soil fertility of this soil.

391 In addition, an association between calcium and SOM has been reported  
392 by some authors who describe the formation of cation bridges with products  
393 derived from decomposition of the applied biomass (Moore & Turunen, 2004;  
394 Whittinghill & Hobbie, 2012). The major cation involved in the formation of  
395 bridges is Ca<sup>2+</sup>, whose concentrations can affect fluxes of dissolved organic matter  
396 by stabilizing negatively charged organic matter through sorption to positively  
397 charged cations. The bond between polyvalent cations and negatively charged  
398 organic matter functional groups is not easily reversed and surfaces of organic  
399 materials will be less accessible for microbial activity. This flocculation prevented  
400 biological, chemical, or physical breakdown, which explains the spatial  
401 association between calcium and SOM (Figure 3) (Oste *et al.*, 2002). On the other  
402 hand, a correlation between SOM and contents of clay and silt was reported by  
403 Azlan *et al.* (2012). SOM tends to increase as the clay content increases due to  
404 bonds between the surface of clay particles and SOM, which retard the

405 decomposition process. In addition, soils with higher clay content increase the  
406 potential for aggregate formation. Macro-aggregates physically protect organic  
407 matter molecules from further mineralization caused by microbial attack (Rice,  
408 2002). Influence of increased pH on SOM content through stimulation of  
409 microbial activity was reported by Whittinghill & Hobbie (2012). These findings  
410 suggest that in soil prone to hardsetting, suitable calcium content could be  
411 achieved by gypsum application rather than lime material with carbonate or  
412 bicarbonate associated with the  $\text{Ca}^{2+}$ . According to Moura *et al.*, (2016), the  
413 application of gypsum plus leguminous residues can enhance the soil rootability  
414 when used on tropical soils prone to hardsetting. These effects facilitate greater  
415 root growth into deeper soil layers and greater nutrient uptake. In contrast, in  
416 tropical soil with low pH buffering capacity large quantity of lime application  
417 can raise pH to level, which favour the fast organic matter decomposition (Aye  
418 *et al.*, 2016).

419 In 2013 and 2014, around 4 % of the mapped field showed grain yield  
420 higher than 5 Mg ha<sup>-1</sup>, which is less than half the potential of the sown maize  
421 hybrid. Inadequate soil physical conditions and high atmospheric evaporative  
422 demand prevent crops from reaching their full productivity potential when  
423 intervals without rainfall greater than four days occur during the growing season  
424 (Moura *et al.*, 2016a). Low water use capacity in conditions with a narrow depth  
425 of the rootable layer may be crucial for crop growth (Atta *et al.*, 2013). Therefore,  
426 in soil tropical conditions of the upper soil layer must be improved by increasing

427 the volume in the soil available for root growth and water and nutrients uptake.  
428 Thus, to reduce soil hardening because of cohesion, plant residues may be  
429 applied in soil surface to preserve soil moisture by reducing evaporative losses  
430 and to increase organic matter and nutrients availability. Therefore, legume  
431 cover increases maize yield by a steady release of N and K during crop cycle,  
432 better distribution and maintenance of adequate levels of base in the root zone as  
433 well as enhanced water and nutrients uptake.

434         It is worth highlighting that a higher grain yield was achieved in the areas  
435 with three years of biomass application of leguminous cover, compared to the  
436 control area. Our experience in soil management of the region has shown that the  
437 paradigm of soil management established for other regions is not adequate for  
438 the conditions of a tropical environment. In the periphery of Brazilian Amazonia,  
439 agricultural practices that are recommended for the Brazilian savannah, such as  
440 improvements in the tilled topsoil cannot be maintained if deterioration of the  
441 porous soil structure is not prevented. Ripping cohesive soil without removing  
442 the causes of compaction might not improve yield. Each time the soil is tilled, it  
443 is aerated, which reduces the organic matter contents and accelerates the process  
444 of re-compaction, increasing strength after a year or less.

445         According to results from Moura *et al.* (2016a) in the same soil, positive  
446 interaction between applied calcium and leguminous biomass can increase  
447 nitrogen uptaken and maize grain yield. On the other hand, the low *coefficients of*  
448 *determination* (0.30 and 0.31) suggest that other predictors that were not

449 measured, such as the available nitrogen, may have affected the grain yields,  
450 which can help explain higher grain yield in leguminous areas.

#### 451 CONCLUSION

452 Our results suggest the possibility of intensifying the production and mitigating  
453 land degradation in the Amazonian periphery. For this must be recommended  
454 biomass and gypsum application to increasing soil organic matter and calcium  
455 contents to improve fertility and enhance crop productivity in farmland; thus,  
456 preventing deforestation of new areas. The main findings that support this  
457 conclusion are: i) high correlation of grain yield with the contents of soil organic  
458 matter and calcium, which can be increased by agronomic practices; ii) very low  
459 contents of these soil components in already cultivated soil; iii) a spatial  
460 association and high correlation between calcium and SOM and: iv) their  
461 influence on other significant indicators like macroporosity, bulk density and soil  
462 penetration strength. It is worth highlighting that due to the negative influence  
463 of pH on total independent contributions to increased soil organic matter, higher  
464 calcium contents must be achieved by gypsum application rather than using  
465 liming material.

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631

**Soil physical changes and maize growth in a structurally fragile  
tropical soil due to mulching and duration between irrigation intervals**

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CAPÍTULO III

1 **Soil physical changes and maize growth in a structurally fragile tropical soil due to**  
2 **mulching and duration between irrigation intervals**

3

4 V. R. A. Macedo<sup>1</sup>, E. G. Moura<sup>1</sup>

5

6 <sup>1</sup> *Agroecology Program, State University of Maranhao, CEP 65054-970 São Luís, MA,*  
7 *Brazil,*

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10 *Running title: Soil physical changes and maize growth*

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12

13

14 **Abstract**

15 Under tropical meteorological conditions, the volume of soil explored by plant roots is  
16 crucial for crop growth as it allows increased water and nutrient use efficiency. We  
17 hypothesized that, under different irrigation intervals, leguminous mulch can extend the  
18 duration between irrigation events but maintain crop performance, because decreased  
19 evaporative fluxes also reduce constraints to root exploration imposed by mechanical  
20 stress. We evaluated the combined effects of leguminous mulch and irrigation intervals  
21 on soil physical properties to determine whether the growth and productivity of maize  
22 were modified in a structurally fragile tropical soil. The experiment involved following  
23 treatments: 4-day irrigation intervals with soil mulched (4C) or bare (4S), 6-day irrigation  
24 intervals with soil mulched (6C) or bare (6S), 8-day irrigation intervals with soil mulched  
25 (8C) or bare (8S), 10-day irrigation intervals with soil mulched (10C) or bare (10S).  
26 Mulch decreased soil penetration resistance and increased to four days the favourable  
27 time for root development in drying soil. Relative to bare soil, mulch with a 6-day  
28 irrigation interval almost doubled nitrogen uptake post-tasseling, which decreased  
29 nitrogen remobilization and increased the crop growth rate during this stage. These  
30 conditions had a positive effect on the transpiration rate and stomatal conductance as well  
31 as on the growth and yield of maize. A 6-day irrigation interval with mulch compared to  
32 4-days with bare soil resulted in similar conditions for root development, but greater  
33 uptake of nitrogen (102.73 to 78.70 kg/ha) and better yield (6.2 to 5.3 t/ha), which means  
34 greater efficiency in nitrogen and water use.

35

36 **Keywords:** Water use efficiency, soil physical attributes, nitrogen uptake, physiological  
37 parameters, crop growth

38

39



40 **Introduction**

41 Under tropical meteorological conditions, due to the high evaporative demand, the depth  
42 of soil that roots can explore is crucial for water uptake, crop growth and nutrient use  
43 efficiency (Atta *et al.*, 2013). As water and nutrient availability are major requirements  
44 for crops, they greatly affect the volume of soil explored by plant roots (Wong & Asseng,  
45 2007). In tropical sandy loam soils, both, available water capacity and nutrient availability  
46 are often limited. In addition, due to the generally small contents of free-iron and organic  
47 carbon, combined with large proportions of fine sand and silt, most of this soil tends to  
48 harden during drying, which adversely affects root development (Daniells, 2012).  
49 Therefore, due to negative interactions between unsuitable soil conditions and the large  
50 evaporative demand, crops do not always reach their full productivity potential even  
51 though the availability of water and nutrients would be considered adequate in other  
52 regions (Moura *et al.*, 2013).

53 In soil that tends to harden with drying, effects on root growth are common when  
54 the water potential approaches -100 kPa (Becher *et al.*, 1997), which, according to Moura  
55 *et al.* (2009), may occur around the fourth day without rain or irrigation in tropical  
56 conditions. In addition, root systems are unable to completely exploit available nitrogen,  
57 required in large amounts by the crops. The restriction of root growth limits exploration  
58 to gain access to soil nitrogen, further suppressing crop productivity. This is especially  
59 true for annual crops that require time to develop extensive root systems because N may  
60 be lost gaseously by volatilization or denitrification, by leaching or by uptake by  
61 competing organisms during root development (Lynch, 2013).

62 Thus, mulching the soil surface with plant residues has been recommended by  
63 some researchers because it has the combined effect of protecting surface soil from  
64 erosion, as well as decreasing the rate of soil drying and hence hardening (Mulumba &  
65 Lal, 2008; Moura *et al.*, 2010).

66           However, understanding the relation between improved soil physical conditions,  
67 such as its rootability, and crop productivity may be challenging in rainfed crop  
68 conditions due to uncontrolled variations in soil moisture and the poor correlation  
69 between static physical properties (quantitatively not related to changes in soil moisture)  
70 and plant growth (Häkansson & Lipiec, 2000). Benjamin *et al.* (2003) suggested a soil  
71 physical indicator expressed as ‘water stress day’ to determine the total number of days  
72 when soil moisture contents are sufficiently reduced to cause crop water stress. In  
73 cohesive sandy loam soil in tropical regions, a water stress day might begin to be observed  
74 on the fourth day without rain or irrigation (Moura *et al.*, 2012).

75           We hypothesized that, under different irrigation intervals, leguminous mulch can  
76 extend the duration between irrigation events but maintain crop performance, because  
77 decreased evaporative fluxes also reduce constraints to root exploration imposed by  
78 mechanical stress. Thus, the objective of this work was to evaluate the effects of  
79 leguminous mulch and irrigation intervals on soil physical properties and how the growth  
80 and productivity of maize were modified in a structurally fragile tropical soil from the  
81 humid tropics. Increased uptake of water and nutrients is crucial for agricultural  
82 production in a tropical climate with a high potential evapotranspiration and soils with  
83 little available nutrient content, particularly with precipitation changes under global  
84 warming.

85

## 86 **Material and Methods**

### 87 *Experimental site and background information*

88 The experiment was conducted at Maranhão State University, Brazil (2°30'4" S,  
89 44°18'33" W), which had a hot, semi humid, equatorial climate with a mean precipitation  
90 of 2,100 mm year<sup>-1</sup> and two well-defined seasons, a rainy season extending from January

91 to June and a dry season without rain and with a pronounced water deficit extending from  
92 July to December. The evapotranspiration during the experimental period was 622.2 mm.  
93 The average temperature is approximately 27°C, with a maximum temperature of 37°C,  
94 and a minimum temperature of 23°C. The average potential evapotranspiration rate of the  
95 experimental period was 6.5 mm per day.

96 The soil showed cohesive characteristics (determined by relationship between  
97 penetration resistance and volumetric water content of soil) (Moura et al., 2009) and was  
98 classified as Arenic Hapludults (Soil Survey Staff, 2014). The A horizon had the  
99 following properties: pH 4.0 (in CaCl<sub>2</sub>); organic-C 20 g/kg (Walkley–Black); P 15  
100 mg/dm<sup>3</sup>; (Al + H) 25 mmol/dm<sup>3</sup>; Ca 15, Mg 9, K 1 mmol/dm<sup>3</sup> (resin), respectively;  
101 CEC 50 mmol/dm<sup>3</sup>; base saturation 46.2 %; coarse sand 300, fine sand 545, silt 61, clay  
102 90 g/kg, respectively. The soil properties were measured using standard methods (Carter  
103 & Gregorich, 2008).

104 The area had been fallow since 1990 and supported a native species of grass,  
105 which was killed using glyphosate. Lime and gypsum were broadcast by hand in  
106 December 2012 at rate of 1 t/ha and 6 t/ha, respectively. These application rates were  
107 calculated to raise base saturation to 55% and calcium to 45 mmol/dm<sup>3</sup>. Maize (cultivar  
108 AG 1051) was sown in the beginning of the rainy season in 2012, 2013, and 2014, with a  
109 between-row spacing of 80 cm and 25 cm between plants. The fertilizer application  
110 consisted of 50 kg/ha N as urea, 35 kg/ha of P as triple superphosphate, 100 kg/ha of K as  
111 potassium chloride and 5 kg/ha of Zn in the form of zinc sulphate.

112

### 113 *Trial set up*

114 The experiment was conducted under no-tillage conditions, and the experimental  
115 layout was established in October 2015 with 4, 6, 8, and 10-day irrigation intervals, with

116 soil mulched or bare. Four replicates of the treatments were arranged in a completely  
117 randomized block design. Irrigation water was conveyed to the experimental plots  
118 through a CPVC plastic pipe and applied in closed-ended furrows through a windowed  
119 tube. A water outlet port controlled the volume of water applied on each occasion and the  
120 total applied to treatments 4C, 4S was 440mm; 300 mm to treatments 6C, 6S; 220 mm to  
121 treatments 8C, 8S; 180 mm to treatments 10C, 10S. No rain-fell during the experimental  
122 period and the average daily evapotranspiration was 6.6 mm (October), 6.8 mm  
123 (November) and 7.1 mm (December).

124 Plot size was 8 x 5 m and the sampling area was 10 m<sup>2</sup>. The same maize cultivar  
125 was sown on October 10, 2015, with a density of four plants per square meter. The soil  
126 received the same rates of nutrients at planting as in preview years. Five days after  
127 emergence of the maize a total of 12 t/ha of dry biomass (branches and leaves) of *Acacia*  
128 *mangium* legume was applied between the rows. Biomass of *Acacia mangium* had been  
129 applied in the years 2012, 2013, and 2014 in the same plots and at the same rate (typical  
130 for alley cropping systems), according to Aguiar *et al.* (2010). The quality parameters of  
131 the acacia residue was as follows: a C/N ratio of 23.5; 22.2 g/kg of N; 3.1 g/kg of P; 5.1  
132 g/kg of K; 16.4 g/kg of Ca; 3.2 g/kg of Mg. The amount of nitrogen applied via legume  
133 was 216 kg/ha. In addition, 50 kg/ha N was applied as side dressing, when the fourth pair  
134 of corn leaves appeared.

135 In September 2015 the soil showed the following chemical properties: pH 5.1 (in  
136 CaCl<sub>2</sub>); organic-C 25 g/kg (Walkley–Black); P 25 mg/dm<sup>3</sup>; (Al + H) 18 mmol/dm<sup>3</sup>; Ca  
137 45, Mg 11, K 1,2 mmol/dm<sup>3</sup> (resin), respectively; CEC 75.2 mmol/dm<sup>3</sup>; base saturation  
138 76.1 %.

139

140

141 *Soil and plant measurements*

142 The soil penetration resistance (SPR) was measured with 5-cm gradations, at  
143 depths of 0-5 cm, 6-10 cm, 11-15 cm and 16-20 cm, using a digital penetrometer (Falker,  
144 Porto Alegre, Brazil) with three replicates per plot at stages: V8, tasseling and maturity  
145 of the maize. The table of critical level defined by Hazelton and Murphy (2007) was used  
146 to construct a graph of soil penetration resistance.

147 We sampled soil in volumetric rings with a 100-cm<sup>3</sup> capacity. We collected three  
148 replicates per plot from a depth of 5 to 12 cm. Samples were saturated, weighed, placed  
149 on a tension table and equilibrated at a potential of -6 kPa to drain pores >50 μm. After  
150 weighing, each replicate was oven-dried at 105°C. Soil dry bulk density ( $\rho_b$ ) was  
151 calculated as  $m/v$ , where:  $m$  = dry soil mass at 105°C and  $v$  = ring volume. Total porosity  
152 ( $\phi_t$ ) was calculated as  $[1 - (\rho_b/\rho_p)]$ , where:  $\rho_b$  = soil density (Mg/m<sup>3</sup>);  $\rho_p$  = particle density  
153 (Mg/m<sup>3</sup>), which was determined according to Embrapa (1997). We divided the total  
154 porosity into macroporosity and microporosity. We calculated the macroporosity,  
155 corresponding to the volume of pores with equivalent diameter larger than 50 μm as the  
156 mass difference between the samples at saturation and after equilibration at -6 kPa, and  
157 we calculated the microporosity, corresponding to the volume of pores with equivalent  
158 diameter narrower than 50 μm as the mass difference between the sample equilibrated at  
159 -6 kPa and the oven-dry sample (Thomasson 1978). All measurements were done  
160 immediately before irrigation of each treatment.

161 We determined the volumetric water content of soil at V8, tasseling, and maturity  
162 for the maize, on a daily basis for the 0-10cm and 10-20cm layers. Measurements were  
163 always made between 0700 and 0900 hours, with three replicates per plot. A TRIME-FM  
164 instrument (IMKO, Ettlingen, Germany) was used to collect time domain reflectometry

165 (TDR) data. All soil physical measurements were done immediately before irrigation of  
166 each treatment.

167 In the evaluation of the stomatal conductance ( $g_s$ ) and transpiration rate ( $E$ ), a  
168 Portable Measurement System for Gaseous Exchanges, an Infrared Gas Analyzer (IRGA)  
169 was used, LI-6400® model, LI-COR, (Lincoln, NE, USA). These physiological  
170 parameters were measured in two new fully expanded leaves, for three plants chosen at  
171 random in each plot, in the upper part of the canopy exposed to full sunlight, between  
172 0800 to 1000 hours. Three measurements were recorded automatically every 2 min for  
173 each leaf to ensure a steady-state condition for the gas exchange flow. The light units (the  
174 diode array contained blue and red LEDs), with the upper jaw enclosing the leaf, were  
175 used to ensure constant irradiance to replicate sunlight ( $1600 \mu\text{mol}/\text{m}^2/\text{s}$ ). The  
176 measurements were carried out in the tasseling stage immediately before irrigation.

177 Crop growth dynamics were monitored by weekly evaluations of crop growth rates  
178 (CGR) using leaf area (LA) and dry matter (DM) variations with time ( $t$ ). To accomplish  
179 this, five plants were collected per plot at seven-day intervals, from the 25th day after  
180 emergence. Seven collections were performed; the last collection coincided with the corn  
181 milky grain stage. Leaf area was measured with a Licor device; dry matter was determined  
182 after drying the plants in a forced-air circulation oven at  $70^\circ\text{C}$ .

183 Nitrogen content was measured on two occasions: at tasseling (or approximately  
184 one week before anthesis) and at the physiological maturity stage. At each sampling, five  
185 plants from each plot were randomly selected and separated into leaves, stalks and, at the  
186 second sampling, reproductive components. All of these plant materials were dried at  
187  $60^\circ\text{C}$  for 3-4 days to obtain a constant weight. Three subsamples were collected and  
188 ground to pass through a 1-mm screen. Total N concentration was determined in the maize  
189 following  $\text{H}_2\text{O}_4$  -  $\text{H}_2\text{O}_2$  digestion according to the standard distillation method described

190 by Cottenie (1980). Based on the measurements of plant dry matter (DM) and N uptake,  
191 we calculated the N remobilization: [Vegetative DM (N content) at silking - vegetative  
192 DM (N content) at maturity]. At the final harvest or at physiological maturity, the grain  
193 yield components were separately assessed in a 10-m<sup>2</sup> area and the values were adjusted  
194 according to moisture level of 145 g/kg.

195

#### 196 *Statistical analyses*

197 The data were analysed via analysis of variance (ANOVA), and the means were compared  
198 using Duncan post hoc test at a  $P = 0.05$  significance level. The data were analysed using  
199 InfoStat software (InfoStat Group, College of Agricultural Sciences, National University  
200 of Córdoba, Argentina). Except for the crop growth rate that was estimated and modeled  
201 with the ANACRES software (Portes & Castro Júnior 1991).

202

## 203 **Results**

### 204 *Changes in soil physical attributes and nitrogen uptake*

205 Physical measurements, with less variation in the time scale, such as soil density, and  
206 macroporosity were not affected by mulching or by irrigation intervals. However,  
207 microporosity was lower in the treatments 4S, 8S and 10S than in the treatments with  
208 mulching. However, total porosity was only statistically less in the bare soil treatments  
209 4S and 10S than in their mulched comparisons (Table 1).

210

211

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213

214

215

216 **Table 1** Soil dry bulk density ( $\rho_b$ ), microporosity, macroporosity and total porosity ( $\Phi_t$ )  
 217 in the experimental treatments in 5-12 cm layer.

Treatments	$\rho_b$	Microporosit y	Macroporosit y	$\Phi_t$
	g/cm <sup>3</sup>	----- cm <sup>3</sup> /cm <sup>3</sup> -----		
4S	1.44 a	0.15 c	0.29 a	0.44 b
4C	1.39 a	0.17 b	0.30 a	0.47 a
6S	1.44 a	0.16 bc	0.29 a	0.46 a
6C	1.40 a	0.18 ab	0.30 a	0.47 a
8S	1.45 a	0.15 c	0.30 a	0.46 a
8C	1.39 a	0.19 a	0.29 a	0.48 a
10S	1.43 a	0.14 c	0.30 a	0.44 b
10C	1.40 a	0.17 b	0.32 a	0.47 a

4S= 4 days irrigation intervals in bare soil; 4C= 4 days irrigation intervals with soil covered; 6S= 6 days irrigation intervals in bare soil; 6C= 6 days irrigation intervals with soil covered; 8S= 8 days irrigation intervals in bare soil; 8C= 8 days irrigation intervals with soil covered; 10C= 10 days irrigation intervals with soil covered, 10S= 10 days irrigation intervals in bare soil. Different letters in the same column indicate differences at the 5% level by Duncan's test.

218

219 In the 0–10cm layer, mulching significantly increased water volumetric contents only  
 220 when the interval of irrigation was 4 days, at V8 and at the tasseling stage. In the 10–  
 221 20cm layer, volumetric water content was increased by mulching at V8 (8 leaves  
 222 produced) and at the tasseling stage in the 4- and 6-day irrigation intervals (Table 2).

223

224

225



226 **Table 2** Volumetric water content ( $\text{m}^3/\text{m}^3$ ) in the experimental treatments in V8 stage, at  
 227 tasseling and at maturity, in 0 – 10 cm and 10 to 20 cm layer.

	V8	At tasseling	At maturity
0 – 10 cm	----- $\text{m}^3/\text{m}^3$ -----		
4S	0.11 b	0.15 b	0.11 a
4C	0.16 a	0.18 a	0.13 a
6S	0.07 c	0.06 c	0.08 b
6C	0.06 c	0.07 c	0.09 ab
8S	0.06 c	0.05 c	0.07 bc
8C	0.06 c	0.08 c	0.09 ab
10S	0.04 c	0.05 c	0.05 c
10C	0.04 c	0.07 c	0.06 bc
10 – 20 cm			
4S	0.06 ab	0.06 b	0.07 b
4C	0.07 a	0.08 a	0.11 a
6S	0.06 ab	0.04 c	0.04 c
6C	0.06 ab	0.06 b	0.06 b
8S	0.04 bc	0.04 c	0.04 c
8C	0.06 ab	0.04 c	0.07 b
10S	0.03 c	0.04 c	0.03 c
10C	0.03 c	0.04 c	0.04 c

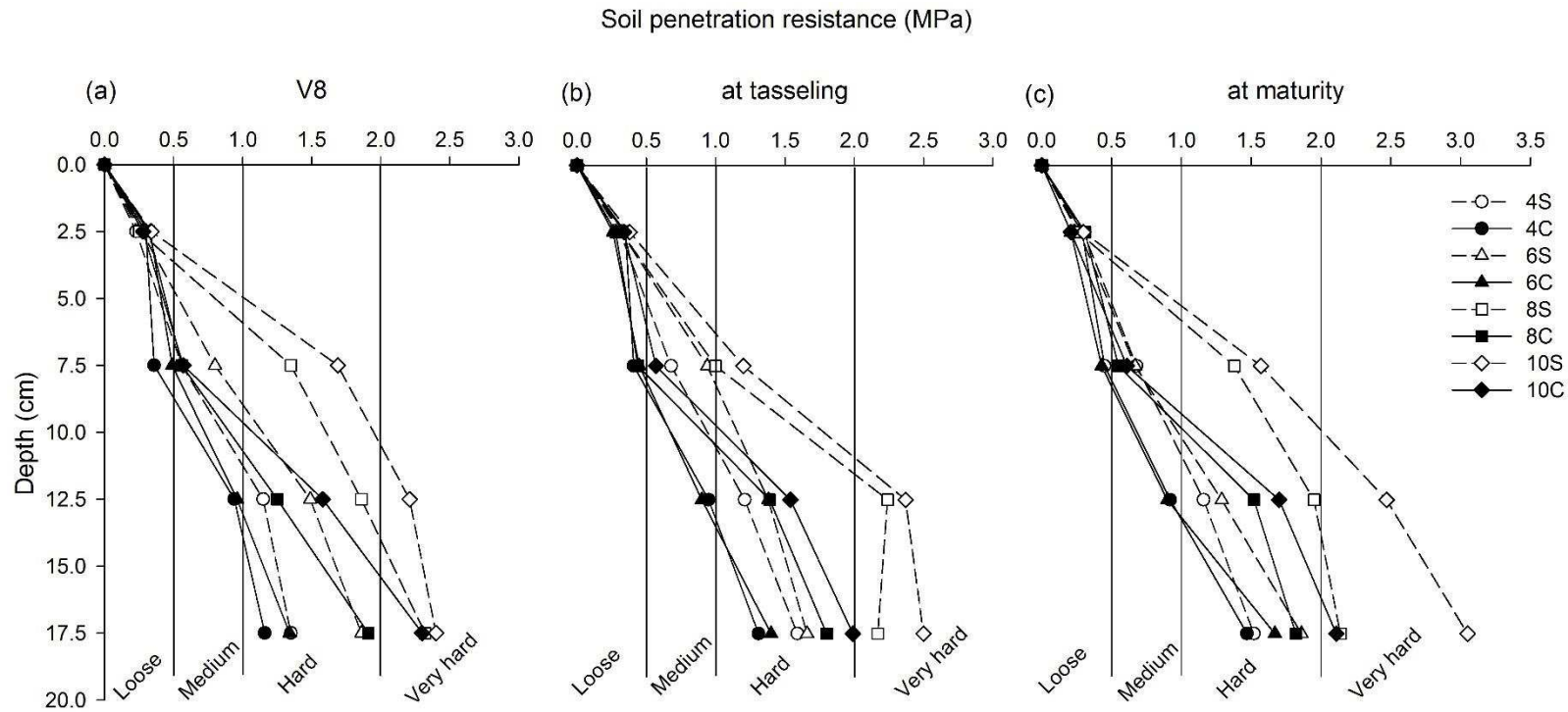
4S= 4 days irrigation intervals in bare soil; 4C= 4 days irrigation intervals with soil covered; 6S= 6 days irrigation intervals in bare soil; 6C= 6 days irrigation intervals with soil covered; 8S= 8 days irrigation intervals in bare soil; 8C= 8 days irrigation intervals with soil covered; 10C= 10 days irrigation intervals with soil covered, 10S= 10 days irrigation intervals in bare soil. Different letters in the same column indicate differences at the 5% level by Duncan's test.

228

229 In contrast, penetration resistance (SPR) was strongly influenced by soil water

230 content, varied considerably over time and was significantly affected by mulching (Figure

231 1). However, effects were significantly only in the 0-5 cm layer. In the 5–15 cm layers,  
232 the effects of mulch on decreasing SPR was significant, when comparing treatments with  
233 and without mulch for the same irrigation intervals ( $P < 0.05$ ). Furthermore, in the 5 –  
234 10cm layer, SPR was smaller for treatment 10C than in 6S at all stages. In 10 – 15 cm  
235 layer SPR was greater in 8S than in 10C during the entire cycle of the maize, while in the  
236 15 – 20cm layer there was no difference between 4S to 6C, 6S to 8C and 8S to 10C ( $P >$   
237  $0.05$ ). In the treatments without mulch, although the effect of irrigation was always more  
238 significant when the intervals were more than four days, SPR was greater in 8S than in  
239 6S in the 5 – 20 cm layer at first and third stage, and in the 10 – 20 cm layer at the second  
240 stage ( $P < 0.05$ ). In the plots with mulch, differences were only observed below the 10–  
241 15cm layer, where SPR was greater in 10C and 8C than in 4C in all stages ( $P < 0.05$ ).  
242 According to Hazelton and Murphy (2007), in the treatments 8S and 10S, the SPR  
243 indicated that the soil was hard in the 5-10cm layer and very hard in the 10–20cm layer  
244 in all stages. Below 10cm, SPR was hard in 4S, 6C, 6S and 8C. In the treatments 4C and  
245 6C, SPR was moderate in 10–15 cm layer and hard of 15–20cm layer. There were no  
246 interactions between irrigation intervals and mulching regarding variation in SPR.  
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**Figure 1** Soil penetration resistance at three stage of maize: (a) V8, (b) at tasseling and (c) at maturity, in the experimental area in 2015. 4S= 4 days irrigation intervals in bare soil; 4C= 4 days irrigation intervals with soil covered; 6S= 6 days irrigation intervals in bare soil; 6C= 6 days irrigation intervals with soil covered; 8S= 8 days irrigation intervals in bare soil; 8C= 8 days irrigation intervals with soil covered; 10C= 10 days irrigation intervals with soil covered, 10S= 10 days irrigation intervals in bare soil. Vertical bars mean the critical levels by Hazelton and Murphy (2007).

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251 Uptake of nitrogen was affected by mulching and irrigation intervals, with  
252 different intensities in the stages evaluated (Table 3). At the pre-tasseling stage, the  
253 differences were small and just 4C, and 10S treatments showed significant differences in  
254 nitrogen content. In 4C it was enhanced, while in 10S it was less than in all the other  
255 treatments ( $P < 0.05$ ). Nitrogen uptake post tasseling widely varied. At all irrigation  
256 intervals, it was always greater in plots with residue. The largest nitrogen content at post-  
257 tasseling was observed in the 6C treatment, where it was almost 50% greater than in 6S  
258 and was larger than in the 4C treatment ( $P < 0.05$ ). In 10S the nitrogen uptake post  
259 tasseling was less than for all other treatments. There was no significant difference  
260 between 8C and 4S treatments ( $P < 0.05$ ). The most nitrogen remobilized occurred in 8S,  
261 whereas in 6C the amount was similar to that in 10S but in both it was less than in all  
262 other treatments ( $P < 0.05$ ). The total uptake of nitrogen was small, even in treatments  
263 with a greater nitrogen content. The greatest amount of nitrogen accumulated was in  
264 treatments 4C and 6C, and was more than twice that in 10S, where accumulation was  
265 least. In 4S, 6S, 8S, 8C and 10C the total nitrogen was less than in 4C and 6C ( $P < 0.05$ ).  
266 In 4C, 6C, 6S and 8C more than half of nitrogen was accumulated post – tasseling,  
267 whereas in 8S and 10S it was less than 40%.

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274 **Table 3** Amount of nitrogen at tasseling (NT), amount of nitrogen remobilized (NR),  
 275 amount of nitrogen uptake post-tasseling (NPT), total nitrogen accumulated (TN),  
 276 percentage of nitrogen uptake post-tasseling (NPT%), in the experimental area in 2015.

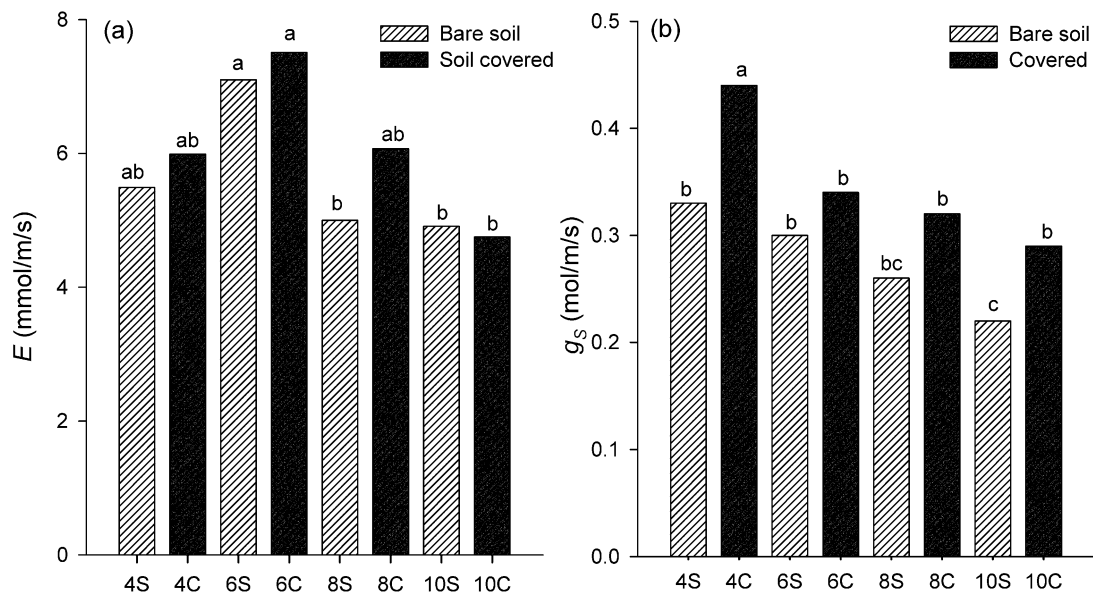
Treatments	NT	NR	NPT	TN	NPT%
(kg/ha)					
4C	49.5 a	13.8 bc	51.1 b	100.6 a	50.3 b
4S	45.4 ab	18.0 ab	33.2 cd	78.7 b	42.1 bc
6C	40.8 ab	5.6 d	61.9 a	102.7 a	60.3 a
6S	39.8 ab	10.6 c	42.5 bc	82.3 b	51.0 b
8C	38.1 b	11.0 c	38.2 c	76.4 b	50.2 bc
8S	49.2 ab	20.1 a	24.9 d	74.1 b	33.6 d
10C	39.7 ab	15.9 ab	30.2 d	69.9 b	44.1 bc
10S	25.4 c	7.0 cd	14.8 e	40.1 c	37.7 cd

4S= 4 days irrigation intervals in bare soil; 4C= 4 days irrigation intervals with soil covered; 6S= 6 days irrigation intervals in bare soil; 6C= 6 days irrigation intervals with soil covered; 8S= 8 days irrigation intervals in bare soil; 8C= 8 days irrigation intervals with soil covered; 10C= 10 days irrigation intervals with soil covered, 10S= 10 days irrigation intervals in bare soil. Different letters in the same column indicate differences at the 5% level by Duncan's test.

277

278 *Changes in physiological parameters and crop growth*

279 The transpiration rate was not affected by mulching when comparing the same irrigation  
 280 intervals (Figure 2). However, when comparing different intervals 6C = 4S, 8C = 6S and  
 281 10C = 8S ( $P < 0.05$ ). The stomatal conductance was smaller in 10S than in 6S and also  
 282 smaller than in all plots with mulch. Conductance in treatment 4C was greater than in all  
 283 other treatments and was double that in 10S, which was similar to that for 8S. There were  
 284 significant differences between 4S, 6S, 6C, and 8C ( $P < 0.05$ ). There were no interactions  
 285 between irrigation intervals and mulching regarding variation in transpiration rate.



**Figure 2** (a) Transpiration rate ( $E$ ) and (b) stomatal conductance ( $g_s$ ) of maize in the experimental area in 2015. 4S= 4 days irrigation intervals in bare soil; 4C= 4 days irrigation intervals with soil covered; 6S= 6 days irrigation intervals in bare soil; 6C= 6 days irrigation intervals with soil covered; 8S= 8 days irrigation intervals in bare soil; 8C= 8 days irrigation intervals with soil covered; 10C= 10 days irrigation intervals with soil covered, 10S= 10 days irrigation intervals in bare soil. Different letters indicate differences at the 5% level by Duncan's test.

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The crop growth rate (CGR) was affected by mulching in different intensities

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dependent on the irrigation interval or stage of the maize (Figure 3). From 44 days after

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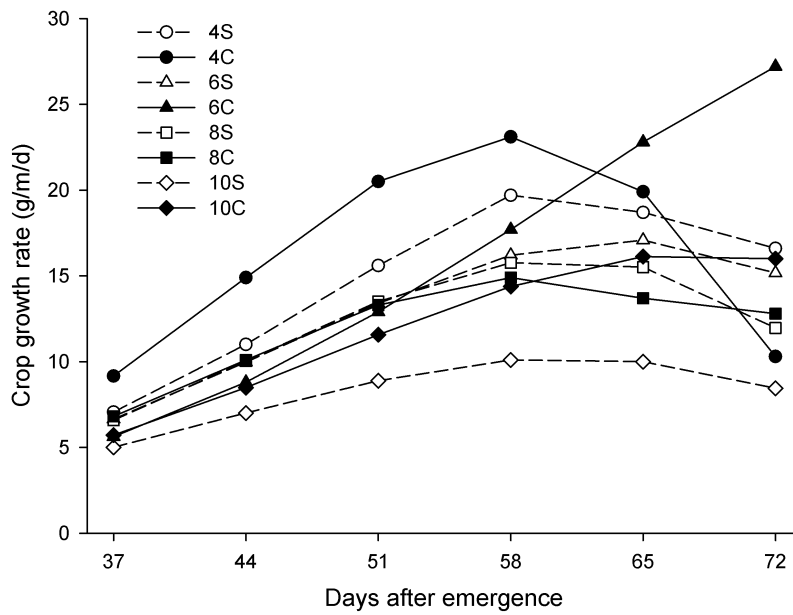
emergence until at tasseling stage (65 days after emergence) CGR was greater in 4C and

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4S ( $P < 0.05$ ). CGR continued to increase after tasseling only in the plants from plots with

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residue. 10S had the lowest CGR during the entire evaluated period ( $P < 0.05$ ).



**Figure 3** Crop growth rate of maize in the experimental area in 2015. 4S= 4 days irrigation intervals in bare soil; 4C= 4 days irrigation intervals with soil covered; 6S= 6 days irrigation intervals in bare soil; 6C= 6 days irrigation intervals with soil covered; 8S= 8 days irrigation intervals in bare soil; 8C= 8 days irrigation intervals with soil covered; 10C= 10 days irrigation intervals with soil covered, 10S= 10 days irrigation intervals in bare soil.

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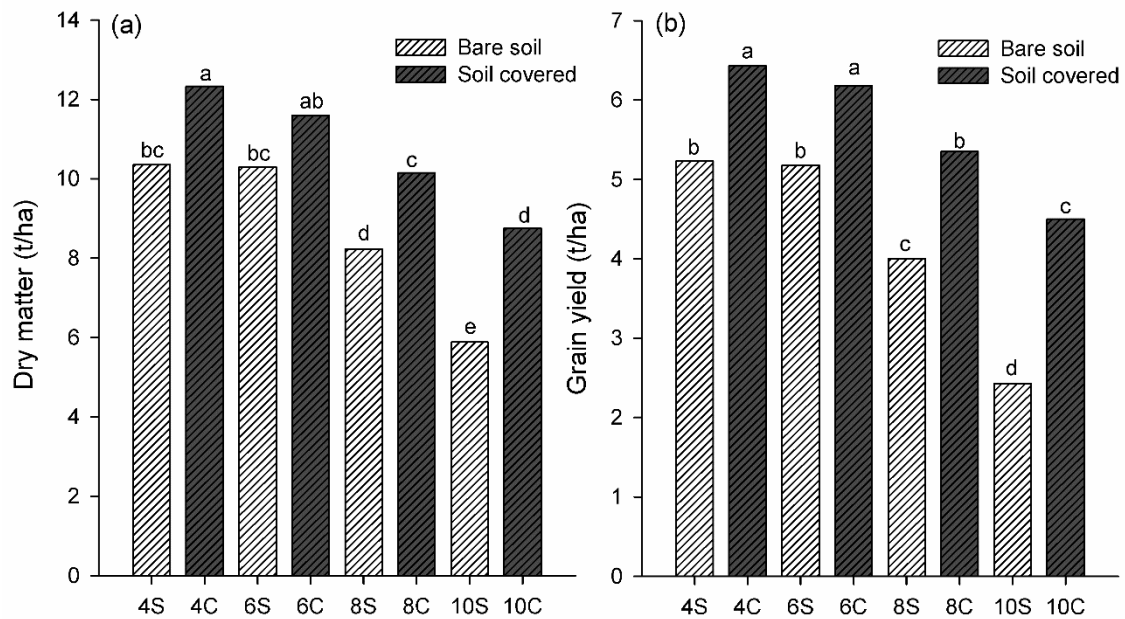
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The mulch significantly increased the dry matter production in 4-, 8-, and 10-day irrigation intervals (Figure 4). In plots without residue the dry matter production was as follows: 4S = 6S > 8S > 10S, whereas in plots with residues was 4C = 6C > 8C > 10C. ( $P < 0.05$ ). Mulching increased yield at all irrigation intervals. There was no significant difference between 4 and 6 days of irrigation intervals when they are compared with or without mulch. In contrast, the grain yield was higher in 8S than in 10S as well as in 8C than in 10C ( $P < 0.05$ ). There were no interactions between irrigation intervals and mulching to variation in grain yield.



**Figure 4** (a) Dry matter and (b) grain yield of maize in the experimental area in 2015. 4S= 4 days irrigation intervals in bare soil; 4C= 4 days irrigation intervals with soil covered; 6S= 6 days irrigation intervals in bare soil; 6C= 6 days irrigation intervals with soil covered; 8S= 8 days irrigation intervals in bare soil; 8C= 8 days irrigation intervals with soil covered; 10C= 10 days irrigation intervals with soil covered, 10S= 10 days irrigation intervals in bare soil. Different letters indicate differences at the 5% level by Duncan's test.

302

### 303 Discussion

304 Although the effect of mulching on soil penetrometer resistance (SPR) decreased  
 305 with depth, comparing 4S to 6C, 6S to 8C and 8S to 10C, our results showed that the  
 306 damaging effect on SPR in drying soil could be delayed for two days by use of mulch. In  
 307 the 0 – 10cm layer the delay may be up to four days if comparing 6S to 10C. Lack of  
 308 differences in the transpiration rate between treatments 6C/4S, 8C/6S and 10C/8S also  
 309 confirms a physiological effect of the mulch in decreasing the number of water stress days  
 310 by two. This delay was also observed in the stomatal conductance in the irrigation  
 311 intervals of four and 10 days. This can also be important for rainfed crops in tropical  
 312 conditions when intervals larger than four days without rain can significantly impair the  
 313 crop root growth, water uptake and crop yield (Moura *et al.*, 2016).



314 According to Mulumba & Lal (2008), residues on the soil surface shade the soil,  
315 serve as a vapour barrier against moisture losses from the soil and therefore more water  
316 is conserved in the soil profile during the early growth period with mulch than without it.  
317 However, in this experiment the small differences in volumetric water contents cannot  
318 explain the differences in the maize crop. Furthermore, residue application in previous  
319 years may have contributed to an environment favourable to root growth and water uptake  
320 promoting the formation of ‘ephemeral structures’ in the structurally fragile soil by  
321 increasing the free light fraction of soil organic matter (Shepherd *et al.*, 2002). It is worth  
322 highlighting that even in plots with mulch, the 15 – 20 cm layer was hard four days after  
323 irrigation, which shows importance of this practice in enhancing rootability in this soil.

324 Several critical conditions may explain the differences in nitrogen uptake in this  
325 experiment. First, the application of nitrogen in sidedressing at the V8 maize stage  
326 decreased differences in nitrogen content among treatments, at the pre-tasseling stage.  
327 Second, post tasseling, the nitrogen uptake was greater in plots with mulch due to  
328 enhanced soil rootability with lower penetration strength, as well as, due to increased  
329 availability of nitrogen released by leguminous residue.

330 Absorption of N is highly dependent on root development and therefore on soil  
331 rootability, while increased root system growth results in greater uptake and less N loss  
332 (Garnett *et al.*, 2009). In addition, a steady nitrogen release by leguminous biomass during  
333 the entire crop cycle is more important in humid tropical conditions compared to the rapid  
334 early availability of nitrogen that results from application of synthetic fertilizers (Aguiar  
335 *et al.*, 2010). Lastly, the greater water application rate in 4C may have impaired the  
336 nitrogen uptake, by leaching induction, compared to 6C. Some authors (Bronson, 2008;  
337 Wang *et al.*, 2012) report that loss of nitrogen by leaching increases with water infiltration  
338 rates and amount of water applied by irrigation. According to Moura *et al.* (2009) the

339 water infiltration rate may achieve up to 70 mm/h in the same soil with residue. These  
340 results may be very important to water use efficiency and nitrogen, since greater water  
341 application rates can result in a wasted water and nitrogen. On the other hand, for rainfed  
342 crops the humid tropics, which receive large amounts of rainfall, this means that the  
343 fragmentation of the fertilizer application can lead to greater efficiency than a single  
344 application at the V8 stage.

345         The smallest nitrogen remobilization in 6C is related to the greatest post tasseling  
346 nitrogen uptake under this treatment. Remobilization in different plant organs is needed  
347 to route nitrogen to grains during the period of grain-fill, when nitrogen uptake is  
348 generally insufficient for the high demand of the seeds (Borrel *et al.*, 2000). In contrast,  
349 the relative flow and remobilization of nitrogen to the grain during grain-fill period can  
350 also be analysed in terms of the source of nitrogen to redistribution and sites for  
351 reutilization and storage. Therefore in 10S, without N from legumes and with small cobs,  
352 remobilization was less than in other treatments. On the other hand, N depletion, due to  
353 remobilization, tends to accelerate the senescence in the leaves, reducing the  
354 photosynthesis decreasing the dry matter and yield (Masclaux-Daubresse *et al.*, 2010).

355         The increase in growth rate of the plants after tasselling, on the plots with mulch,  
356 reflected the greater availability and uptake of nitrogen promoted by the leguminous  
357 residue. However, greater dry matter and grain yield was obtained in treatments with  
358 larger amounts of total nitrogen uptake. For yield, the results 10C = 8S, 8C = 6S suggest  
359 that the effect of mulch on decreasing water stress days and on nitrogen uptake increased  
360 maize production. In addition, the results of 6C = 4C > 4S indicated that this effect was  
361 influenced by nitrogen uptake post tasseling when the irrigation interval was not harmful  
362 to maize growth. Thus, the greater growth and yield of the maize in plots with mulch can  
363 be accounted for most likely by its effect on water stress and nitrogen uptake.

364 In the Amazonian periphery with soil that is difficult to manage, small intervals  
365 without rain can decrease maize grain yield due to water stress and poor nutrient uptake.  
366 From a practical point of view, our results mean that, the use of mulch in a quantity similar  
367 to that commonly produced in alley cropping system can increase the viability of a crop  
368 even in cultivated and cohesive soils predominant in the region. From an environmental  
369 point of view, this may reduce expansion of farming into forested or in fallow areas, and  
370 it may prevent more deforestation, which is a critical issue in this region.

371

### 372 **Conclusion**

373 Mulch used on tropical soil surfaces may decrease soil penetration resistance, thereby  
374 expanding the time that root systems can develop in drying soil, which has a positive  
375 effect on the physiological parameters as well as on the growth and yield of maize.  
376 Furthermore, mulch of leguminous residue increased nitrogen uptake at the post tasseling  
377 stage, which decreased the remobilization and increased total uptake nitrogen. The use of  
378 6-day irrigation intervals with mulch compared to 4 days without mulch resulted in the  
379 same conditions for root development, greater amounts of nitrogen uptake and larger  
380 yields, which means greater water use efficiency.

381

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