

UNIVERSIDADE ESTADUAL DO MARANHÃO

CENTRO DE CIÊNCIAS AGRÁRIAS

PROGRAMA DE PÓS-GRADUAÇÃO EM AGROECOLOGIA

CURSO DE DOUTORADO EM AGROECOLOGIA

**COMPOSIÇÃO DA COMUNIDADE DE VESPAS E ABELHAS SOLITÁRIAS EM
VEGETAÇÃO DE RESTINGA DE SERGIPE**

DANÚBIA MARIA DA COSTA

SÃO LUÍS - MA

2019

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Orientador: Prof. Dr. Adenir Vieira Teodoro

Co-orientadora: Dra. Michela Costa Batista

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O conhecimento emerge apenas através da invenção e reinvenção, através da inquietante, impaciente, contínua e esperançosa investigação que os seres humanos buscam no mundo, com o mundo e uns com os outros.

Paulo Freire

Dedico

Aos meu familiares, em especial aos meus pais Gláucia e Antônio, aos meus irmãos Henrique e Mariana e minha cunhada Fabíola por todo apoio.

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RESUMO

Abelhas e vespas solitárias são himenópteros que nidificam em cavidades pré-existentes em ambientes naturais, exercem um papel chave no funcionamento dos ecossistemas promovendo serviços ambientais como, polinização e controle biológico natural. O objetivo do estudo foi avaliar a composição da comunidade de abelhas e vespas solitárias em vegetação de restinga, e determinar como as condições climáticas influenciam na nidificação em restinga da Reserva Particular do Patrimônio Natural do Caju (RPPN). Usamos ninhos-armadilha para a coleta de ninhos e as avaliações foram mensais no período de dois anos. Foram coletados 735 ninhos de 4 espécies de abelhas e 15 de vespas. A nidificação ocorreu principalmente em áreas abertas durante a estação seca, reduzindo na estação chuvosa. A composição de espécies diferiu entre área aberta e vegetação secundária, e no geral, a contribuição dos componentes da beta diversidade (turnover e aninhamento) foi similar entre os tipos de vegetação, no entanto, na estação seca o turnover foi maior em vegetação secundária e arbustiva. Por outro lado, vegetação arbustiva foi completamente aninhada à área aberta em ambas estações e o turnover foi o principal responsável pela beta diversidade ao longo do tempo. Além disso, nós avaliamos como as condições climáticas, o tipo de vegetação e o estrato arbóreo influenciam na nidificação pela abelha *Centris tarsata*. A nidificação foi maior durante a estação seca, em áreas abertas a 1.5m de altura. No entanto, em vegetação arbustiva e secundária a nidificação parece ser favorecida por uma variação específica de temperatura e pluviosidade. Além disso, a quantidade de células de cria, adultos e comprimento de ninho foram maiores em áreas abertas que outros tipos de vegetação. Os resultados sugerem que abelhas e vespas solitárias respondem diretamente às condições climáticas e preferem nidificar em locais com grandes quantidades de recursos como material de construção dos ninhos e alimento para as crias.

Palavras-chave: ninhos-armadilha, controle biológico natural, polinizadores, Mata Atlântica.

ABSTRACT

Cavity-nesting bees and wasps are hymenopterans that nest in pre-existing cavities in natural environments, and play a key role in ecosystem functioning providing environmental services such as pollination, and biological control. Here, we aimed at assessing the species composition of cavity-nesting bee and wasp community in restinga vegetation, and determine how climatic conditions influence nesting and assembly composition in restinga vegetation from Reserva Particular do Patrimônio Natural do Caju (RPPN). We used trap-nests and evaluations were conducted monthly over two years. We collected 735 nests from 4 bee species and 14 wasp species. Mostly nesting occurred in open areas during dry season, and decreased in the rainy season. The species composition differed for open areas and secondary vegetation, overall, contribution of beta diversity components turnover and nestedness was similar between vegetation types, but, in the dry season turnover was higher in secondary and shrubby vegetation. Conversely, shrubby vegetation was completely nested to open areas in both seasons and turnover was the main responsible for beta diversity over the time. In addition, we assessed how climatic conditions, vegetation type and arboreal strata influence *Centris tarsata* nesting, the most abundant bee species in RPPN do Caju. Overall, nesting was higher during the dry season, mostly in open areas at 1.5m height. However, in shrubby and secondary vegetations nesting seems to be favored by a specific rainfall and temperature range. Furthermore, the amount of brood cells, total number of adults and nest length were higher in open areas than other vegetation types. The results suggest that bees and wasps respond directly to climate conditions, and these arthropods show preference for nesting in places with greater resource availability such as building materials and food for their offspring.

Keywords: trap-nests, natural biological control, pollinators, Atlantic Rainforest.

CAPÍTULO 1

Referencial teórico

1 - REFERENCIAL TEÓRICO

1.1 - Vegetação de restinga

A Mata Atlântica já foi uma das maiores florestas tropicais das Américas, originalmente cobriu cerca de 150 milhões de hectares em condições ambientais altamente heterogêneas. Extende do Rio Grande do Norte ao Rio Grande do Sul por todas as formações florestais ao longo da costa brasileira (LANDIM & SANTOS, 2001). No Brasil, apesar da constante exploração, por mais de 500 anos ainda abriga uma alta biodiversidade (RIBEIRO *et al.*, 2009; ROCHA-FILHO *et al.*, 2017). Dentro deste bioma está inserido o ecossistema de restinga, o qual é considerado como uma vegetação de regiões costeiras tropicais e subtropicais que cresce em planícies arenosas sob influência marinha. No entanto, esse ecossistema está em constante ameaça devido à expansão urbana e especulação imobiliária (NOGUEIRA JUNIOR, *et al.*, 2013; OLIVEIRA & LANDIM, 2014; SERRA *et al.*, 2016).

A composição florística das restingas ao longo do litoral brasileiro é altamente variável, tanto em escala regional quanto local. Os diferentes tipos de vegetação variam desde formações herbáceas, passando por formações arbustivas abertas ou fechadas, chegando a florestas, cujo dossel varia em altura, geralmente não ultrapassando 20m (SILVA, 2000).

As formações herbáceas ocorrem nas faixas de praia e locais que podem ser atingidos pelas marés mais altas ou depressões alagáveis. Nestas, predominam espécies herbáceas, em alguns casos com pequenos arbustos e árvores que podem ocorrer de forma isolada ou formando agrupamentos mais densos, com variações fisionômicas, composição e graus de cobertura (SILVA, 2000). Formações arbustivas são os tipos fisionômicos que mais chamam atenção, variando desde densos emaranhados de arbustos misturados (trepadeiras, bromélias terrícolas e cactáceas) a moitas com extensão e altura variáveis, intercaladas por áreas abertas expondo diretamente a areia que forma o substrato (ZICKEL *et al.*, 2004).

Já as formações florestais variam ao longo de toda a costa, tanto em aspectos florísticos como estruturais, sendo influenciadas pela vegetação adjacente e pelas características do substrato (origem, composição e condições de drenagem). Esse tipo de vegetação a altura do estrato superior pode variar de 15 a 20m de altura (SILVA, 2000).

1.2 - A Reserva Particular do Patrimônio Natural (RPPN) do Caju

A Reserva Particular do Patrimônio Natural (RPPN) do Caju é uma Unidade de Conservação pertencente à Embrapa Tabuleiros Costeiros criada em 17 de janeiro de 2011, era uma antiga propriedade conhecida regionalmente como Fazenda do Caju. Atualmente, são realizadas atividades de pesquisas científicas, visitas de educação ambiental e atividades com comunidades ao entorno. A RPPN do Caju está localizada no município de Itaporanga D'Ajuda ($11^{\circ}06'16.19"S$; $37^{\circ}11'05.89"O$) a 29 km de Aracaju - SE, abrangendo uma área de 763,37 hectares, sendo 416,08 de zona de proteção (Figura 1). O clima da região é tropical com verão seco, estação chuvosa no outono e precipitação acima de 1.250 mm^3 . O bioma em que está inserida é o de Mata Atlântica, em seu interior são encontrados apicuns (áreas sem vegetação com substrato de areia que recebem influência direta do regime de marés), mangue e vegetação de restinga em mosaico. O relevo é predominantemente plano com pequenas áreas onduladas, solos arenosos e baixa fertilidade natural (NOGUEIRA JUNIOR *et al.*, 2013).

A vegetação de restinga da RPPN do Caju pode ser caracterizada em três tipos:

Vegetação de área aberta - incidência direta de luz solar, presença de vegetação herbácea, predominância de gramíneas como o capim-gengibre (*Paspalum maritimum* Trind), plantas herbáceas, o cacto coroa-de-frade (*Melocactus zehntneri* Britton & Rose), e presença esparsa de espécies arbóreas.

Formação de vegetação arbustiva - sombreamento parcial com alta luminosidade, predominância de espécies arbustivas e em menor número espécies arbóreas de porte médio como o murici (*Byrsonima crassifolia* L. Kunth), cambuí (*Myrciaria tenella* D.C. Berg), aroeira da praia (*Schinus terebinthifolius* Radji), piaçava (*Attalea funifera* Martius) e gramíneas esparsas.

Formações florestais (vegetação secundária) - composta por vegetação secundária com solo coberto por grande quantidade de serapilheira, sombreamento com baixa luminosidade e alta umidade relativa do ar, vegetação densa com predominância de espécies arbóreas de médio a grande porte como o cajueiro (*Anacardium occidentale* L), mangaba (*Hancornia speciosa* Gomes), pau-pombo (*Tapira guianensis* Aublet) e ingá (*Inga* sp.).



Figura 1 - Mapa da RPPN do Caju, localizada em Itaporanga D'Ajuda, SE, usando imagem do satélite Sentinel-2.

1.3 - Escalas de diversidade

Compreender os padrões de diversidade de espécies é vital para avaliar os impactos do manejo florestal e para desenvolver estratégias de conservação (RUBENE *et al.*, 2014). A diversidade de espécies em sistemas naturais e paisagens fragmentadas é afetada por múltiplos processos operando em múltiplas escalas (GAVISH *et al.* 2019). Whittaker (1972) definiu três importantes escalas de diversidade: alfa (riqueza de espécies no habitat), beta (variação nas identidades das espécies entre os habitats), e a diversidade gama que é a riqueza de espécies regional, entre uma variedade de habitats (uma paisagem, área geográfica, ilha).

O estudo da diversidade beta permite testar diferentes hipóteses sobre os processos que direcionam a distribuição de espécies e há duas formas em que as comunidades de espécies podem ser “diferentes”. A primeira é a substituição de espécies (turnover) em um local por diferentes espécies em outro local, de modo que as duas comunidades são diferentes entre si (BASELGA, 2012). A segunda é a perda (ou ganho) de espécies em somente um dos locais, o que implica a eliminação (ou adição) de espécies, fazendo com que a comunidade mais pobre seja um subconjunto da mais rica (um padrão chamado aninhamento) (BASELGA & ORME, 2012), a ocorrência de espécies tende à sobrepor uma à outra e compartilhar locais em comum (ULRICH & GOTELLI, 2007) (Figura 2).

Portanto, a seleção da medida de dissimilaridade para quantificar as diferenças entre as assembléias é crucial pois, existem diferentes maneiras de medi-la (BASELGA & ORME, 2012). Nesse sentido, Baselga (2010, 2012) propôs um método para a partição da dissimilaridade total (beta diversidade) em dois componentes separados: a dissimilaridade da substituição de espécies (turnover) e a dissimilaridade do aninhamento (BASELGA & ORME, 2012). Na natureza, a substituição de espécies pode refletir a separação de espécies pelo meio ambiente ou a dispersão, enquanto o aninhamento é frequentemente relacionado à dinâmicas de extinção-colonização ordenadas (SI *et al.*, 2016, SOININEN *et al.*, 2017). Portanto, a diversidade beta interage com os gradientes de diversidade alfa (riqueza), e ambos os componentes da biodiversidade (alfa e beta) resultam da montagem da comunidade por meio de filtros locais e regionais. Assim, a diversidade beta pode capturar a natureza dinâmica dos padrões de diversidade melhor do que medidas simples de diversidade alfa (SOININEN, *et al.*, 2017).

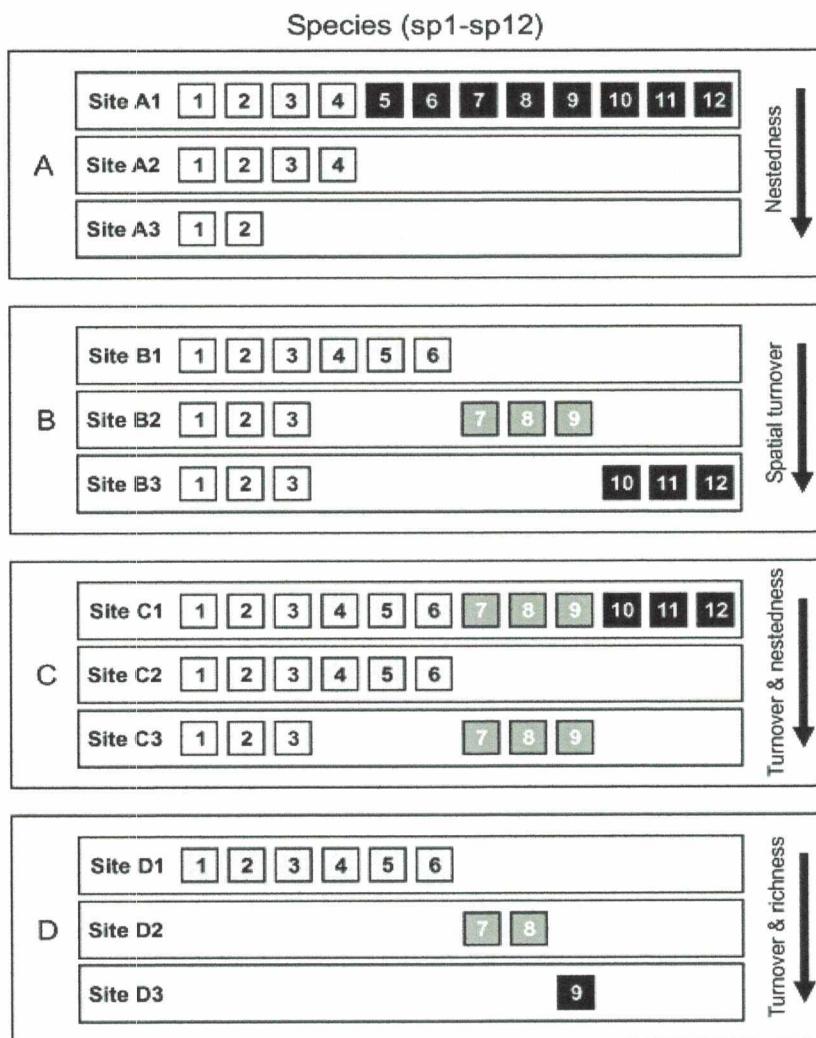


Figura 2 - Exemplos hipotéticos envolvendo quatro ilhas (A - D) e três sítios amostrais em cada. Biotas dos sítios A1 - A3 são completamente aninhadas, porque biotas pobres são subconjuntos das biotas ricas. Sítios B1 - B3 têm a mesma riqueza (seis espécies cada) com três espécies em comum a todos os três locais e três espécies exclusivas para cada sítio, ou seja, exibindo um padrão de turnover espacial. Sítios C1 - C3 ambos padrões presentes, porque C2 e C3 são subconjuntos de C1 (anhinhamento), mas algumas espécies são substituídas entre C2 e C3, que não são subconjuntos uma da outra. Sítios D1 - D3 turnover espacial e não são obviamente aninhada, mas apresentam diferenças na riqueza (BASELGA, 2010).

2.4 - Abelhas e vespas solitárias

As interações plantas-polinizadores são relações mutualísticas de extrema importância, ao coletar recursos florais como pólen, néctar e óleos, os polinizadores facilitam a reprodução das plantas, proporcionando a formação de frutos e sementes (GIANNINI *et al.*, 2015). As abelhas são os mais importantes e efetivos polinizadores (KLEIN *et al.*, 2007), são ecologicamente e funcionalmente diversas, forrageiam uma ampla variedade de formas florais e nidificam em diversos substratos (MICHENER, 2007).

Apesar de sua importância, os números de riqueza de abelhas são imprecisos por diversas razões como erros de identificação, sinonímias, homônimias, muitas áreas a serem amostradas (FREITAS *et al.*, 2009). Estima-se que o número de espécies de abelhas descritas está em torno de 18.000, mas que o número total pode ser em torno de 20.000 a 30.000 espécies (MICHENER, 2007) sendo que 85% são solitárias (SILVA *et al.*, 2001; BUSCHINI, 2006). A fêmea de espécies de abelhas solitárias constrói seu próprio ninho, abastece as células de cria sem ajuda de outras abelhas, e geralmente morre ou deixa o ninho antes da emersão de sua cria (MICHENER, 2007).

Quanto às vespas, cerca de 26.000 espécies foram descritas no mundo todo, sendo 90% consideradas de hábito solitário (LOYOLA & MARTINS, 2006; BUSCHINI & WOISKI, 2008; BATISTA, 2010). Várias destas espécies de vespas solitárias são consideradas inimigos naturais de artrópodes atuando no controle biológico natural de insetos (afídeos, ortópteros e lepidópteros) (TYLIANAKIS, *et al.*, 2005; HOLZSCHUH *et al.*, 2009, CAMPBELL, *et al.*, 2017), e outros artrópodes (baratas e aranhas) (EBELING *et al.* 2012, ROCHA-FILHO, *et al.*, 2017).

Vespas e abelhas solitárias podem nidificar em cavidades pré-existentes no solo, em troncos e galhos ocos ou perfurados, podendo ser estudadas por meio de ninhos-armadilha (TSCHARNTKE *et al.*, 1998; MORATO & CAMPOS, 2000; AGUIAR & MARTINS, 2002; MENDES & RÊGO, 2007). Ninhos-armadilha são espaços tubulares, os quais espécies de abelhas e vespas solitárias podem usar para nidificação (Figura 3) (TSCHARNTKE *et al.*, 1998; MORATO & CAMPOS, 2000; AGUIAR & MARTINS, 2002; TYLIANAKIS *et al.* 2005; KLEIN *et al.* 2007). Ninhos-armadilha são um sistema adequado para estudar interações multitróficas complexas devido à abelhas e vespas solitárias possuírem tamanho

pequeno, riqueza de espécies alta, complexidade trófica, e por sua importância no funcionamento do ecossistema (TSCHARNKE *et al.*, 1998; KLEIN *et al.* 2007). Cerca de 5% de todas as abelhas e vespas são potenciais espécies que nidificam em cavidades (Figuras 3, 4 e 5) (GATHMANN, 1998), e os recursos alimentares para ambas guildas dependem da diversidade de plantas presentes no habitat. Há evidências de que a abundância e a riqueza local das espécies de abelhas e vespas correlacionam-se com a diversidade de recursos alimentares como flores, néctar e presas.(TSCHARNKE *et al.*, 1998).

Segundo Morato & Martins (2006) abelhas e vespas solitárias gastam muito tempo de sua vida adulta construindo ninhos e coletando alimento para suas crias e, fatores que afetam essas atividades como local de nidificação e disponibilidade de alimento são por sua vez afetados pela estrutura ambiental. Consequentemente, mudanças na diversidade e abundância desses insetos podem afetar as relações do ecossistema através de seus papéis como predadores e polinizadores (TAKI *et al.* 2008). Algumas espécies de abelhas e vespas têm requerimentos de habitats específicos, para a construção de seus ninhos necessitam de cavidades pré-existentes, que por sua vez dependem da estrutura da vegetação (FLORES *et al.*, 2018). O período de nidificação é uma fase crítica no ciclo de vida de abelhas e vespas solitárias (MORATO & MARTINS, 2006; FLORES *et al.*, 2018) pois, esses himenópteros são sensíveis aos efeitos de perturbações ambientais (MORATO & CAMPOS, 2000) portanto, têm sido usadas como bioindicadores da qualidade do ambiente em diferentes ecossistemas (TYLIANAKIS *et al.*, 2005).



Figura 3- (A) Ninhos-armadilha instalados na RPPN do Caju; (B) ninhos fundados por abelhas (setas vermelhas) em cavidades de ninhos-armadilhas.

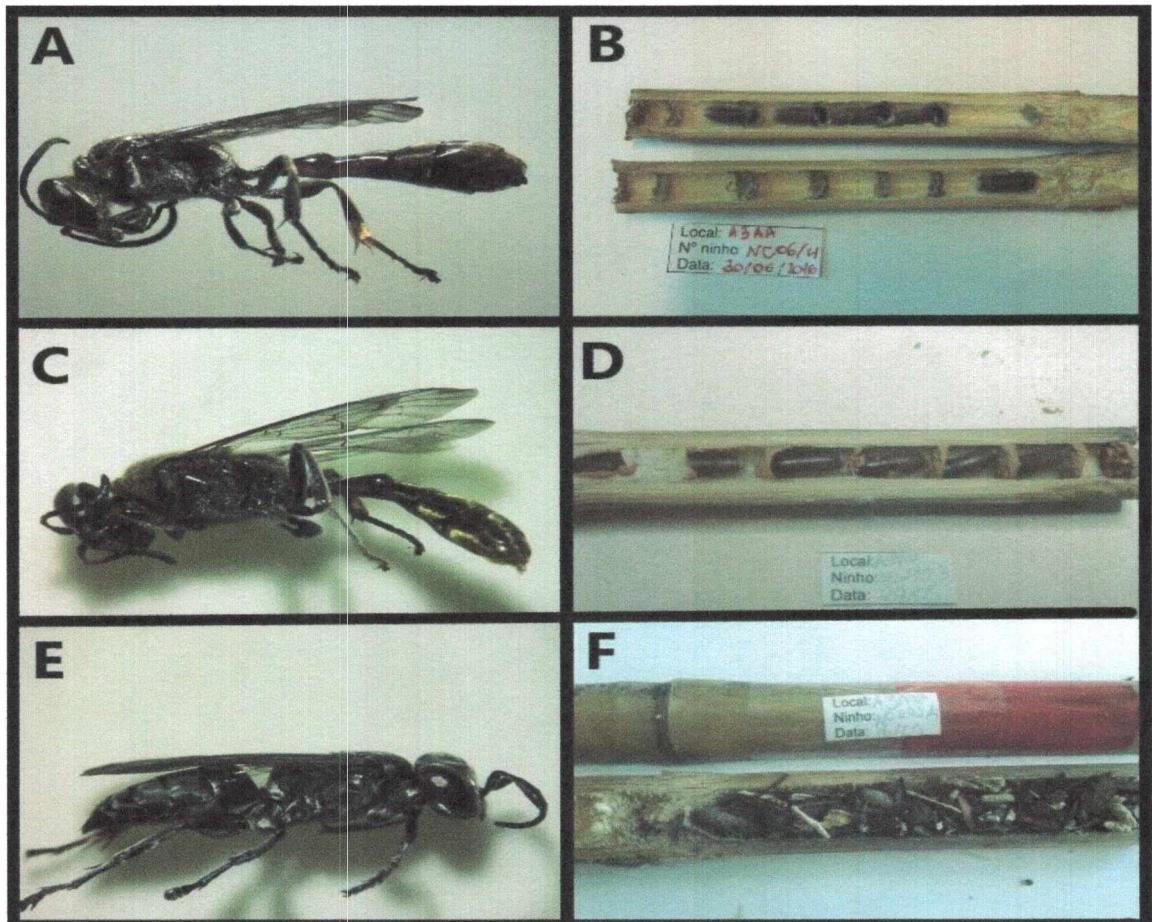


Figura 4 - Espécies e ninhos de vespas solitárias em vegetação de restinga na RPPN do Caju: Adulto (A) e ninho com pupas (B) de *Trypoxyylon aff. nitidum*; Adulto (C) e ninho fundado (D) por *T. ausuncicola*; Adulto (E) e ninho com pupas (F) de *Liris sp.*

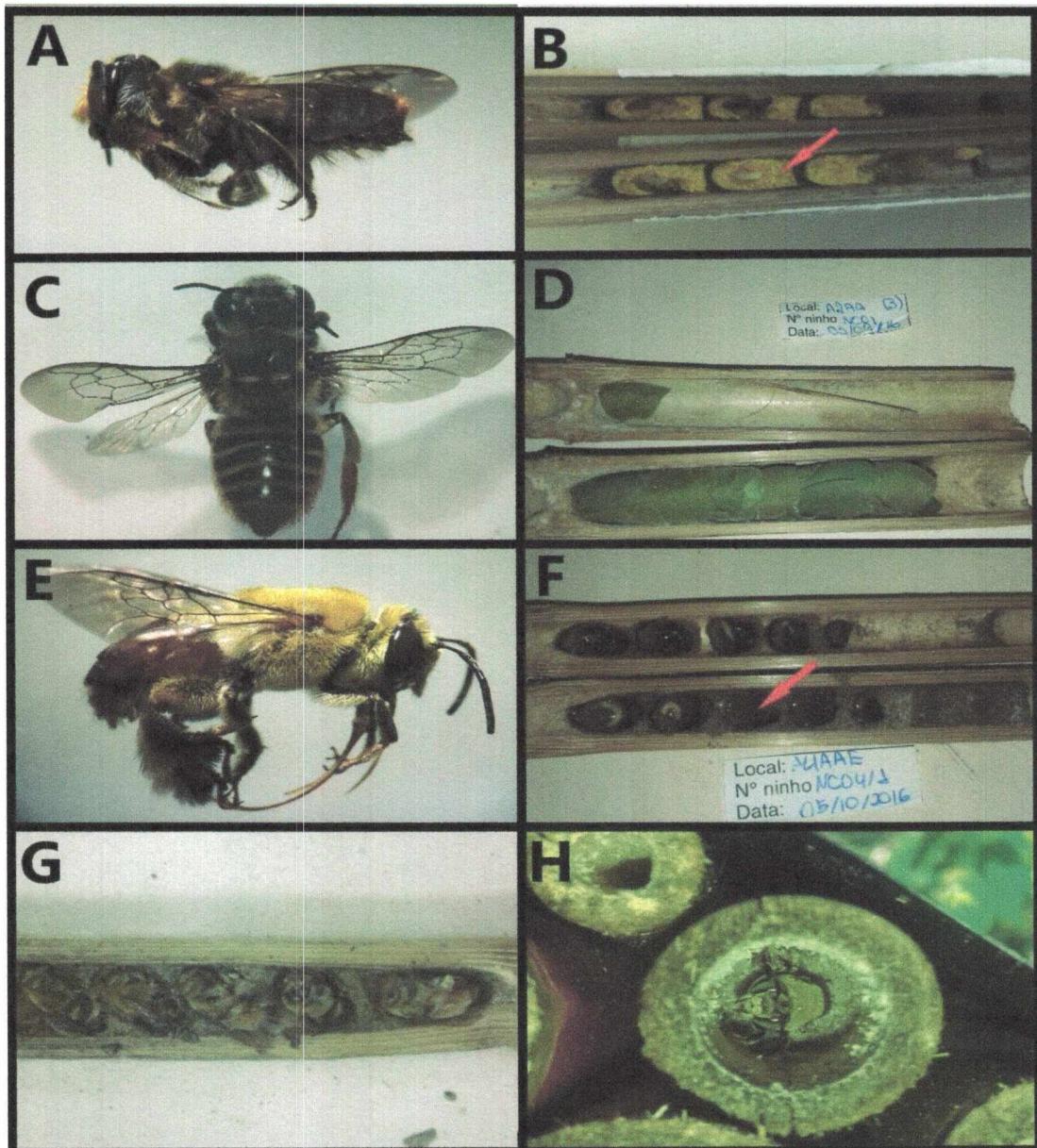


Figura 5- Espécies e ninhos de abelhas solitárias registrados em vegetação de restinga na RPPN do Caju: Adulto (A) e ninho (B) de *Centris* sp., é possível observar pólen disposto nas células como fonte de alimento para as larvas; Adulto (C) e ninho (D) de Megachilidae; Adulto (E) e ninho (F) de *C. tarsata*, nota-se a alimentação das larvas nas células, (G) Pré-adultos de *C. tarsata* dispostos nas células de cria antes da emergência; e (H) Ninho sendo construído por fêmea de *C. tarsata*.

2 - OBJETIVOS

2.1 Objetivo geral

Avaliar a composição de espécies da comunidade de vespas e abelhas solitárias em áreas abertas, vegetação arbustiva e secundária de restinga.

2.2– Objetivos específicos

- Investigar como as condições climáticas e o tipo de vegetação influenciam a nidificação e a composição de espécies da comunidade de vespas e abelhas solitárias em restinga.
- Investigar se a associação entre tipo de vegetação e condições climáticas influenciam na nidificação da abelha *Centris tarsata*, bem como nas características biológicas e de arquitetura dos ninhos dessa espécie.

3 - ESBOÇO DOS CAPÍTULOS

O presente estudo apresenta os resultados de uma pesquisa de dois anos com início em maio de 2016, com abelhas e vespas solitárias nidificantes em vegetação de restinga na RPPN do Caju no município de Itaporanga D'Ajuda, Sergipe.

Capítulo 2 – Aborda padrões de diversidade e estrutura da comunidade de espécies de abelhas e vespas solitárias que nidificam em ambiente de restinga por meio da participação da diversidade beta nos componentes substituição de espécies (turnover) e aninhamento. Foram abordados neste estudo os seguintes questionamentos: condições climáticas são fatores determinantes na riqueza e abundância de abelhas e vespas solitárias? A diversidade beta e seus componentes, substituição de espécies e aninhamento, são influenciados pelos diferentes tipos de vegetação? Os padrões de diversidade da comunidade de vespas e abelhas solitárias mudam ao longo do tempo?

Capítulo 3 - Investiga se a vegetação em mosaico de restinga e as condições climáticas regionais têm efeito sobre a nidificação de *Centris tarsata*, a abelha solitária mais abundante na região estudada abordando as seguintes questões: há uma associação entre as condições climáticas, tipo de vegetação e nidificação por *C. tarsata*? A arquitetura dos ninhos, características biológicas e o parasitismo por outros artrópodes variam com o tipo de vegetação? *Centris tarsata* apresenta preferências de nidificação nos diferentes estratos da vegetação?

4 - REFERÊNCIAS BIBLIOGRÁFICAS

AGUIAR, A. J. C.; MARTINS, C. F. Abelhas e vespas solitárias em ninhos-armadilha na Reserva Biológica Guaribas (Mamanguape, Paraíba, Brasil). **Revista Brasileira de Zoologia**, v. 19, n. p. 101–116, 2002.

BASELGA A. Partitioning the turnover and nestedness components of beta diversity. **Global Ecol. Biogeogr.** v.19, n.1, 2010, p. 134 – 143.

BASELGA A., ORME, C. D. L. Betapart: an R package for the study of beta diversity. **Methods in Ecology and Evolution**, v.3, n.1, 2012, p. 808 – 812.

BATISTA, M. C. **Resposta da comunidade de vespas e abelhas solitárias** (Insecta: Hymenoptera) ao uso da terra. Dissertação, UEMA. p. 100, 2010.

BUSCHINI, M. L. T. Species diversity and community structure in trap-nesting bees in southern Brazil. **Apidologie**, 37, 58–66, 2006.

BUSCHINI, M. L. T.; WOISKI, T. D. Alpha-beta diversity in trap-nesting wasps (Hymenoptera: Aculeata) in Southern Brazil. **Acta Zoologica**, v. 89, n. 4, p. 351–358, 2008.

CAMPBELL, J. W.; SMITHERS, C.; IRVIN, A.; KIMMEL, C. B.; STANLEY-STAHR, C.; DANIELS, J. C.; ELLIS, J. D.. Trap nesting wasps and bees in agriculture: a comparison of sown wildflower and fallow plots in Florida. **Insects**, v.8, n.4, p. 1-10, 2017.

EBELING, A.; KLEIN. A. M.; WEISSER, W.W.; TSCHARNTKE, T. Multitrophic effects of experimental changes in plant diversity on cavity-nesting bees, wasps, and their parasitoids. **Oecologia**, v. 169, n. 2, p. 453–465, 2012.

FLORES, L. M. A.; ZANETTE, L. R. S.; ARAUJO, F. S. Effects of habitat simplification on assemblages of cavity nesting bees and wasps in a semiarid neotropical conservation area. **Biodiversity and Conservation**, v. 27, n. 2, p. 311–328, 2018.

FREITAS, B. M.; IMPERATRIZ-FONSECA, V.L.; MEDINA, L. M.; PEIXOTO, A. M.; GALETTO, L.; NATES-PARRA, G.; QUEZADA-EUÁN, J.J.G. Diversity , threats and conservation of native bees in the Neotropics. **Apidologie** v. 40, p. 332–346, 2009.

GAVISH, Y., GILADI, I., ZIV, Y. Partitioning species and environmental diversity in fragmented landscapes: do the alpha, beta and gamma components match? **Biological and Conservation**, v. 28, 2019, p. 769 - 786. DOI <https://doi.org/10.1007/s10531-018-01691-7>

GATHMANN A. Bienen, Wespen und ihre Gegenspieler in der Agrarlandschaft: Artenreichtum und Interaktionen in Nisthilfen. Aktionsrunden und Habitatbewertung. Cuvillier, Göttingen, 1998.

GIANNINI, T.C.; BOFF,S.; CORDEIRO, G.D.; CARTOLANO JR, E.A.; VEIGA, A.K.; IMPERATRIZ-FONSECA, V.L.; SARAIVA, A.M. Crop pollinators in Brazil: a review of reported interactions. **Apidologie**, v. 46, n. 2, p. 209–223, 2015.

HOLZSCHUH, A.; STEFFAN-DEWENTER, I.; TSCHARNTKE, T. Grass strip corridors in agricultural landscapes enhance nest-site colonization by solitary wasps. **Ecological Applications**, v. 19, n. 1, p. 123–132, 2009.

KLEIN A.M., VAISSIÈRE, B.E., CANE J.H., STEFFAN-DEWENTER I., CUNNINGHAM S.A., KREMEN C., TSCHARNTKE T. Importance of pollinators in changing landscapes for world crops, **Proc. R. Soc.. London B, Biol. Sci.** 274, 303–313. 2007.

LANDIM, M.F.; SIQUEIRA, E.R. Caracterização florística e ecológica da Mata Alântica de Sergipe. In: SIQUEIRA, E. R.; RIBEIRO, F. E. (org.). **Mata Atlântica de Sergipe**. Aracaju: Embrapa Tabuleiros Costeiros, 2001, v.1., p.9-50.

LOYOLA, R. D.; MARTINS, R. P. Trap-nest occupation by solitary wasps and bees (Hymenoptera: Aculeata) in a forest urban remnant. **Neotropical Entomology**, v. 35, n. 1, p. 41–48, 2006.

MENDES, F. N.; RÉGO, M. M. C. Nidificação de *Centris* (Hemisiella) *tarsata* Smith

(Hymenoptera, Apidae, Centridini) em ninhos-armadilha no Nordeste do Maranhão, Brasil. **Revista Brasileira de Entomologia**, v. 51, n. 3, p. 382–388, 2007.

MICHENER, C. D. (2007). **The bees of the world**. The Johns Hopkins University Press, Baltimore.

MORATO, E. F.; CAMPOS, I. A. O. Efeitos da fragmentação florestal sobre vespas e abelhas solitárias na Amazônia Central. II. Estratificação vertical. **Revista Brasileira de Zoologia**, v. 17, n. 2, p. 429–444, 2000.

MORATO, E. F.; MARTINS, R. P. An overview of proximate factors affecting the nesting behavior of solitary wasps and bees (Hymenoptera: Aculeata) in preexisting cavities in wood. **Neotropical Entomology**, v. 35, n. 3, p. 285–298, 2006.

NOGUEIRA JUNIOR, L. R.; DOMPIERI, M. H. G.; RANGEL, M. S. A. RODRIGUES, R. F. A.; MELO, A. F. R. et al. **Plano de Manejo Reserva Particular do Patrimônio Natural do Caju**. Embrapa Tabuleiro Costeiros. Aracaju, 2013

OLIVEIRA, E. V. S.; LANDIM, M. F. Caracterização fitofisionômica das restingas da Reserva Biológica de Santa Isabel, litoral norte de Sergipe. **Scientia Plena**, v.10, n.10, p. 1 - 10, 2014

RIBEIRO, M. C., METZGER, J. P., MARTENSEN, A. C., PONZONI, F. J., HIROTA, M. M. The Brazilian Atlantic Forest: how much is left, and how is the remaining forest distributed? Implications for conservation. **Biological Conservation**, v. 142, n.1, p. 1141 - 1153, 2009.

ROCHA-FILHO, L. C., RABELO, L. S., AUGUSTO, S. C., GARÓFALO, C. A. Cavity-nesting bee and wasps (Hymenoptera: Aculeata) in a semi-deciduous Atlantic forest fragment immersed in a matrix of agricultural land. **J. Insect. Conserv.** 2017, v. 21, n.4, 2017, p. 727-736. DOI 10.1007/s10841-017-0016-x

RUBENE, D., SCHROEDER, M., RANIUS, T. Estimating bee and wasp (Hymenoptera: Aculeata) diversity on clear-cuts in forest landscapes: an evaluation of sampling methods.

Insect Conservation and Diversity, v. 8, n. 3, 2014 p. 261-271, DOI: 10.1111/icad.12105

SERRA, F. C. V.; LIMA, P. B.; ALMEIDA JR, E. B. DE. Species richness in restinga vegetation on the eastern Maranhão State, Northeastern Brazil. **Acta Amazonica**, v. 46, n. 3, p. 271–280, 2016.

SI, X., BASELGA, A., LEPRIEUR, F., SONG, X., DING, P. Selective extinction drives taxonomic and functional alpha and beta diversities in island bird assemblages. **Journal of Animal Ecology**, v. 85, n.2, 2016, p. 4019-418. DOI: 10.1111/1365-2656.12478

SILVA, S. M. **Diagnóstico das restingas do Brasil**. In: Fundação BIO RIO, Workshop Avaliação e ações prioritárias para a conservação da biodiversidade na Zona Costeira e Marinha (2000). Porto Seguro, Disponível em: http://rodadas.anp.gov.br/arquivos/Round7/arquivos_r7/PERFURACAO_R7/refere/Restingas.pdf . acesso em fev. 2019

SILVA, F. O.; VIANA, B. F.; NEVES, E. L. Biologia e arquitetura de ninhos de *Centris (Hemisiella) tarsata* Smith (Hymenoptera: Apidae: Centridini). **Neotropical Entomology**, v. 30, p. 541–545, 2001.

SOININEN, J., HEINO, J., WANG, J. A meta-analysis of nestedness and turnover components of beta diversity across organisms and ecosystems. **Global Ecol. Biogeogr.** v. 27, n.1, 2017, p.96-109. DOI: <https://doi.org/10.1111/geb.12660>

STANGLER, E. S.; HANSON, P. E.; STEFFAN-DEWENTER, I. Interactive effects of habitat fragmentation and microclimate on trap-nesting Hymenoptera and their trophic interactions in small secondary rainforest remnants. **Biodiversity and Conservation**, v. 24, n. 3, p. 563–577, 2015.

TAKI, H.; VIANA, B.F.; KEVAN, P.G.; SILVA, F.O.; BUCK, M. Does forest loss affect the communities of trap-nesting wasps (Hymenoptera: Aculeata) in forests? Landscape vs. local habitat conditions. **Journal of Insect Conservation**, v. 12, n. 1, p. 15–21, 2008.

TSCHARNTKE, T.; GATHMANN, A.; STEFFAN-DEWENTER, I. Bioindication using trap-nesting bees and wasps and their natural enemies: Community structure and interactions.

Journal of Applied Ecology, v. 35, n. 5, p. 708–719, 1998.

TYLIANAKIS, J. M; KLEIN, A.M.; TSCHARNTKE, T. Spatiotemporal variation in the diversity of Hymenoptera across a tropical habitat gradient. v. 86, n. 12, p. 3296–3302, 2005.

TUOMISTO, H. & RUOKOLAINEN, K. Analyzing or explaining beta diversity? Understanding the targets of different methods of analysis. **Ecology**, v.87, n.11, p. 2697–2708, 2006, . DOI: [https://doi.org/10.1890/0012-9658\(2006\)87\[2697:AOEBDU\]2.0.CO;2](https://doi.org/10.1890/0012-9658(2006)87[2697:AOEBDU]2.0.CO;2)

ULRICH, W. & GOTELLI, N.J. Disentangling community patterns of nestedness and species co-occurrence. **Oikos**, v. 11, n.12, p. 2053 - 2061, 2007.

WHITTAKER, R. H. Evolution and measurement of species diversity. **TAXON** v.2, n.2, 1972, p. 213 - 251.

ZICKEL, C.S.; VICENTE, A.; ALMEIDA Jr., E.B.; CANTARELLI, J.R.; SACRAMENTO, A.C. Flora e vegetação das restingas no Nordeste Brasileiro. In E. ESKINAZI-LEÇA, S. NEUMANN-LEITÃO; M.F. COSTA (eds.), Oceanografia: um cenário tropical. **Bargaço**. Recife, p. 689 – 701, 2004.

CAPÍTULO 2

Community composition of cavity-nesting bees and wasps according to restinga vegetation type

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1 **Community composition of cavity-nesting bees and wasps according to restinga
2 vegetation type**

3
4 **ABSTRACT**

5
6 Solitary bees and wasps are hymenopterans that nest in pre-existing cavities in natural
7 environments, and play a key role in ecosystem functioning providing important ecosystem
8 services such as pollination and natural biological control. Here, we aimed to assess the
9 species composition of cavity-nesting bees and wasps community in restinga vegetation, and
10 determine how climatic conditions influence nesting and assembly composition in three
11 vegetation types (open-areas, shrubby and secondary vegetation) of Brazilian restinga. The
12 hymenopterans were sampled with trap-nests and evaluations were conducted monthly over
13 two years. We collected 735 nests belonging to 4 bee species and 15 wasp species. Most
14 nesting occurred in open areas followed by secondary vegetation in the dry season, decreasing
15 in the rainy season. Species composition differed between open areas and secondary
16 vegetation. Overall, the contribution of beta diversity components, turnover and nestedness,
17 was similar among vegetation types. However, in the dry season, turnover was higher in
18 secondary and shrubby vegetations. Conversely, a completely nested pattern occurred in open
19 areas in both seasons. Turnover was the main responsible for beta diversity between the first
20 and second years, being high in secondary and shrubby vegetations. Accordingly, climatic
21 conditions influence species composition of cavity-nesting bees and wasps in restinga
22 vegetation favouring nesting activity in open areas and secondary vegetations. Furthermore,
23 secondary vegetation playing the role of shelter during warm and dry season, contributing for
24 the conservation of these essential ecosystem service providers.

25
26 **Key-words:** Atlantic Rainforest, beta diversity; climatic conditions; trap-nests

27

28 **1. INTRODUCTION**

29

30 Habitat fragmentation and land use intensification has led to the decline of global
31 biodiversity. The Brazilian Atlantic Forest biome, famed for its high biodiversity, has been
32 continuously exploited over the past 500 years and replaced by agriculture, livestock and
33 urban areas. The restinga is an ecosystem contained within the Atlantic forest biome, which is
34 characteristic of tropical and subtropical coastal regions, composed of herbaceous plants,
35 shrubs and an arboreal stratum growing in sandy plains under a marine influence. However,
36 this ecosystem is under constant threat as urban sprawl, real estate speculation and
37 deforestation directly affects the local flora and fauna (Fischer & Lindenmayer, 2007;
38 Oliveira & Landim, 2014; Rocha-Filho et al., 2017; Serra et al., 2016).

39 Bees and wasps are important in ecosystem functioning, play the role of providing
40 ecosystem services, are effective pollinators, ecologically and functionally diverse and they
41 forage a wide variety of floral plant species. Otherwise, several wasps species are considered
42 natural enemies of diverse arthropod species, thereby acting as natural biological control of
43 phytophagous species (Buschini & Woiski, 2008; Ebeling et al., 2012; Holzschuh et al., 2009;
44 Klein et al., 2007; Michener, 2007; Tylianakis et al., 2005).

45 Around 5% of all bees and wasps are potential cavity-nesting species that use hollow
46 trunks and branches to nest, and therefore can be studied using trap-nests (Aguiar & Martins,
47 2002; Mendes & Rêgo, 2007; Morato & Campos, 2000; Tscharntke et al., 1998, Veddeler et
48 al., 2010). These insects require pre-existing cavities to build their nests, which means that
49 they rely on the availability of particular vegetation structure and, therefore, they can be used
50 as indicators of environmental quality or changes (Flores et al., 2018; Tscharntke et al., 1998).
51 Understanding and preserving the habitats of cavity-nesting bees and wasps is important for

52 the maintenance of these populations of Aculeata (Morato & Martins, 2006). Therefore,
53 recognising diversity patterns and how biotic and abiotic factors influence the community
54 structure of these arthropods is essential for their management and conservation.

55 Beta diversity is defined as the variation in community composition between sites and
56 it may capture the dynamic nature of diversity patterns more thoroughly than measurements
57 of alpha diversity (species richness) (Soininen et al., 2017; Tuomisto & Ruokolainen, 2006).
58 Analysing beta diversity allows different hypotheses regarding the processes driving species
59 distribution to be tested. Therefore, Baselga (2010, 2012) proposed a method for partitioning
60 total dissimilarity (beta diversity) into two components accounting for the dissimilarities
61 derived from turnover and from nestedness. Turnover is the replacement of species in one site
62 with different species in another site. Nestedness is the species loss (or gain) resulting from
63 the elimination (or addition) of species in one location, making the poorer community a
64 subset of the richer (Baselga & Orme, 2012). In nature, species turnover may reflect species
65 segregation by environment or dispersal, while nesting is often related to ordered extinction-
66 colonization dynamics (Si et al., 2016; Soininen et al., 2017).

67 Here, we addressed the following questions: (a) whether climatic conditions influence
68 the nesting periods and vegetation type preferences of cavity-nesting bees and wasps; (b) what
69 the relative contributions are of turnover and nestedness to variation in the community
70 composition of cavity-nesting bees and wasps; (c) whether there is a relationship between
71 climatic conditions and turnover and nestedness for communities of cavity-nesting bees and
72 wasps within the restinga.

73 **2. METHODS**74 **2.1 Study Area**

76
77 The study was carried out in a protected area (763.37-ha) belonging to the Brazilian
78 Agricultural Research Corporation (Embrapa Tabuleiros Costeiros) in the municipality of
79 Itaporanga D'Ajuda ($11^{\circ} 06' 16.19''$ S, $37^{\circ} 11' 05.89''$ W), in the north-eastern Brazilian State
80 of Sergipe. According to the Köppen-Geiger classification, the predominant climate in the
81 region is tropical with marked dry (September to February) and rainy (March to August)
82 seasons, with average annual rainfall above 1250 mm. The biome is Atlantic rainforest
83 composed of a matrix of mangrove, restinga and apicum (sandy open areas influenced by the
84 tidal regime). The study was conducted within the restinga domain and rainfall and
85 temperature, respectively, were obtained from a meteorological station located in the area and
86 from the Instituto Nacional de Meteorologia (2018), on a per-month basis.

87

88 **2.2 Sampling**

89

90 Trap-nests were constructed from 15 pieces of hollow bamboo cane (20.0 cm long), with one
91 node in the middle, forming two internodes (9.0 to 11.0 cm deep). Bamboo canes had internal
92 diameters ranging from 0.3 to 2.0 cm evenly distributed (5 hollow bamboo canes of 0.3 to 0.8
93 cm, 5 of 0.9 to 1.4 cm and 5 of 1.5 to 2.0 cm diameter) and they were wrapped together with a
94 22.0 cm-wide plastic sheet and wire, making a total of 30 cavities (Batista Matos et al., 2012).
95 Trap-nests were set up in three vegetation types in the restinga: (a) open areas, with
96 dominance of grasses, such as *Paspalum maritimum* Trind., weeds, cactus (*Melocactus*
97 *zehntneri*, Britton & Rose) and some sparsely distributed trees, direct sunlight; (b) shrubby
98 vegetations, with dominance of shrubs and a few tree species, such as *Anacardium*

99 *occidentale* L., *Byrsonima crassifolia* (L.) Kunth, *Myrciaria tenella* (D.C.) O.Berg, grasses,
100 partial shading with high luminosity; and (c) secondary vegetations, with dominance of
101 medium to large tree species such as *A. occidentale*, *Hancornia speciosa* Gomes, *Tapirira*
102 *guianensis* Aublet and *Inga* sp., dense vegetations, soil covered by a thick litter layer, shaded
103 with low luminosity. Four replicates for each vegetation type were chosen, totalling 12
104 sampling areas. Study sites were usually clustered in groups of three vegetation types to avoid
105 spatial autocorrelation. Four trap-nests were hung by a wire at each study site and attached to
106 a tree branch, 1.5 m above the ground and distanced 1.0 m from each other. Sampling areas
107 were at least 100 m distant from each other. Entomological glue was applied to the wire
108 monthly to deter crawling arthropods. Trap-nests were inspected monthly for two years (from
109 July 2016 to June 2018) and nested bamboo canes were removed and replaced by empty ones
110 with similar internal diameter. Nested bamboo canes were placed in a plastic cage
111 individually and kept indoors at 27°C ($\pm 2^\circ\text{C}$) for three months. After adult emergence,
112 cavity-nesting bees and wasps were separated according to morphotype and identified using
113 Goulet & Huber (1993), Michener (2007) and DNA sequencing. Voucher specimens were
114 deposited in the collection of Laboratório de Entomologia, Embrapa Tabuleiros Costeiros,
115 Aracaju, Brazil, and Laboratório de Genética Evolutiva e de Himenópteros, Universidade
116 Federal de São Carlos, São Carlos, Brazil.
117

118 **2.3 Data Analysis**

119
120 Species richness was calculated using the estimators Chao1 and ACE (Abundance-based
121 Coverage Estimator) to verify the sufficiency of sampling effort using EstimateS software
122 version 9.1.0 (Colwell, 2013). Afterwards, sample-based accumulation curves were plotted
123 for observed and expected richness using SigmaPlot software version 12.5 (Systat Software,

124 2011). We also performed the Spearman correlation between observed and estimated richness
125 using the *ISwR* package (Dalgaard, 2015). Shannon-Wiener diversity (H') and Pielou's
126 evenness index (J') were calculated for cavity-nesting bees and wasps and submitted to
127 analysis of variance (ANOVA). Also, the number of singletons/doubletons in the vegetation
128 types was obtained using the *vegan* package (Oksanen et al., 2019). General linear models
129 (GLMs) with Poisson (discrete variables) error distributions were constructed in order to
130 assess differences among vegetation types for species richness. The models were subjected to
131 ANOVA, and *a posteriori* contrasts were carried out to assess differences among treatments
132 (Crawley 2007).

133 Regression tree analyses addressed how climatic conditions and vegetation type
134 influenced the nesting of bees and wasps. Temperature (°C), rainfall (mm³) and vegetation
135 type (open areas, shrubby and secondary vegetations) were treated as exploratory variables
136 and the number of nests built as the response variable. The root node represented total
137 sampling, which was further divided into two or more sets called decision nodes and into sub-
138 nodes and/or terminal nodes. Analyses were performed using the *rpart* package (Therneau et
139 al., 2019).

140 The contribution of beta diversity components was assessed using the dissimilarity
141 partition method (Baselga, 2010), which separates the total beta diversity (Sørensen
142 dissimilarity index) into two components (turnover and nestedness), using the *betapart*
143 package. The Sørensen dissimilarity index (presence/absence) is expressed by the formula:
144 $\beta_{SOR} = \beta_{SIM} + \beta_{NES}$, where β_{SIM} is the Simpson dissimilarity (the turnover component of
145 Sørensen dissimilarity) and β_{NES} is the nestedness component of Sørensen dissimilarity. In
146 order to establish the components of beta diversity between the first and the second year of
147 evaluation, we used the *beta.temp* function which calculates the dissimilarity values (turnover
148 and nestedness) between the combined sites of two data sets (x, y) describing presence and

149 absence at the same set of locations for two separate times. Also, we calculated pairwise
150 dissimilarity, which computes beta diversity and components using the *beta.pair* function
151 (Baselga & Orme, 2012). Unless otherwise stated, the analyses described above were
152 performed using R software version 3.3.1 (R Development Core Team 2016). Additionally,
153 we used the Mantel test to explore the possible relationship between spatial distance and
154 species dissimilarity, using Jaccard distance (presence/absence) for bees' and wasps' species
155 composition and Euclidean distance for the spatial distance matrix, using Pearson's
156 correlation and 999 permutations (Perillo et al., 2017).

157 Non-metric multidimensional scaling (NMDS) analysis was performed to verify
158 differences in the community composition of cavity-nesting bees and wasps among vegetation
159 types using the Bray-Curtis similarity index (abundance). Moreover, similarity analyses
160 (ANOSIM) were conducted to compare the differences between two or more groups of
161 vegetation using the Bray-Curtis similarity index and Bonferroni correction at the 5% level.
162 Analyses were performed using the software PAST (Hammer et al., 2001).

163 3. RESULTS

164

165 A total of 735 nests and 2.147 individuals were collected, of which 1.353 were wasps,
166 comprising 15 species and/or morphospecies (Table 1), belonging to the families Crabronidae
167 and Sphecidae. The remaining 794 individuals that emerged were bees, encompassing 4
168 species from the families Apidae and Megachilidae (Table 1).

169 In open areas, we found 607 individuals from 11 wasp species and 441 from 4 bee
170 species. In shrubby vegetations, there were 328 individuals belonging to 5 species of wasps
171 and 277 individuals from 3 bee species. In secondary vegetations 418 individuals of 10 wasp
172 species and 76 comprising 3 bee species were found. Secondary vegetations also presented

173 more singletons (7) compared to open areas (5) and shrubby vegetations (4), whereas
174 doubletons were found only in open areas and secondary vegetations (2 and 1, respectively).

175 Sample-based accumulation curves show that the observed species richness of bees
176 and wasps increased with the number of individuals (Figure 1). Observed and estimated
177 species richness were highly correlated with the Chao1 and ACE estimators (Spearman
178 correlation; $\rho = 0.99$, $P < 0.0001$), with 97.56% and 91.86% efficiency, respectively,
179 suggesting that sampling effort was sufficient. However, the Shannon-Wiener (H') and Pielou
180 (J') indexes did not differ among vegetation types (open areas: $H' = 1.64$; $J' = 0.74$; shrubby
181 vegetations: $H' = 1.28$; $J' = 0.74$; secondary vegetations: $H' = 1.21$; $J' = 0.71$; $P > 0.05$). In
182 contrast, vegetation types influenced species richness, in that open areas had more species in
183 comparison with shrubby and secondary vegetations ($\chi^2 = 26.51$, $df = 1$, $P < 0.0001$; $\chi^2 =$
184 20.76, $df = 1$, $P < 0.001$, respectively).

185 Cavity-nesting bees and wasps preferred to nest in open areas during the dry season,
186 followed by shrubby and secondary vegetations. Nesting peaked from November to
187 December 2016 and from November 2017 to June 2018 (Figure 2). The regression tree
188 revealed 5 levels of depth and 15 nodes, these being 7 decision nodes and 8 terminals. Two
189 groups were formed according to temperature and vegetation types. In the first group, nesting
190 was favoured at temperatures below 27°C in shrubby and secondary vegetations, with 0.25
191 nests in 33% of the samples. On the other hand, in open areas, for temperatures below 27°C
192 and rainfall lower than 148 mm³ the mean number of nests was 0.72 in 11% of the samples.
193 However, when rainfall increased, average nesting was 1.8 in 6% of the samples. In the
194 second group (temperature ≥ 27 °C), the mean number of nests was 1.2 in 33% of the samples
195 in shrubby and secondary vegetations. Conversely, nesting in open areas was favoured in
196 temperatures below 28°C. Dry periods with rainfall ranging from 41 to 132 mm³, led to a

197 mean number of 3.4 nests in 4% of the samples. For rainfall below 41 mm³, the mean number
198 of nests was 1.4 in 6% of the samples (Figure 3).

199 Overall, the components of beta diversity indicated that species turnover and
200 nestedness contributed similarly (51.4% and 48.6%, respectively), which was also true during
201 the dry season (57.1%, and 42.9%, respectively) (Figure 4A). However, during the rainy
202 season, nestedness (65.8%) overrode species turnover (34.2%) (Figure 4B). Between the first
203 and second years of evaluation, higher bee and wasp species turnover was recorded in
204 secondary vegetations (84.4%), followed by shrubby vegetations (70.8%), while in open areas
205 (53.3%) species composition variation matched the nestedness pattern (46.7%) (Table 2).

206 According to the pairwise dissimilarity, shrubby vegetations was completely nested
207 within open areas (100%). Secondary vegetations and open areas had high species turnover
208 (82.5%) while between shrubby and secondary vegetations (62.5%) was lower. During both
209 the dry and rainy seasons, open areas and shrubby vegetations remained completely nested
210 (100%), unlike open areas and secondary vegetations where nestedness was 13.2 % during the
211 dry season and 63.6 % in the rainy season. Conversely, during the dry season species turnover
212 between open areas and secondary vegetations was high (86.1%), as well as during the rainy
213 season between shrubby and secondary vegetations (57.1%) (Table 2). The Mantel test
214 showed that there was no relationship between spatial distance and species dissimilarity
215 (Pearson correlation; $r = -0.09013$; $P > 0.05$).

216 The NMDS revealed that open areas and secondary vegetations were dissimilar in
217 their cavity-nesting bee and wasp community composition, while shrubby and secondary
218 vegetations, and open areas and shrubby vegetations, were similar (stress = 0.13, Figure 5).
219 However, according to ANOSIM, there was no difference in the rank similarities for the 12
220 sites in the three vegetation types ($P > 0.05$).
221

222 **4. DISCUSSION**

223

224 Our results showed that wasp species richness was higher than bee species richness for
225 the three restinga vegetation types, which has also been observed in tropical regions in
226 Indonesia, Costa Rica, Amazonia and north-east Brazil. This pattern may be due to bees being
227 more sensitive than wasps to habitat modification (Batista Matos et al., 2012; Klein et al.,
228 2002; Morato & Campos, 2000; Stangler et al., 2015).

229 Open areas harboured greater species richness and diversity of cavity-nesting bees and
230 wasps, followed by secondary vegetations, possibly due to the influence of generalist species
231 on the overall species richness. Generalist species may benefit from a wide variety of
232 resources and thus are prone to be less sensitive to simplified habitats (Flores et al., 2018). In
233 addition, open areas may provide greater availability of resources such as exposed nest
234 building materials like clay, sand grains, resins, leaves, filaments, twigs, and organic matter,
235 as well as food resources such as pollen, nectar, floral oils and prey (Loyola & Martins, 2006;
236 Morato & Martins, 2006).

237 Overall, the nesting of cavity-nesting bees and wasps seems to be affected by climatic
238 conditions, as it was higher during the dry season in open areas, followed by secondary
239 vegetations. As rainfall increased, nesting decreased, except for in shrubby vegetations, where
240 nesting was higher during the rainy season. Nesting was lower during the first year compared
241 to the second, possibly due to higher rainfall during the latter (INMET, 2018).

242 However, the regression tree showed that there are ranges of temperature and rainfall
243 that may be beneficial to cavity-nesting bees and wasps nesting. Nesting in shrubby and
244 secondary vegetations was favoured by higher temperatures in drier periods, while in open
245 areas were temperatures below 28°C and rainfall up to 132 mm³. Seasonal changes in
246 temperature and relative humidity can affect the population density of several arthropod

247 species, including cavity-nesting bees and wasps, and play an important role in determining
248 species' preference for a given habitat, (Buschini et al., 2006; Matos et al., 2016; Philpott et
249 al., 2006; Teodoro et al., 2009; Tylianakis et al., 2005). Additionally, climatic conditions can
250 directly affect resource availability, and the higher humidity inside forests can negatively
251 influence the activity of cavity-nesting bees and wasps due to their limited ability to
252 thermoregulate (Loyola & Martins, 2006; Stangler et al., 2015).

253 In a study on differential habitat occupation by *Trypoxylon* spp. wasps, Santoni et al.
254 (2009) observed a synchronic temporal occurrence of these species, especially in the warm
255 and rainy season. According to these authors, the influence of seasonality on nesting biology
256 of neotropical cavity-nesting species is possibly associated with monthly average temperature.
257 Moreover, in a study of *Trypoxylon lactitarse* from southern Brazil (pronounced dry season,
258 mild summers and frosts in the winter), Buschini & Wolff (2006) observed that this species
259 started nesting in forested areas during spring and summer and switched to grassland and
260 swamps during autumn and winter. However, *Centris tarsata*, the most abundant oil-
261 collecting bee in the present study, showed a preference for nesting during the dry season, a
262 pattern observed in several studies in different regions of Brazil (Aguiar & Martins, 2002;
263 Buschini & Wolff, 2006; Mendes & Rêgo, 2007; Costa et al., 2019).

264 The species composition was dissimilar between open areas and secondary
265 vegetations, endorsing results obtained in other regions of Brazil that species composition of
266 cavity-nesting bee and wasp assemblages varies along a habitat complexity gradient (Buschini
267 & Woiski, 2008; Flores et al., 2018).

268 Total beta diversity among vegetation types was low and nestedness and species
269 turnover made similar contributions. However, when it was calculated for each season,
270 different contributions from turnover and nestedness were observed. Species replacement
271 (turnover) was favoured during the dry season, unlike the rainy season, when species richness

272 was low and the poorer cavity-nesting bees and wasps community became nested within the
273 richer. Regardless of season, the community of cavity-nesting bee and wasp species from
274 shrubby vegetation was completed nested within that of the open areas and secondary
275 vegetations, but, in the latter case only during the dry season. Additionally, species
276 replacement was higher than nestedness for the two years, possibly due to the high rainfall
277 recorded during the second year of sampling.

278 The relationship between precipitation and variation in fauna composition may explain
279 the covariation between species richness and precipitation variation (Baselga, 2008). A study
280 of the structure of bee and wasp diversity in burned and unburned areas in clear-cut forest in
281 Sweden demonstrated that 72% to 90% of beta diversity was explained by species turnover,
282 while nestedness explained only 10% to 28% (Rubene et al., 2015). Furthermore, Perillo et al.
283 (2017), investigating the species composition of the bee and wasp community over an altitude
284 gradient in the neotropical region, found that beta diversity partitioning was mainly influenced
285 by species turnover (81%). According to these authors, species turnover was influenced by the
286 high rate of singletons and doubletons, a pattern commonly seen in tropical arthropod studies.
287 In fact, in our study, secondary vegetations presented a higher number of singletons and
288 doubletons compared to the other vegetation types, and that environment had species turnover
289 as the most representative component of beta diversity.

290 Our results indicate that solitary wasps and bees in the restinga respond strongly to
291 climatic conditions and vegetation type. These species prefer to nest during the warmer dry
292 season in open areas and secondary vegetations. Although the variation in species
293 composition was low for open areas, shrubby and secondary vegetations, the patterns that
294 drive species distribution varied according to vegetation type and climatic conditions.
295 Therefore, understanding the species composition of cavity-nesting bees and wasps is crucial
296 for the management and conservation of these essential ecosystem service providers.

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298

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306 **REFERENCES**

307

- 308 Aguiar, A. J. C.; Martins, C. F. (2002). Abelhas e vespas solitárias em ninhos-armadilha na
309 Reserva Biológica Guaribas (Mamanguape, Paraíba, Brasil). *Rev. Bras. Zool.* 19(1),
310 101-116.
- 311 Baselga, A., 2008. Determinants of species richness, endemism and turnover in European
312 longhorn beetles. *Ecography* 31(2), 263-271.
- 313 Baselga, A. (2010). Partitioning the turnover and nestedness components of beta
314 diversity. *Global Ecol. Biogeogr.* 19(1), 134-143.
- 315 Baselga, A., (2012). The relationship between species replacement, dissimilarity derived from
316 nestedness, and nestedness. *Global Ecol. Biogeogr.* 21(12), 1223-1232.
- 317 Baselga, A., Orme, C. D. L. (2012). Betapart: an R package for the study of beta diversity.
318 *Methods in Ecology and Evolution*, 3(5), 808-812.
- 319 Batista Matos, M. C., Souza-Souto, L., Almeida, R. S., Teodoro, A. V. (2012). Contrasting
320 patterns of species richness and composition of solitary wasps and bees (Insecta:
321 Hymenoptera) according to land-use. *Biotropica* 45(1), 73-79.

- 322 Buschini, M. L. T. (2006). Species diversity and community structure in trap-nesting bees in
323 southern Brazil. *Apidologie*, 37, 58–66.
- 324 Buschini, M. L. T.; Wolff, L. L. (2006). Nesting biology of *Centris (Hemisiella) tarsata*
325 Smith in southern Brazil (Hymenoptera, Apidae, Centridini). *Braz. J. Biol.* 66(4)
326 1091-1101.
- 327 Buschini, M. L. T.; Woiski, T. D. (2008). Alpha-beta diversity in trap-nesting wasps
328 (Hymenoptera: Aculeata) in Southern Brazil. *Acta Zoologica*, 89(4), 351-358.
- 329 Colwell, R.K. (2013). EstimateS: Statistical estimation of species richness and shared species
330 from samples. Version 9.1.0.
- 331 Crawley, M.J. (2007). The R book. John Wiley & Sons Ltd, West Sussex.
- 332 Costa, D. M., Batista, M. C., Brito, A. S. Barros, I. Teodoro, A. V. (2019). Rainfall,
333 temperature and vegetation type influence nesting by the oil-collecting bee *Centris*
334 (*Hemisiella*) *tarsata* in Brazilian restinga. *Apidologie*.
335 <https://doi.org/10.1007/s13592-019-00688-7>
- 336 Dalgaard, P. (2015). Introductory statistics with R. Springer Verlag. <https://cran.r-project.org/web/packages/ISwR/ISwR.pdf>
- 338 Ebeling, A., Klein, A.M., Weisser, W. W., Tscharntke, T. (2012). Multitrophic effects of
339 experimental changes in plant diversity on cavity-nesting bees, wasps, and their
340 parasitoids. *Oecologia*, 169(2), 453–465.
- 341 Fischer, J., Lindenmayer, D. B. (2007). Landscape modification and habitat fragmentation: a
342 synthesis. *Global Ecol. Biogeogr.* 16(3), 265-280.
- 343 Flores, L. M. A.; Zanette, L. R. S.; Araujo, F. S. (2018). Effects of habitat simplification on
344 assemblages of cavity nesting bees and wasps in a semiarid neotropical conservation
345 area. *Biodiversity and Conservation*, 27(2),: 311–328.

- 346 Goulet, H., Huber, J. T. (1993). Hymenoptera of the world: An identification guide to
347 families. Research Branch, Agricultural Canada Publication. Canada Communication
348 Group-Publishing, Ottawa.
- 349 Hammer, Ø., Harper, D. A. T., Ryan, P. D. (2001). PAST: Paleontological software package
350 for education and data analysis. *Palaentologia Electronica*, 4(1), 1-9.
- 351 Holzschuh, A.; Steffan-Dewenter, I.; Tscharntke, T. (2009). Grass strip corridors in
352 agricultural landscapes enhance nest-site colonization by solitary wasps. *Ecological
353 Applications*, 19(1), 123–132.
- 354 Instituto Nacional de Meteorologia, (2018). Dados Climatológicos de 2016 a 2018. 2018.
355 Digital Repository. <http://www.inmet.gov.br/portal/index.php?r=bdmep/bdmep>
- 356 Klein, A. M., Steffan-Dewenter, I., Buchori, D., Tscharntke, T. (2002). Effects of land-use
357 intensity in tropical agroforestry systems on coffee flower-visiting and trap-nesting
358 bees and wasps. *Conservation Biology*, 16(4), 1003-1014
- 359 Klein A.M., Vaissière B.E., Cane J.H., Steffan-Dewenter., Cunningham S.A., Kremen C.,
360 Tscharntke T. (2007). Importance of pollinators in changing landscapes for world
361 crops. *Proc. R. Soc.. London B, Biol. Sci.* 274(1608), 303–313.
- 362 Loyola, R. D.; Martins, R. P. (2006). Trap-nest occupation by solitary wasps and bees
363 (Hymenoptera: Aculeata) in a forest urban remnant. *Neotropical Entomology*, 35(1),
364 41–48.
- 365 Matos, M. C., Silva, S. S., Teodoro, A. V. (2016). Seasonal population abundance of the
366 assembly of solitary wasps and bees (Hymenoptera) according to land-use in
367 Maranhão state, Brazil. *Rev. Bras. Entomol.* 60(2), 171-176.
- 368 Mendes, F. N.; Rêgo, M. M. C. (2007). Nidificação de *Centris (Hemisiella) tarsata* Smith
369 (Hymenoptera, Apidae, Centridini) em ninhos-armadilha no Nordeste do Maranhão.
370 *Brasil. Rev. Bras. Entomol.* 51(3), 382-388

- 371 Michener, C. D. (2007). The bees of the world. The Johns Hopkins University Press,
372 Baltimore.
- 373 Morato; E.F.; Campos, L.A.O. (2000). Efeitos da fragmentação florestal sobre vespas e
374 abelhas solitárias na Amazônia Central. *Rev. Bras. Zool.* 17(2), 429-444.
- 375 Morato, E. F.; Martins, R. P. (2006). An overview of proximate factors affecting the nesting
376 behavior of solitary wasps and bees (Hymenoptera: Aculeata) in preexisting cavities in
377 wood. *Neotropical Entomol.* 35(3), 285–298.
- 378 Oksanen, J., Blanchet, F. G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D.,...Wagner,
379 H. (2019). Community ecology package. <https://cran.r-project.org/web/packages/vegan/vegan.pdf>
- 380 Oliveira, E. V. S.; Landim, M. F. (2014). Caracterização fitofisionômica das restingas da
381 Reserva Biológica de Santa Isabel, litoral norte de Sergipe. *Scientia Plena*, 10, 1-10.
- 382 Perillo, L. N., Neves, F. S., Antonini, Y., Martins, R. P.(2017). Compositional changes in bee
383 and wasp communities along Neotropical mountain altitudinal gradient. *Plos One*.
- 384 Philpott, S. M., Perfecto, I., Vandermeer, J. (2006) Effects of management intensity and
385 season on arboreal ant diversity and abundance in coffee agroecosystems. *Biodiversity
386 and Conservation* 15(1), 139-155.
- 387 R Core Team (2016). R: A language and environment for statistical computing. R Foundation
388 for Statistical Computing, Vienna, Austria. <https://www.R-project.org/>.
- 389 Rubene , D., Schroeder, M., Ranius, T. (2015). Diversity patterns of wild bees and wasps in
390 managed boreal forests: effects of spatial structure, local habitat and surrounding
391 landscape. *Biological Conservation* 184, 201-2018.
- 392 Rocha-Filho, L. C., Rabelo, L. S., Augusto, S. C., Garófalo, C. A. (2017). Cavity-nesting bees
393 and wasps (Hymenoptera: Aculeata) in a semi-deciduous Atlantic forest fragment
394 immersed in a matrix of agricultural land. *J. Insect. Conserv.* 21(4) 727-736.

- 396 Santoni, M., M., Brescovit, A. D., Del Lama, M. A. (2009). Ocupação diferencial do habitat
397 por vespas do gênero Trypoxylon (*Trypargilum*) Latreille (Hymenoptera,
398 Crabronidae). *Rev. Bras. Entomol.* 53(1), 107-114.
- 399 Serra, F. C. V.; Lima, P. B.; Almeida Jr., E. B. (2016). Species richness in restinga vegetation
400 on the eastern Maranhão State, Northeastern Brazil. *Acta Amaz.* 46 (3), 271–280.
- 401 Si, X., Baselga, A., Leprieur, F., Song, X., Ding, P. (2016). Selective extinction drives
402 taxonomic and functional alpha and beta diversities in island bird assemblages.
403 *Journal of Animal Ecology*, 85(2), 4019-418.
- 404 Soininen, J., Heino, J., Wang, J. (2017). A meta-analysis of nestedness and turnover
405 components of beta diversity across organisms and ecosystems. *Global Ecol.*
406 *Biogeogr.* 27(1),96-109.
- 407 Stangler, E. S.; Hanson, P. E.; Steffan-Dewenter, I. (2015). Interactive effects of habitat
408 fragmentation and microclimate on trap-nesting Hymenoptera and their trophic
409 interactions in small secondary rainforest remnants. *Biodiversity and Conservation*,
410 24(3), 563–577.
- 411 Systat Software Inc. (2011). Version 12.5. San Jose California.
- 412 Teodoro, A. V., Klein, A. M., Tscharntke, T. (2009). Temporally mediated responses of the
413 diversity of coffee mites to agroforestry management. *J. Appl. Entomol.* 133, 659-665.
- 414 Therneau, T., Atkinson, B., Ripley, B. (2019). Recursive partitioning and regression trees.
415 <https://cran.r-project.org/web/packages/rpart/rpart.pdf>
- 416 Tscharntke, T.; Gathmann, A.; Steffan-Dewenter I. (1998). Bioindication using trap-nesting
417 bees and wasps and their natural enemies: Community structure and interactions.
418 *Journal of Applied Ecology*, 35(5), 708–719.

- 419 Tuomisto, H. & Ruokolainen, K. (2006). Analyzing or explaining beta diversity?
420 Understanding the targets of different methods of analysis. *Ecology*, 87(11), 2697–
421 2708.
- 422 Tylianakis, J. M; Klein, A.M.; Tscharntke, T. (2005). Spatiotemporal variation in the diversity
423 of Hymenoptera across a tropical habitat gradient. *Ecology*, 86(12), 3296–3302.
- 424 Veddeler, D., Tylianakis, J., Tscharntke, T., Klein, A. M. (2010). Natural enemy diversity
425 reduces temporal variability in wasp but not bee parasitism. *Oecologia*, 162(3), 755–
426 762.
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Table 1 - Total number of cavity-nesting bees (Apidae and Megachilidae) and wasps (Sphecidae and Crabronidae) in open areas, shrubby and secondary vegetations in restinga domain in Caju RPPN, Itaporanga D'Ajuda, Sergipe.

Family	Morphospecies	Nests founded	Total emerged adults		
			Open area	Shrubby vegetation	Secondary vegetation
Sphecidae	Sphecidae 1	15	15	25	5
	Sphecidae 2	1	0	0	2
	Sphecidae 3	2	10	0	0
	<i>Liris</i> sp	87	69	44	31
	<i>Isodontia</i> sp 1	1	1	0	0
	<i>Isodontia</i> sp2	1	1	0	0
	<i>Isodontia</i> sp3	2	0	0	2
	<i>Isodontia</i> sp4	8	7	0	0
Crabronidae	<i>Trypoxylon lactitarse</i>	71	16	0	46
	<i>Trypoxylon nitidum</i>	199	238	220	280
	<i>Trypoxylon ausuncicola</i>	47	119	9	17
	<i>Trypoxylon</i> sp	83	94	30	27
	<i>Trypoxylon</i> sp1	1	0	0	6
	<i>Trypoxylon</i> sp2	9	37	0	0
	<i>Trypoxylon</i> sp 3	2	0	0	2
Apidae	<i>Centris tarsata</i>	162	363	217	61
	<i>Centris</i> sp 1	1	1	0	0
Megachilidae	<i>Megachile</i> sp	27	33	40	8
	Anthidiini	16	44	20	7
Total		735	1048	605	494

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Table 2 - Proportion of Beta diversity components for cavity-nesting bee and wasp community in restinga domain. Bold numbers indicate contribution over 50% in beta diversity. β SIM (Simpson dissimilarity, i.e. species turnover), β NES (nestedness) and β SOR Sørensen dissimilarity, i.e. total beta diversity).

Solitary bees and wasps communities	Beta diversity components		Total beta diversity
	β SIM	β SNE	β SOR
First and second year of sampling			
Open areas	0.22 (53.3%)	0.19 (46.7%)	0.42 (100%)
Shrubby vegetation	0.12 (70.83%)	0.05 (29.17%)	0.17 (100%)
Secondary vegetation	0.37 (84.37%)	0.07 (15.65%)	0.44 (100%)
Pairwise among vegetation types			
Open areas – Shrubby vegetation	0 (0%)	0.30 (100%)	0.30 (100%)
Open areas – Secondary vegetation	0.33 (82.5%)	0.07 (17.5%)	0.40 (100%)
Shrubby vegetation - Secondary vegetation	0.25 (62.5%)	0.15 (37.5%)	0.40 (100%)
Pairwise among vegetation for dry season			
Open areas – Shrubby vegetation	0 (0%)	0.25 (100%)	0.25 (100%)
Open areas – Secondary vegetation	0.33 (86.1%)	0.049 (13.2%)	0.36 (100%)
Shrubby vegetation - Secondary vegetation	0 (0%)	0.18 (100%)	0.18 (100%)
Pairwise among vegetation for rainy season			
Open areas – Shrubby vegetation	0 (0%)	0.20 (100%)	0.20 (100%)
Open areas – Secondary vegetation	0.16 (36.4%)	0.27(63.6%)	0.44 (100%)
Shrubby vegetation - Secondary vegetation	0.16 (57.1%)	0.12 (42.9%)	0.28 (100%)

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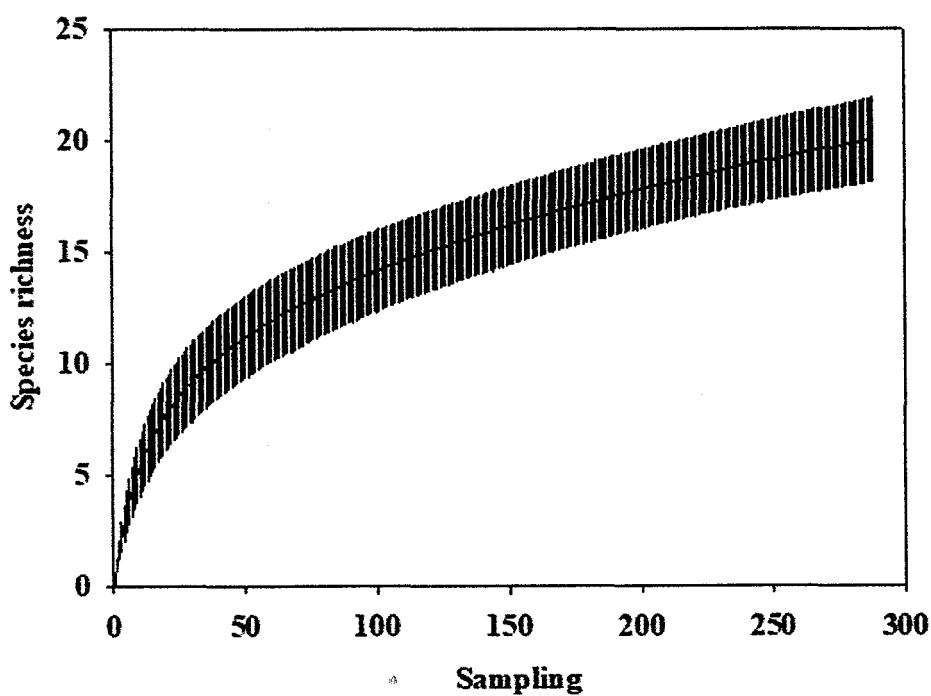
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Figure 1 - Sample-based species accumulation curve for cavity-nesting bees and wasps in restinga domain. Error bars denote 95% *CI*.

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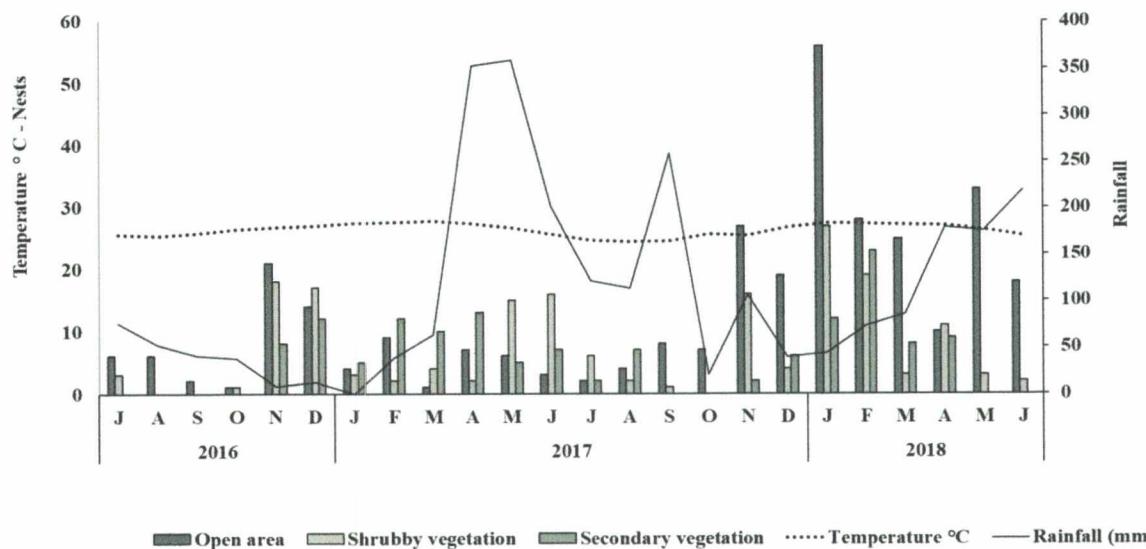


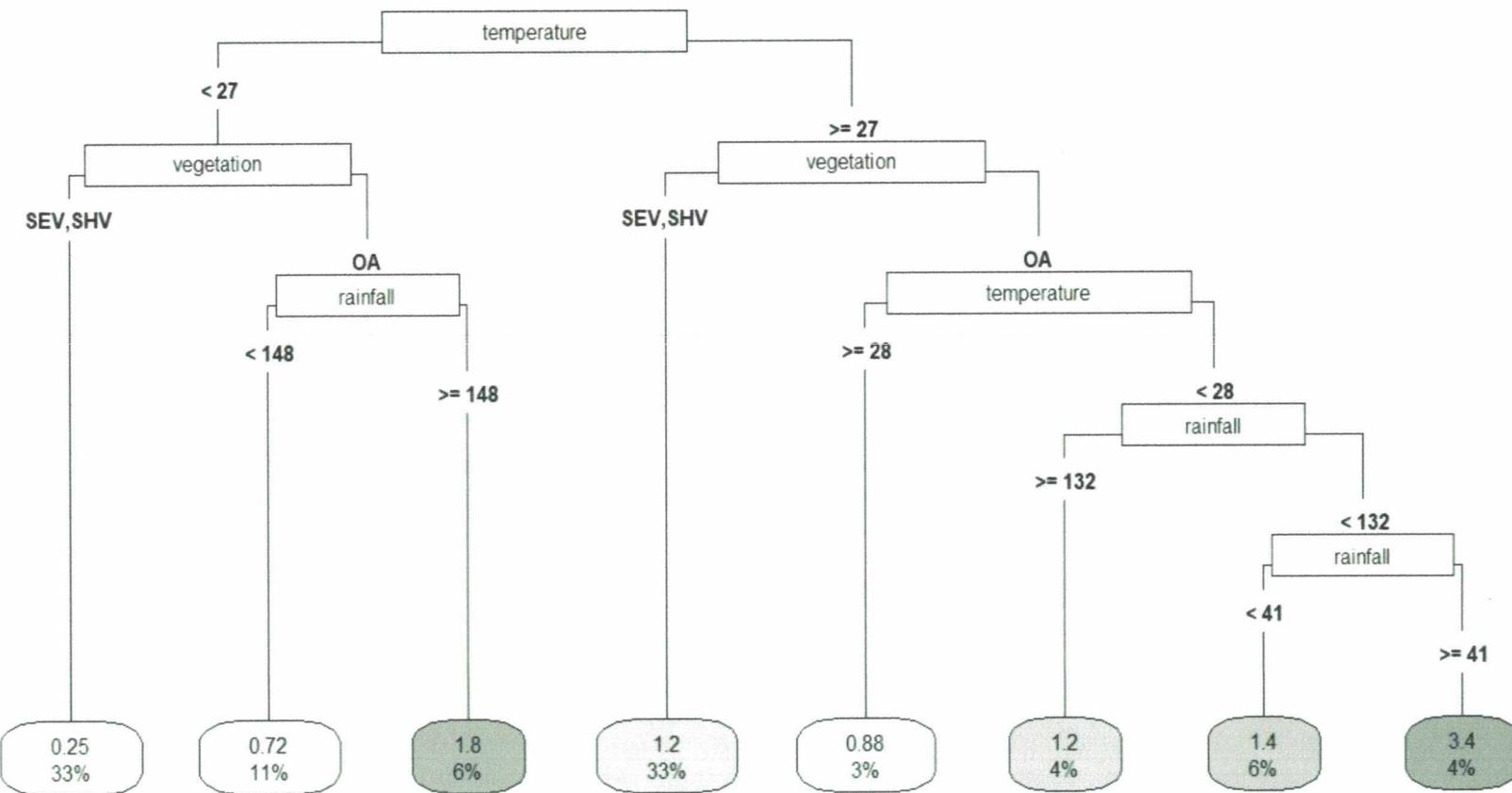
Figure 2 - Temperature (°C), rainfall (mm³) and nests founded by cavity-nesting bees and wasps in open areas, shrubby and secondary vegetation in restinga domain over two year.

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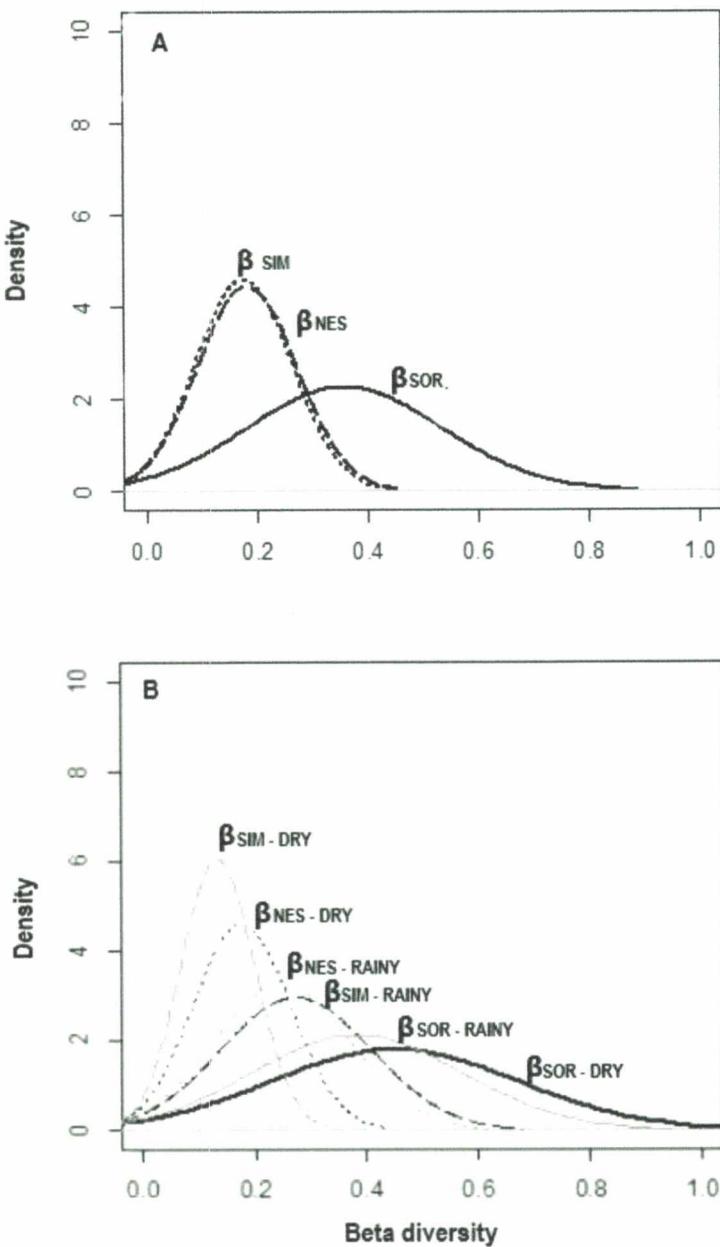
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Figure 1 - Regression tree showing the influence of climatic conditions and vegetation types (OA- open area; SHV – shrubby; SEV - secondary vegetation) on nesting by cavity-nesting bees and wasps in restinga domain over two year.



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Figure 2 - Beta diversity components for cavity-nesting bee and wasp community. (A) Among the vegetation types of restinga domain (open area, shrubby and secondary vegetation); (B) During dryer and rainy season. β_{SIM} (Simpson dissimilarity, i.e. species turnover), β_{NES} (nestedness) and β_{SOR} (Sørensen dissimilarity, i.e. total beta diversity).

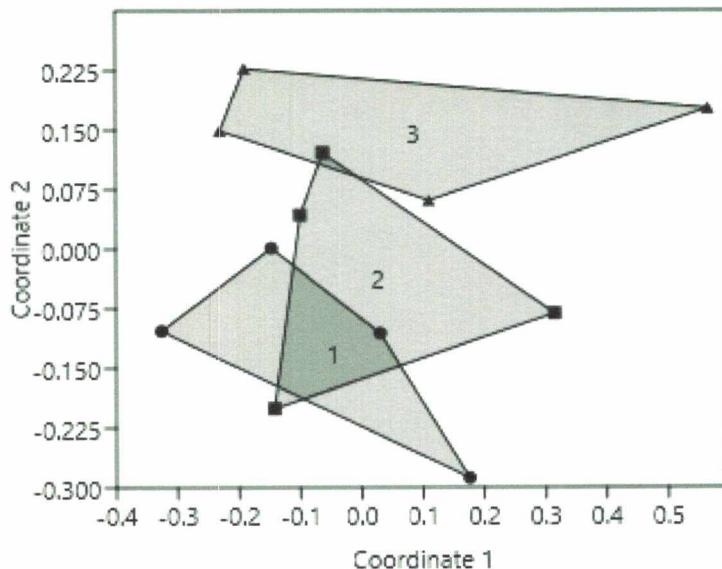


Figure 3- Non-metric scaling analysis of similarity (NMDS) ordination for species abundance of cavity-nesting bees and wasps in open areas (1), shrubby (2) and secondary vegetation (3) in restinga domain. NMDS were calculated based on Bray-Curtis similarity and Bonferroni correction at 5% level.

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CAPÍTULO 3

**Rainfall, temperature and vegetation type influence nesting by
the oil-collecting bee *Centris (Hemisiella) tarsata* in Brazilian restinga**

Manuscrito aceito pelo periódico Apidologie

24 **Rainfall, temperature and vegetation type influence nesting by**
25 **the oil-collecting bee *Centris (Hemisiella) tarsata* in Brazilian restinga**

26

27 **Nesting by *Centris tarsata* in restinga**

28 **Abstract**

29 Solitary bees are the main pollinators of native plant species and crops, therefore
30 understanding how they respond to environment is essential to maintain ecosystem
31 function and services. Here, we assessed how climatic conditions and vegetation
32 type influence *Centris tarsata* nesting in Brazilian restinga. Evaluations were
33 conducted using trap-nests placed in open areas, shrubby and secondary vegetations,
34 and at understory and canopy in open areas and secondary vegetations. Overall,
35 nesting was higher during the dry season, mostly in open areas at 1.5m height.
36 However, in shrubby and secondary vegetations nesting seems to be favored by a
37 specific rainfall and temperature range. Furthermore, the amount of brood cells, total
38 number of adults and nest length were higher in open areas compared to shrubby and
39 secondary vegetations. Therefore, our results show that rainfall, temperature and
40 vegetation structure are determining factors in *C. tarsata* nesting in the Brazilian
41 restinga.

42 **Keywords:** Trap-nests / cavity-nesting bees / pollinators / climate conditions /
43 **Malpighiaceae**

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48 **1 – INTRODUCTION**

49

50 Mutualistic interactions between plants and pollinators play a fundamental
51 role in the maintenance of ecosystem functions and services. Plants provide food
52 resources such as pollen and nectar to pollinators, which in turn help fruit and seed
53 formation through pollination services (Giannini et al. 2015). Bees are the most
54 important and effective pollinators. Furthermore, they are ecologically and
55 functionally diverse, forage on a wide variety of floral forms and nest on a myriad of
56 substrates (Klein et al. 2007). The number of bee species worldwide could be as high
57 as 30.000 species (Michener 2007), from which 85% have a solitary habit (Silva et
58 al. 2001; Buschini & Wolff 2006). Solitary bee species have specific habitat
59 requirements such as pre-existing cavities for nesting as well as flower resources
60 (pollen, nectar and oil), and thus, they heavily rely on vegetation structure (Ebeling
61 et al. 2012; Flores et al. 2018). Additionally, the reproductive success of solitary
62 bees can be indirectly affected by climatic conditions owed to their influence on host
63 plant community richness, composition, physiognomy and bee nesting biology.
64 Therefore, habitat simplification has adverse effects on bee diversity, which makes
65 preservation efforts pivotal for the maintenance of Aculeata populations (Morato &
66 Campos 2000, Morato & Martins 2006; Faria & Gonçalves 2013; Flores et al. 2018).

67 Several studies evaluating solitary bee communities in areas with well-
68 defined dry and rainy seasons revealed that *Megachile* sp. (Megachilidae), *Eufriesea*
69 *nordestina* Moure, *Euglossa cordata* Linnaeus, *Tetrapedia diversipes* Klug,
70 *Xylocopa frontalis* Olivier, and species belonging to the genus *Centris* such as
71 *Centris tarsata* Smith, *C. analis* Fabricius and *C. vittata* Lepeletier (Apidae) are the
72 most common species (Aguiar & Martins 2002; Aguiar et al 2005, Mesquita &

73 Augusto 2011, Flores et al 2018). The genus *Centris* encloses medium to large sized
74 solitary bee species, it is predominantly occidental, and distributed from tropical
75 areas of Argentina and Bolivia to the United States (Silveira et al. 2002; Michener
76 2007). These species need a rich flora to obtain food resources such as nectar, pollen
77 and floral oils for both larvae and adults. *Centris (Hemisiella) tarsata* (Smith), 1874
78 is widely distributed in Brazil and it has been recorded in different biomes and
79 climatic conditions countrywide (Silva et al. 2001; Aguiar & Martins 2002; Aguiar
80 & Garófalo 2004; Buschini & Wolff 2006, Mendes & Rêgo 2007, Mesquita &
81 Augusto 2011, Flores et al. 2018).

82 Although some studies have investigated the vertical nesting of bees in
83 tropical forests (Morato 2001; Mendes & Rêgo 2007; Stangler et al. 2015), they were
84 carried out in the Amazon forest and in semi-deciduous vegetation in Brazil, and in
85 forest remnants in Costa Rica. Information on bee nesting preference for understory
86 or canopy is scarce, and to our knowledge no study has addressed the vertical nesting
87 behavior of solitary bees in restinga. This tropical and subtropical ecosystem is
88 typical on coastal regions, and it is composed by herbaceous plants, shrubs and
89 arboreous strata on sandy plains under marine influence (Souza et al. 2008; Serra et
90 al. 2016). Accordingly, we addressed the following questions: (1) is there an
91 association among climatic conditions, type of vegetation and *C. tarsata* nesting
92 behavior? (2) Do nest architecture (brood cell number built, nest length, cavity
93 width, thickness of entrance and partition walls, and volume of cell) biological
94 characteristics and parasitism vary among vegetation types? (3) Is there preference
95 for *C. tarsata* nesting at different vegetation strata?

96

97

98 **2 – MATERIAL AND METHODS**

99

100 **2.1 – Study area**

101

102 The study was carried out in a protected area (763.37 ha) belonging to
103 Empresa Brasileira de Pesquisa Agropecuária (Embrapa) in the municipality of
104 Itaporanga D'Ajuda ($11^{\circ} 06'16.19''S$; $37^{\circ} 11'05.89''W$), in the northeastern Brazilian
105 State of Sergipe. According to Köppen classification, predominant climate in the
106 region is tropical with marked dry (September to February) and rainy (March to
107 August) seasons, with average annual rainfall above 1250 mm. The biome is Atlantic
108 rainforest composed by a matrix of mangrove, restinga and apicum (sandy open
109 areas influenced by the tidal regime). The study was conducted within the restinga
110 domain (Nogueira Junior et al. 2013). Rainfall and temperature were respectively
111 obtained from a meteorological station located in the area and from the Instituto
112 Nacional de Meteorologia (INMET), on a per-month basis.

113

114 **2.2 - Sampling**

115

116 Trap-nests consisted of 15 pieces of hollow bamboo canes (20.0cm long)
117 with one node in the middle, forming two internodes (9.0 to 11.0cm deep). Bamboo
118 canes had internal diameters ranging from 0.3 to 2.0cm evenly distributed (5 hollow
119 bamboo canes of 0.3 to 0.8cm, 5 of 0.9 to 1.4cm and 5 of 1.5 to 2.0cm of diameter)
120 and they were wrapped together with a 22.0cm wide plastic sheet and wire, totaling
121 thirty cavities. Trap-nests were set up in three vegetation types in restinga, viz. (1)
122 open areas - dominance of grasses as *Paspalum maritimum* Trind., weeds, cactus
123 (*Melocactus zehntneri*, Britton & Rose) and some sparsely distributed trees, direct

124 sunlight; (2) shrubby vegetations - dominance of shrubs and few tree species as
125 *Anacardium occidentale* L., *Byrsonima crassifolia* L. Kunth, *Myrciaria tenella* D.C.
126 Berg, some grasses, partial shading with high luminosity; and (3) secondary
127 vegetations – dominance of medium to large tree species as *A. occidentale*,
128 *Hancornia speciosa* Gomes, *Tapira guianensis* Aublet and *Inga* sp., dense
129 vegetation, soil covered by a thick litter layer, shaded with low luminosity. Four
130 replicates for each vegetation type were chosen, totaling twelve sampling areas.
131 Study sites were usually clustered in groups of three vegetation types to avoid spatial
132 autocorrelation (Teodoro et al. 2011). At each study site, four trap-nests were hung
133 by a wire and attached to a tree branch at 1.5m above the ground and distanced 1.0m
134 from each other. Entomological glue was monthly applied to the wire to avoid
135 crawling arthropods. Trap-nests were monthly inspected throughout two years (from
136 July 2016 to June 2018) and nested bamboo canes were removed and replaced with
137 empty ones with similar internal diameter.

138 Nested bamboo canes were individualized in a plastic cage and kept indoors
139 at $27^{\circ}\text{C} \pm 2^{\circ}\text{C}$ until the emergence of *C. tarsata* adults and their cleptoparasites, also
140 known as cuckoo bees or cuckoo wasps that loot brood cells storage (Ebeling et al.
141 2012) and parasitic flies. Emerging adults were killed, sexed and the morphometry
142 was performed. Nest architecture features such as the count of brood cells and
143 measures of total nest length (from the first to last cell built), nest width, and
144 thickness of the brood cell partition and entrance wall were also evaluated with a
145 caliper ruler. Brood cell volume was calculated by the formula: $\pi.r^2.h$, where r is the
146 width and h the length of the cell. Wings of three adults of each sex were measured
147 with a scale-magnifier model MG13100, totaling 12 adults (females and males) from
148 each vegetation type, except for secondary vegetation where in one of the sampling

149 areas, no adults emerged. Identification of nest building material and counting of
150 dead immatures (no emerging individuals of *C. tarsata*, cleptoparasites or offspring
151 who died in early stages) were performed three months after the nests were taken to
152 the laboratory.

153

154 2.3. - Nesting stratification

155

156 An additional nesting stratification preference study was performed by lining
157 up two trap-nests tied together with a wire and attached to a rope at 1.5, 3.5 and 5.5m
158 height in open areas and secondary vegetations, with four replicates. The rope was
159 attached to a pulley at the tip of a 7.0m high bamboo pole. Entomological glue was
160 applied to bamboo pole base to prevent crawling arthropods. Evaluations were
161 monthly conducted for one year, beginning in September 2016, and the nested
162 bamboo canes were taken to the laboratory and replaced by empty ones with similar
163 internal diameters. Nested cavities were kept indoors as explained above.

164

165 2.4 - Statistical analyses

166

167 Data were nested and analyzed using generalized linear mixed-effects
168 models (GLMEs) with Poisson (discrete variables) or Gaussian (continuous
169 variables) error distributions (Crawley 2007). The models were subjected to analysis
170 of variance (ANOVA), and *a posteriori contrasts* were performed to assess
171 differences among treatments (Crawley 2007). Spearman correlations were
172 performed between the number of nests built and temperature, rainfall and tree
173 abundance using ISwR package. Pearson's chi square test (χ^2) was performed to

174 compare *C. tarsata* sex ratios among vegetation types. All analyses were performed
175 with R software version 3.3.1 (R Development Core Team 2016) and subjected to
176 residual analyses in order to assess the suitability of the models and error
177 distributions (Crawley 2007).

178 A classification and regression tree analysis was further performed for a
179 better understanding of how climatic conditions influenced *C. tarsata* nest building
180 on different vegetation types. Evaluation period (month), temperature (°C), rainfall
181 (mm) and vegetation types were considered as explanatory variables, while the
182 number of nests built by *C. tarsata* was treated as response variable. The root node
183 represents total sampling, which further divides into two or more sets called decision
184 nodes and into sub-nodes and/or terminal nodes. Differences among sub-nodes and
185 parent node were accessed through the Chi-Squared Automatic Interact Detection
186 (CHAID) method (Kass 1980). The analysis was performed in XLSTAT 2014.5.3
187 and adjusted by Bonferroni correction, at 5% significance level.

188

189 **3 - RESULTS**

190

191 3.1 – *Centris tarsata* nesting according to vegetation type

192

193 A total of 162 nests of *C. tarsata* were collected, the majority in open areas
194 (52.47%), 37.03% in shrubby and only 10.50% in secondary vegetations. Altogether,
195 585 adults of *C. tarsata* emerged from the nests, most of them from open areas
196 (59.6%), followed by 32.14% from shrubby vegetations and 8.20% from secondary
197 vegetations. The cuckoo bee *Mesocheira bicolor* Fabricius (54.05%) was the main *C.*
198 *tarsata* natural enemy emerging from nests, followed by an Anthidiini species

199 (37.84%) and *Coelioxys* sp. (1.8%) (Megachilidae). Other natural enemies emerging
200 from nests were the dipterans *Anthrax* sp1 (4.5%) and *Anthrax* sp2 (1.8%)
201 (Bombyliidae).

202 *Centris tarsata* nesting period lasted from October 2016 to February 2017,
203 peaking in November and December during the dry season. A new nesting cycle was
204 recorded from September 2017 until January 2018, peaking in November and in
205 January (Fig. 1). Nesting was negatively correlated with rainfall ($\rho = -0.581$, P
206 <0.01) and abundance of tree species ($\rho = -0.755$, $P <0.01$), but not with
207 temperature ($\rho = 0.2047$, $P>0.05$).

208 Classification and regression tree analysis yielded five levels of depth as
209 shown in Figure 2. In higher temperatures, the mean number of nests was 0.81 in
210 36% of samples ($P<0.001$) contrasting with a lower number (0.28) in 64% of
211 samples in lower temperatures ($P<0.05$). In shrubby and secondary vegetations,
212 nesting mostly occurred when temperatures were high and associated to a low
213 rainfall ($P<0.001$). Also, there was no nesting in 16% of samples in shrubby and
214 secondary vegetations with increasing rainfall ($P<0.001$). In open areas, the main
215 explanatory factor was the sampling period, being the mean number of nests (1.25)
216 higher in December in 1.3% of samples ($P<0.001$) followed by September and
217 October. Moreover, the mean number of nests was higher (4.0) in 1.3% of samples
218 ($P<0.001$) when the temperature increased in June, July, August and November. In
219 those months, there was a decrease in cumulative rainfall in relation to the previous
220 period (Fig. 1).

221

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225 3.2 Nest architecture and bionomical characteristics

226

227 Nests were constructed with a mixture of sand and vegetal oil which covered
228 the outer wall of the entrance, hardening after dry. Brood cells were arranged
229 according to cavity width, linear in smaller sized cavities, perpendicular in medium
230 sized and in some cases, overlapping each other in larger cavities. The inner brood
231 cell was smooth and hard in appearance, covered by a cellophane-like film. Brood
232 cells contained a dark yellow to brown mixture of pollen and nectar as food supply
233 for larvae.

234 The number of brood cells built per nest differed among vegetation types:
235 open areas had the highest numbers, followed by shrubby and secondary vegetations.
236 There was no difference in nest width, wall thickness, and brood cell volume among
237 vegetation types. The partition wall thickness of the brood cells was higher in
238 secondary vegetations compared to open areas and shrubby vegetations (Table I).

239 After field collection, emergence of adults from nests ranged from 2.0 to 49.0
240 days for males and from 4.33 to 40.6 days for females. The total number of emerged
241 adults and males of *C. tarsata* was influenced by vegetation types ($\chi^2 = 4.84$, d.f.=1,
242 $P<0.001$; $\chi^2 = 7.75$, d.f. = 1, $P<0.01$, respectively), being highest in shrubby
243 vegetations, followed by open areas and secondary vegetations. The mean number of
244 females, cleptoparasites and dead immatures were also influenced by vegetation
245 types ($\chi^2 = 15.67$ d.f. = 1, $P<0.001$; $\chi^2 = 6.62$, d.f. = 1, $P<0.01$; $\chi^2 = 31.67$, d.f. = 1,
246 $P<0.01$, respectively), with higher values for open areas, followed by shrubby and
247 secondary vegetations. There was no difference in the size of male or female wings
248 ($\chi^2 = 0.48$, d.f. = 2, $P>0.05$; $\chi^2 = 0.08$, d.f. = 2, $P>0.05$, respectively) (Table I) among

249 vegetation types. Sex ratio was significantly male biased in secondary vegetations
250 ($\text{sr} = 1:0.37$, $\chi^2 = 13.13$, d.f. = 5, $P < 0.05$), but not in open areas ($\text{sr} = 1:0.35$, $\chi^2 =$
251 19.33, d.f. = 19, $P > 0.05$) and shrubby vegetations ($\text{sr} = 1:0.27$, $\chi^2 = 17.42$, d.f. = 13,
252 $P > 0.05$).

253 3.3 – Nesting stratification

254

255 A total of 31 nests founded by *C. tarsata* were collected, roughly half of
256 which (14 nests) in open areas and closer to the soil (1.5m height), followed by 9 and
257 3 nests at 3.5m and 5.5m height, respectively. We observed considerable fewer nests
258 in secondary vegetations (16.13%; 1 nest at 1.5m, and 2 nests at 3.5m and 5.5m
259 height). The cuckoo bee *M. bicolor* was the only cleptoparasite that emerged from
260 the nests and it was found only in open areas.

261 The highest number of brood cells and dead immatures ($P < 0.001$) was
262 recorded at 1.5m height and the lowest at 5.5m in open areas. The total of individuals
263 was lower in nests closer to the canopy (at 5.5m height) when compared to those
264 nearer to the soil (1.5m height). Neither vegetation nor trap-nest height influenced
265 the total number of males and females (Table II).

266

267 **4 - DISCUSSION**

268

269 Rainfall was the main climate condition explaining *C. tarsata* nesting
270 patterns over the 2-year sampling period, and it was negatively correlated with
271 nesting. In contrast, temperature had no influence on nesting, possibly owing to its
272 low variation throughout the year, which contrasts with well-defined rainy and

273 drought periods in the study region. However, when temperature and rainfall were
274 analyzed in relation to the type of vegetation and throughout the sampling period,
275 new results emerged. The classification and regression tree shows two distinct
276 groups of *C. tarsata* (Fig. 2): one influenced directly by rainfall and temperature and
277 another influenced by the sampling period. Nesting in shrubby and secondary
278 vegetations was favored at low rainfall and high temperature, increasing rainfall
279 ceases nesting, possibly because high humidity is unfavorable for nesting. In open
280 areas, *C. tarsata* nesting was also generally concentrated during the dry season, in
281 months of low rainfall and higher temperatures. Either drought or rainfall periods
282 beyond normal may disturb floral resources availability, and consequently bee
283 frequency. The frequent annual population fluctuations appear to be a regular feature
284 for bee species that nest in pre-existing cavities (Frankie et al. 1998).

285 This nesting pattern in the dry season is consistent with results from other
286 Brazilian vegetations such as eucalyptus, riparian and mesophytic forests in
287 Maranhão State, semideciduous seasonal forest and open savanna in Paraíba State,
288 and swamps and pastures in the southern region (Aguiar & Martins 2002; Buschini
289 & Wolff 2006; Mendes & Rêgo 2007). However, Aguiar and Garófalo (2004)
290 observed more *C. tarsata* nests during the rainy season in semi-deciduous and
291 semiarid caatinga vegetations. Sunny habitats can provide more resources for the
292 construction and provision of nests, such as floral resources, when compared to
293 shaded habitats, as observed in open areas. Concerning floral resources, we
294 frequently observed *C. tarsata* foraging on *Byrsonima crassifolia* (Malpighiaceae)
295 mainly in shrubby vegetations, which is in line with other studies that recorded bees
296 foraging for oil and pollen on *Byrsonima* spp. (Mendes & Rêgo 2007; Mello et al.
297 2013).

298 Our results show that *C. tarsata*, in restinga vegetation, act as a bivoltine

299 species with two generations per year, each generation lasting approximately two

300 months. Nests were found for only five months per year, in both years. Similar

301 results were found by Buschini & Wolff (2006) for *C. tarsata* and by Martins et al

302 (2014) for *C. flavifrons*. Aguiar & Garófalo (2004) studying *C. tarsata* nesting

303 behavior in caatinga and semi-deciduous vegetation observed some nests diapausing

304 after being established in mid rainy season (caatinga vegetation) or in mid dry season

305 (semi-deciduous vegetation), both stressful periods. These authors suggested that *C.*

306 *tarsata* possibly spends stressful periods either as adults or as prepupae in diapause

307 to withstand harsh environmental conditions. In fact, evidence indicates that

308 diapause is the primary mechanism through which the annual rhythm of insect life-

309 history phases are appropriately synchronized to seasonal conditions, and it is

310 subjected to both genetic and environmental influence (Tauber & Tauber, 1981;

311 Faria & Gonçalves, 2013). Another strategy may be taking shelter in unknown

312 places, as suggested by Pereira et al (1999) and Ramos et al (2010). Adults of *C.*

313 *tarsata* might diapause in restinga vegetation during stressful periods, but this

314 remains yet to be investigated.

315 Regarding nest architecture, *C. tarsata* females built thicker brood cells

316 partition wall in secondary vegetations compared to open areas and shrubby

317 vegetations. This could be related to the smaller amount of resources for nest

318 building (sand) in secondary vegetations, since its ground is covered by a large

319 amount of litter. However, vegetation type did not influence the thickness of

320 entrance plug wall and we observed a layer of oil outside of all nests entrance plug,

321 which may be a characteristic from subgenus *Hemisiella* to protect the nest (Pereira

322 et al. 1999; Buschini & Wolff, 2006).

323 Sex ratio was male biased in secondary vegetation, but not in open areas and

324 shrubby vegetations. Aguiar & Martins (2002) and Aguiar & Garófalo (2004) also

325 reported a male-biased sex ratio for *C. tarsata*. However, Silva et al. (2001),

326 Buschini & Wolff (2006) and Mendes & Rêgo (2007) found a female-biased sex

327 ratio. According to Silva et al. (2001), sex ratio is often variable, and it is associated

328 to the abundance of resources available in the environment for females. Also, a

329 lower number of emerging females might be caused by parasitism, as parasites and

330 cleptoparasites prefer female brood cells, possibly due to higher nutritional quality of

331 food stored (Aguiar & Martins 2002; Buschini & Wolff 2006).

332 Cleptoparasites observed in *C. tarsata* nests were the cuckoo bees *M. bicolor*,

333 *Coelioxys* sp. and Anthidiini, and the dipterans *Anthrax* sp1, *Anthrax* sp2, which

334 were also recorded in other Brazilian biomes (Aguiar & Martins 2002; Aguiar &

335 Garófalo 2004; Buschini & Wolff 2006). According to a study carried out in small

336 secondary forest remnants in Costa Rica, parasitism rates tend to be high in the

337 understory due to the dominance of *Centris*, which was observed nesting at 2.0m

338 height (Stangler et al. 2015). Here, the low parasitism rate recorded in the understory

339 may possibly be due to the low number of nests founded. Therefore, further studies

340 are needed to elucidate the parasitism of *C. tarsata* in restinga, mainly in the

341 understory and canopy.

342 The number of dead immatures was higher in open areas, possibly due to a

343 related high number of nests and brood cells built. Moreover, climate conditions,

344 especially temperature, can reach extreme values in open areas, potentially

345 increasing larvae and juvenile bee mortalities (Jesus & Garófalo 2000; Aguiar &

346 Garófalo 2004; Buschini & Wolff 2006).

347 In open areas, *C. tarsata* preferred to nest in trap-nests near to the ground
348 unlike in secondary vegetations, where only one nest was built at this height. This
349 preference for lower heights in restinga may be favored by proximity from nest
350 building resources such as sand, and floral source such as oil, nectar and pollen. On
351 the other hand, for denser habitats such as secondary forests, microclimatic
352 conditions may be more favorable for nesting at higher strata, which may explain our
353 results for secondary vegetation. Also, Mendes & Rêgo (2007) observed that *C.*
354 *tarsata* built more nests in the canopy of eucalyptus (5 to 12m height) and
355 mesophytic forest (12.0m height) compared to nests at 1.5m heigh.

356 In restinga, rainfall is a determining factor in *C. tarsata* nesting which was
357 concentrated in the dry season, mainly in open areas, where we also recorded the
358 greatest amount of brood cells and highest nest length. In shrubby and secondary
359 vegetations, nesting seems to be favored by specific climatic conditions present in
360 the understory and canopies especially in the harsh dry season. *Centris tarsata*
361 preferred to nest in the understory in secondary vegetations and open areas, however,
362 further studies on nesting behavior and exploitation of floral resources in restinga are
363 needed in order to improve the management and maintenance of related ecosystem
364 functions and services.

365

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373 **AUTHORS CONTRIBUTIONS** DMC and AVT conceived this research and
374 designed experiments; MCB participated in the design, analysis and interpretation of
375 the data; ASB participated in the design; IB performed analysis and interpretation of
376 the data. All authors read and approved the final manuscript.

377 **CONFLICT OF INTEREST**

378 “The authors declare that they have no potential conflict of interest in relation to the
379 study in this paper”.

380 **5 – REFERENCES**

381

382 Aguiar, A. J. C.; Martins, C. F. (2002). Abelhas e vespas solitárias em ninhos-
383 armadilha na Reserva Biológica Guaribas (Mamanguape, Paraíba, Brasil). Rev. Bras.
384 Zool. 19 (1), 101-116. <https://doi.org/10.1590/s0101-81752002000500005>

385 Aguiar, C. M. L.; Garófalo, C. A. (2004). Nesting biology of *Centris (Hemisiella)*
386 *tarsata* Smith (Hymenoptera, Apidae, Centridini). Rev. Bras. Zool. 21 (3), 477-486.
387 <https://doi.org/10.1590/s0101-81752004000300009>

388 Buschini, M. L. T.; Wolff, L. L. (2006). Nesting biology of *Centris (Hemisiella)*
389 *tarsata* Smith in southern Brazil (Hymenoptera, Apidae, Centridini). Braz. J. Biol.
390 66 (4), 1091-1101. <https://doi.org/10.1590/s1519-69842006000600016>

391 Crawley, M.J. (2007). The R book. John Wiley & Sons Ltd, Wests Sussex.

392 Ebeling, A.; Klein, A.M.; Weisser, W.W.; Tscharntke, T. (2012). Multitrophic
393 effects of experimental changes in plant diversity on cavity-nesting bees, wasps, and

- 394 their parasitoids. *Oecologia*. **169** (2), 453-465. <https://doi.org/10.1007/s00442-011-2205-8>
- 396 Faria, L. R. R.; Gonçalves, R. B. (2013) Abiotic correlates of the bee diversity and
397 composition along eastern Neotropics. *Apidologie*. **44** (5), 547-562.
398 <https://doi.org/10.1007/s13592-013-0205-x>
- 399 Flores, L. M. A.; Zanette, L. R. S.; Araújo, F. S. (2018). Effects of habitat
400 simplification on assemblages of cavity nesting bees and wasps in a semiarid
401 neotropical conservation area. *Biod. Conserv.* **27** (2), 311-328.
402 <https://doi.org/10.1007/s10531-017-1436-3>
- 403 Frankie, G. W., Vinson, S. B., Rizzardi, M. A., Griswold, T. L., O'Keefe, S.,
404 Snelling, R.R. (1998). Diversity and abundance of bees visiting a mass flowering
405 tree species in disturbed seasonal dry forest, Costa Rica. *J. Kansas Entomol. Soc.* **70**
406 (4), 281-296.
- 407 Giannini, T. C.; Boff, S.; Cordeiro, G.D.; Cartolano Jr., E. A.; Veiga, A. K.;
408 Imperatriz-Fonseca, V. L.; Saraiva, A. M. (2015). Crop pollinators in Brazil: a
409 review of reported interactions. *Apidologie*. **46** (2), 209 – 223.
410 <https://doi.org/10.1007/s13592-014-0316-z>
- 411 Jesus, B. M. V., Garófalo, C. A. (2000). Nesting behaviour of *Centris*
412 (*Heterocentris*) *analis* (Fabricius) in southeastern Brazil (Hymenoptera, Apidae,
413 Centridini). *Apidologie*, **31** (4), 503-515. <https://doi.org/10.1051/apido:2000142>
- 414 Kass, G. V. (1980). An exploratory technique for investigating large quantities of
415 categorical data. *Applied Statistics*, **29** (2), 119-127
- 416 Klein, A. M., Vaissière, B. E., Cane, J. H., Steffan-Dewenter, I., Cunningham, S. A.,

- 417 Kremen, C., Tscharntke, T. (2007). Importance of pollinators in changing landscapes
418 for world crops. Proc. R. Soc. B. **274** (1608), 303 – 313.
419 <https://doi.org/10.1098/rspb.2006.3721>
- 420 Martins, C. F.; Peixoto, M. P; Aguiar, C. M. L. (2014) Plastic nesting behaviour of
421 *Centris (Centris) flavifrons* (Hymenoptera: Apidae: Centridini) in an urban area.
422 Apidologie. **45** (2), 156-171. <https://doi.org/10.1007/s13592-013-0235-4>
- 423 Mello, M. A. R., Bezerra, E. L. S., Machado, I. C. (2013). Functional roles of
424 Centridini oil bees and Malpighiaceae oil flowers in biome-wide pollination
425 networks. Biotropica. **45** (1), 45-53. <https://doi.org/10.1111/j.1744-7429.2012.00899.x>
- 427 Mendes, F. N.; Rêgo, M. M. C. (2007). Nidificação de *Centris (Hemisiella) tarsata*
428 Smith (Hymenoptera, Apidae, Centridini) em ninhos-armadilha no Nordeste do
429 Maranhão. Brasil. Rev. Bras. Entomol. **51** (3), 382-388.
430 <https://doi.org/10.1590/s0085-56262007000300017>
- 431 Mesquita, T. M. S., Augusto, S. C. (2011). Diversity of trap-nest bees and their
432 natural enemies in the Brazilian savanna. Tropical Zool. **24** (2), 127-144
- 433 Michener, C. D. (2007). The bees of the world. The Johns Hopkins University Press,
434 Baltimore.
- 435 Morato, E. F. (2001). Efeitos da fragmentação florestal sobre vespas e abelhas
436 solitárias na Amazônia Central. II. Estratificação vertical. Rev. Bras. Zool. **18** (3),
437 737-747. <https://doi.org/10.1590/s0101-81752001000300010>
- 438 Morato; E.F.; Campos, L.A.O. (2000). Efeitos da fragmentação florestal sobre
439 vespas e abelhas solitárias na Amazônia Central. Rev. Bras. Zool. **17** (2), 429-444.

- 440 <https://doi.org/10.1590/s0101-81752000000200014>
- 441 Morato, E. F.; Martins, R. P. (2006). An overview of proximate factors affecting the
442 nesting behavior of solitary wasps and bees (Hymenoptera: Aculeata) in preexisting
443 cavities in wood. *Neotropical Entomol.* **35** (3), 285–298.
444 <https://doi.org/10.1590/s1519-566x2006000300001>
- 445 Nogueira Junior, L. R.; Dompieri, M. H. G.; Rangel, M. S. A. Rodrigues, R. F. A.;
446 Melo, A. F. R. et al. (2013). Plano de Manejo Reserva Particular do Patrimônio
447 Natural do Caju. Embrapa Tabuleiro Costeiros. Aracaju.
- 448 Pereira, M.; Garófalo, C. A.; Camillo, E.; Serrano, J. C. (1999). Nesting biology of
449 *Centris (Hemisiella) vittata* Lepeletier in southeastern Brazil (Hymenoptera, Apidae,
450 Centridini). *Apidologie*. **30**, 327-338. <https://doi.org/10.1051/apido:19990409>
- 451 R Core Team (2016). R: A language and environment for statistical computing. R
452 Foundation for Statistical Computing, Vienna, Austria. [Https://www.R-project.org/](https://www.R-project.org/).
- 453 Serra, F. C. V.; Lima, P. B.; Almeida Jr., E. B. De. (2016). Species richness in
454 restinga vegetation on the eastern Maranhão State, Northeastern Brazil. *Acta Amaz.*
455 **46** (3), 271–280. <https://doi.org/10.1590/1809-4392201504704>
- 456 Ramos, M.; Albuquerque, P.; Rêgo, M. (2010). Nesting behaviour of *Centris*
457 (*Hemisiella*) *vittata* Lepeletier (Hymenoptera: Apidae) in an area of the cerrado in
458 the northeast of the State of Maranhão, Brazil. *Neotropical Entomol.* **39** (3), 379-
459 383. <https://doi.org/10.1590/s1519-566x2010000300011>
- 460 Silva, F. O.; Viana, B. F.; Neves, E. L. (2001). Biologia e arquitetura de ninhos de
461 *Centris (Hemisiella) tarsata* Smith (Hymenoptera: Apidae: Centridini). *Neotropical*
462 *Entomol.* **30** (4), 541-545. <https://doi.org/10.1590/s1519-566x2001000400005>

- 463 Silveira, F. A.; Melo, G. A. R.; Almeida, E. A. B. (2002). Abelhas brasileiras:
464 sistemática e identificação. Fund. Araucária, Belo Horizonte.
- 465 Souza, C. R. G., Hiruma, S. T., Sallun, A. E. M., Ribeiro, R. R., Sobrinho, J. M. A.
466 (2008). “Restinga” conceitos e empregos do termo no Brasil e implicações na
467 legislação ambiental. Instituto Geológico, São Paulo
- 468
- 469 Stangler, E. S.; Hanson, P. E.; Steffan-Dewenter, I. (2015). Vertical diversity
470 patterns and biotic interactions of trap-nesting bees along a fragmentation gradient of
471 small secondary rainforest remnants. *Apidologie*. **47** (4), 527-538.
472 <https://doi.org/10.1007/s13592-015-0397-3>
- 473 Tauber, C. A.; Tauber, M. J. (1981). Insect seasonal cycles: genetics and evolution.
474 *Ann. Rev. Ecol. Syst.* **12** (1), 281-308.
475 <https://doi.org/10.1146/annurev.es.12.110181.001433>
- 476 Teodoro, A. V., Muñoz, A., Tscharntke, T., Klein, A. M., Tylianakis, J. (2011).
477 Early succession arthropod community changes on experimental passion fruit plant
478 patches along a land-use gradient in Ecuador. *Agriculture Ecossyst. Environ.* **140**,
479 14-19. <https://doi.org/10.1016/j.agee.2010.11.006>
- 480 XLSTAT (2014). Data analysis and statistical solution for Microsoft Excel.
481 Addinsoft, Paris, France.

482 **Table I** - Nest architecture and bionomical characteristics of *C. tarsata* individuals in relation to restinga vegetation types

	Open areas		Shrubby vegetations		Secondary vegetations		χ^2	d.f.	P
	X ± SD	N	X ± SD	N	X ± SD	N			
Total brood cells	12.87 ± 23.96	515	7.32 ± 12.73	293	2.30 ± 4.45	92	214.6	1	<0.001
Nest lenght (cm)	5.95 ± 1.33	85	5.43 ± 1.54	60	4.48 ± 1.05	17	6.69	1	<0.05
Cavity width (cm)	1.45 ± 1.59	85	1.66 ± 2.30	60	1.07 ± 0.27	17	0.61	2	>0.05
Thickness of partition walls (mm)	0.77 ± 0.004	85	0.78 ± 0.01	60	0.79 ± 0.017	17	5.66	1	<0.05
Thickness of the entrance walls (mm)	0.82 ± 0.063	85	0.79 ± 0.026	60	0.82 ± 0.024	17	0.11	2	>0.05
Volume of cells (cm ³)	2.06 ± 0.56	85	1.89 ± 0.43	60	2.06 ± 0.70	17	0.43	2	>0.05
Total of individuals	13.04 ± 16.46	349	14.35 ± 19.13	188	3.31 ± 5.10	48	4.84	1	<0.001
Total of males	8.27 ± 11.18	224	9.47 ± 11.58	138	2.31 ± 3.56	30	7.75	1	<0.01
Total of females	4.77 ± 6.17	125	4.11 ± 5.36	50	1.5 ± 2.15	18	15.67	1	<0.001
Total of Cleptoparasites	1.32 ± 2.64	52	1.92 ± 3.99	29	0.22 ± 0.61	5	6.62	1	<0.01
Total of dead immatures	2.82 ± 7.0	114	1.9 ± 3.87	76	0.95 ± 2.09	39	31.67	1	<0.001
Male wing lenght (mm)	9.71 ± 0.75	12	9.61 ± 0.72	12	9.83 ± 0.86	9	0.48	2	>0.05
Female wing lenght (mm)	9.94 ± 0.53	12	9.94 ± 0.60	12	10.0 ± 0.68	8	0.08	2	>0.05

483 N = total number, X ± SD = mean ± standard deviation

484

485 **Table II** - Influence of vegetation and height on nesting features by *C. tarsata* in
 486 restinga open areas, shrubby and secondary vegetations.

Response variable	Explanatory variable	χ^2	d.f	P
Brood cells	Vegetation	34.503	1	<0.001
	Height	66.109	2	<0.05
	Vegetation x height	87.237	2	<0.001
Dead immatures	Vegetation	46.145	1	<0.05
	Height	36.669	2	<0.001
	Vegetation x height	82.519	2	<0.05
Total individuals	Vegetation	0.191	1	>0.05
	Height	42.019	2	>0.05
	Vegetation x height	6.224	2	<0.05
Total males	Vegetation	0.411	1	>0.05
	Height	2.598	1	>0.05
	Vegetation x height	0.2507	2	>0.05
Total females	Vegetation	1.901	1	>0.05
	Height	1.626	1	>0.05
	Vegetation x height	2.313	2	>0.05

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ANEXO

ANEXO – A

Types of Papers

Paper (up to 5,000 words): studies that will advance our understanding of evolutionary, ecological, or conservation theory in tropical systems. The data used to test hypotheses can be collected using any rigorous approach (e.g., experimental, observational, genetic, etc.).

Manuscript Format

Submit the entire manuscript, including figures and tables, as a single Microsoft Word document (*.doc or *.docx), or equivalent
for **Do NOT submit papers as pdf files.**

Use 8.5" x 11" page size (letter size) with a 1" margin on all sides. Align left and do not justify the right margin. Number all pages starting with the title page and include continuous line numbers.

Double space throughout the manuscript, including tables, figures and tittle legends, abstract, and literature cited

Use Times New Roman 12-point font throughout except in figures, for which Arial is preferred.

Use the abbreviations provided in Section D (below) throughout the text.

Assemble manuscripts in this order:

1. **Title page**
2. **Abstract (s)**
3. **Data Archiving Policy**
4. **Keywords**
5. **Text**
6. **Tables**
7. **Figure legends**
8. **Figures**
9. **Acknowledgments**
10. **Disclosure Statements**
11. **References**
12. **Supplementary Information (to be supplied as separate files)**

1. TITLE PAGE

Running Heads: The authors' family name should be included as left and right running heads. It is set in small caps. The format is as follows:

LRH and RRH: YAZ and PEIGH

(may not exceed 50 characters, two or more authors use YAZ et al.)

Title: No more than 12 words (usually), flush left, near the middle of the page. Use Bold Type.

Where species names are given in the title, it should be clear to general readers what type(s) of organism(s) are being referred to, either by using Family appellation or common name:

'Invasion of African savanna woodlands by the Jellyfish tree *Medusagyne oppositifolia*', OR 'Invasion of African savanna woodlands by *Medusagyne oppositifolia* (Medusagynaceae)'

Titles that include a **geographic locality** should make sure that this is clear to the general reader:

'Effect of habitat fragmentation on pollination networks on Flores, Indonesia', NOT
'Effect of habitat fragmentation and pollination networks on Flores'.

Authors: Below title, include the author(s) full name(s), affiliation(s), and unabbreviated complete address(es). Use superscript number(s) following author(s) name(s) to indicate current location(s) if different than above. In multi-authored papers, additional footnote superscripts may be used to indicate the corresponding author and e-mail address. Although geographical place names should use the English spelling in the text (e.g., Zurich, Florence, Brazil), authors may use their preferred spelling when listing their affiliation (e.g., Zürich, Firenze, Brasil).

Submission and Acceptance Dates: At the bottom of the title page every article must include: Received: _____; Revised: _____ (optional); Accepted: _____.
(*Biotropica* will fill in the dates.)

2. ABSTRACT PAGE

Abstracts have maximum of 250 words for papers and reviews and 50 words for Insights. There is no abstract for Commentary papers.

The Abstract should include brief statements about the intent or purpose, materials and methods, results, and significance of findings. Abstract can be given as multiple paragraphs (with subheadings such as Aim, Methods, Results, and Conclusion) or as a single paragraph. Do not use abbreviations in the abstract.

Authors are strongly encouraged to provide a second abstract in the language relevant to the country in which the research was conducted. The second abstract will be published in the online versions of the article. This second abstract should follow the first abstract.

3. KEYWORDS

Provide up to eight keywords after the abstract, separated by a comma (,). Keywords should be in English (with the exception of taxonomic information) and listed alphabetically.

Include the location of the study as a key word if it is not already mentioned in the title (see example below). Key words should *not* repeat words used in the title. Avoid words that are too broad or too specific. (e.g., keywords: Melastomataceae, *Miconia argentea*, Panama, seed dispersal, tropical wet forest).

4. TEXT

Headings

Main headings are 1. INTRODUCTION, 2. METHODS, 3. RESULTS, and 4. DISCUSSION in bold, capital letters, numbered, and flush left.

Indent all but the first paragraph of each section.

Leave one blank between main heading and text.

Second level headings should be in Initial caps, bold, numbered, and flush left.
(e.g., 2. Inventory technique.)

First three headings are numbered and fourth and fifth order headings are unnumbered.

Insights submissions do not use any subject headings.

When using previously published data in analyses please cite both the data archive(s) and the original manuscript(s) for which they were collected in the text: "We used previously archived data (Bruna et al., 2011a,b) in our simulations.", where a is the data archive and b is the publication. Be sure both citations are included in the literature cited.

Do not use footnotes in the main text.

Refer to figures as 'Figure 1', and tables as 'Table 1'. Reference to online Supporting Information is referred to as 'Figure S1' or 'Table S1'.**Units, Abbreviations, and style**

Abbreviations: year(s), month(s), week(s), day(s), hr, min, s, km, cm, mm, ha, kg, g, L, g/m²

Units: Use solidus style for simple units (e.g., m/s) and follow negative indices style for compound units (e.g., nmol · hr⁻¹ · mg⁻¹)

Write out other abbreviations the first time they are used in the text and abbreviate thereafter: "El Niño Southern Oscillation (ENSO) . . ."

Numbers: Write out one to 9 unless a measurement or in combination with other numbers: four trees, 6 mm, 35 sites, 7 year, 10 × 5 m, 7 m, ± SE, 5 bees and 12 wasps).

Use a comma as a separator in numbers with four or more digits: 1,000 vs. 10,000

Decimals: 0.13 (leading zero and points, never commas)

Temperature: 21°C (no space after the degree symbol)

Use dashes to indicate a set location of a given size (e.g., 1-ha plot).

Spell out 'percent' when used at the beginning of a sentence and use symbols when used in number combinations (e.g., "there was a 5% increase...", "plants were grown at high light levels (20%)...", 95% CI.)

Statistical abbreviations:

- Use italics for *P*, *N*, *t*, *F*, *R2*, *r*, *G*, *U*, *N*, χ^2 (italics, superscripts non-italics)
- Use italic for: *df*, *SD*, *SE*, *SEM*
- Use roman for CI, two-way ANOVA, ns

Dates: 10 December 1997

Times: 0930 h, 2130 h

Latitude and Longitude: 10°34'21" N, 14°26'12"

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Geographical place names should use the English spelling in the text (Zurich, Florence, Brazil), but authors may use their preferred spelling when listing their affiliation (Zürich, Firenze, Brasil).

Lists: Follow the style... (a)... ; (b)...; and (c)...: “The aims of the study were to: (a) evaluate pollination success in *Medusagyne oppositifolia*; (b) quantify gene flow between populations; and (c) score seed set.”

5. TABLES (Continue page numbering)

While Biotropica does have word limits that differ by manuscript category, there are not have strict limits on the number of tables and/or figures. However, printed manuscripts rarely exceed 32 pages in length, and we encourage authors to submit only necessary tables and figures. Additional information, figures, and tables should appear in the Supporting Information.”

Each table must start on a separate page

Number tables with Arabic numerals followed by a period. Capitalize ‘Table’ (e.g., Table 1, Table 2, etc.).

Indicate footnotes by lowercase superscript letters

Do not use vertical lines in tables.

6. FIGURE LEGENDS (Continue page numbering)

Type figure legends in paragraph form, starting with ‘Figure’ and number.

Do not include symbols (lines, dots, triangles, etc.) in figure legends; either label them in the figure or refer to them by name in the legend

Label multiple plots/images within one figure as a, b, c etc., and please ensure the panels of each plot include these labels and are referred to in the legend (e.g., Figure 1 Fitness of *Medusagyne oppositifolia* as indicated by (a) seed set and (b) seed viability’, making sure to include the labels in the relevant plot.)

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ATBC members can publish graphs and other figures of results in color at no additional charge. Please make sure these figures are accessible by following our **Figures Guidelines**.

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If it is not possible to submit figures embedded within the text file, then submission as *.pdf, *.tif or *.eps files is permissible.

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Line artwork (vector graphics) as *.eps, with a resolution of > 300 dpi at final print size

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Do not use negative exponents in figures, including axis labels.

Each plot/image grouped in a figure or plate requires a label (e.g., a, b). Use lower case letters on grouped figures, and in text references.

8. ACKNOWLEDGEMENTS

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Authors must provide the following disclosures where relevant, in the body of their manuscript (after acknowledgements, before references). If any of the following are not relevant to their research, a null statement is required. If no statement is given, we will publish the article with statements indicating as such.

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Authors must disclose interests that might affect, or appear to affect, their ability to present or review work objectively. These might include relevant financial interests (for example, patent ownership, stock ownership, consultancies, or speaker's fees), or personal, political, or religious interests.

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Bruna, E. M., Izzo, T. J., Inouye, B. D., Uriarte, M., & Vasconcelos, H. L. (2011b). Asymmetric dispersal and colonization success of Amazonian plant-ants queens. PLoS ONE 6, e22937.

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List works by the same author chronologically, beginning with the earliest date of publication.

Insert a period and space after each initial of an author's name; example: Yaz, A. B., & Azy, B. (1980).

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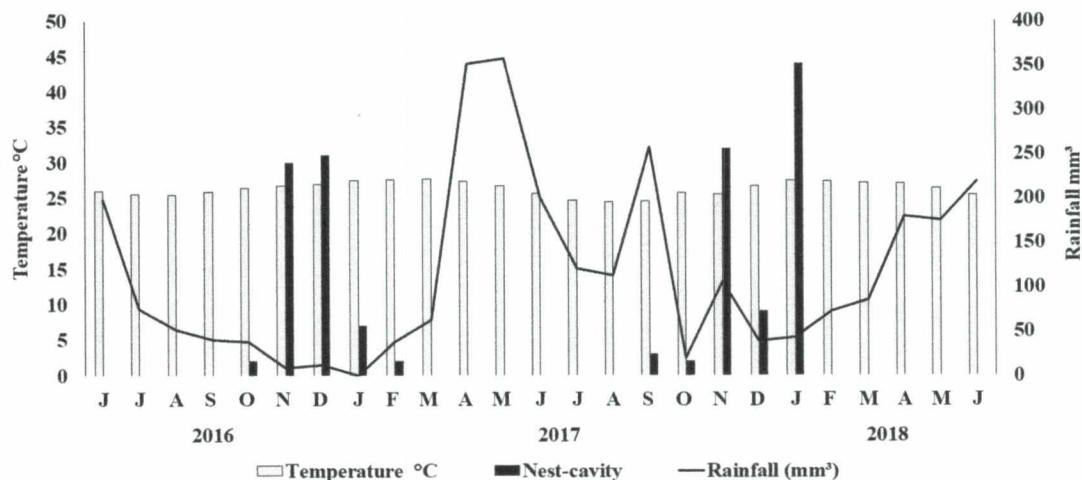
Article in books: Azy, A. B. (1982). Title of book chapter. In: G. Yaz (Ed.). *Book title* (pp. 24–36). Boca Raton, FL: CRC Press.

For theses and dissertations: 'PhD Dissertation' and 'MSc Dissertation'.

Author, A. A. (2003). *Title of doctoral dissertation or master's thesis* (Doctoral dissertation or master's thesis). Retrieved from Name of database. (Accession or Order No.)

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Supporting Information (SI) accompanies the online version of a manuscript and will be fully accessible to everyone with electronic access to *Biotropica*. We urge authors to make use of the SI section to add context or additional information that do not have central relevance to the manuscript. **All such material must be cited in the text of the printed manuscript.**



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491 **Figure 1** - Temperature (°C), rainfall (mm) and number of nests built by *C. tarsata*
492 in restinga open areas, shrubby and secondary vegetations from July 2016 to June
493 2018

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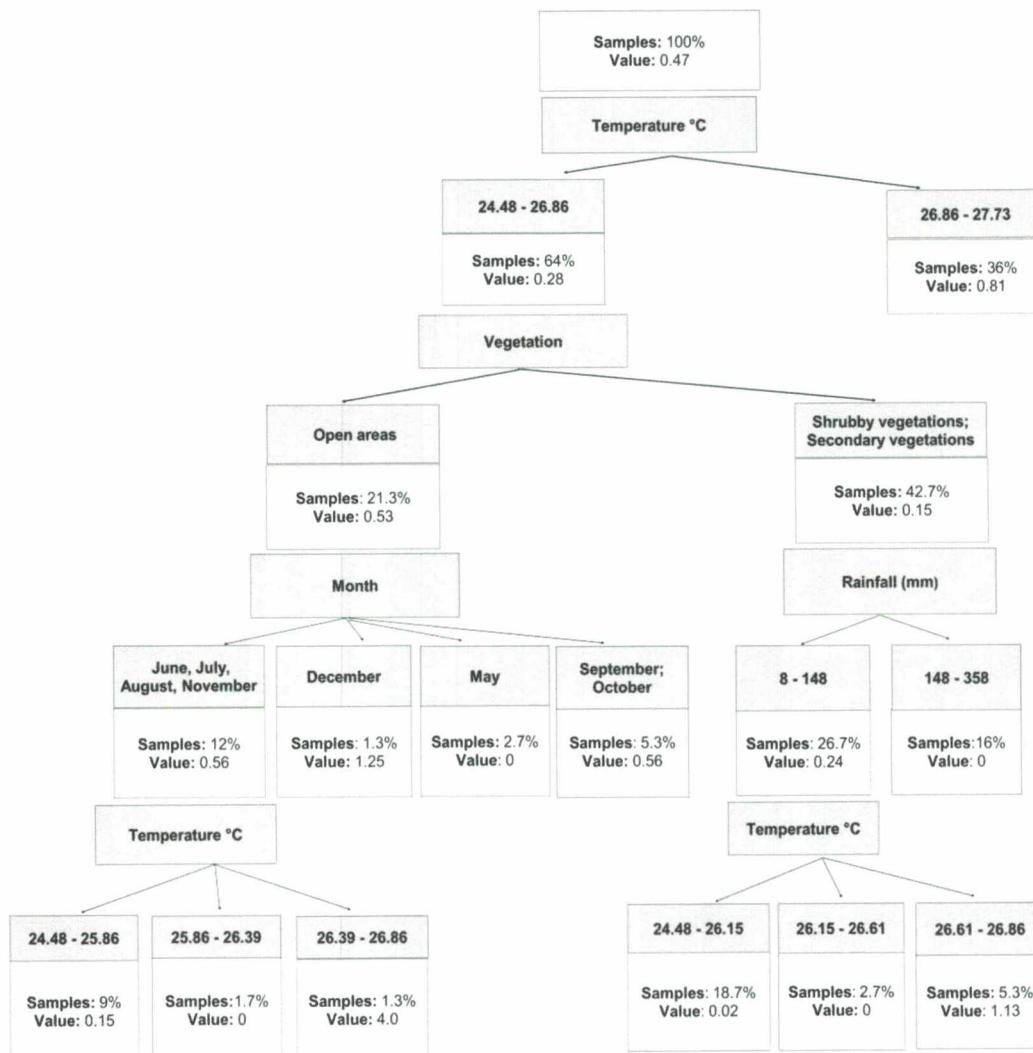
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507 **Figure 2** - Regression tree showing the influence of climatic conditions and
 508 sampling period on nesting by *C. tarsata* in restinga open areas, shrubby and
 509 secondary vegetations

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516 **CONSIDERAÇÕES FINAIS**

517

518 Abelhas e vespas solitárias respondem fortemente às condições climáticas e
519 ao tipo de vegetação em restinga, com preferência por nidificação em áreas abertas e
520 vegetação secundária. A composição de espécies varia ao longo dos níveis de
521 complexidade de habitat, e os padrões que direcionam a distribuição de espécies
522 difere entre os tipos de vegetação e condições climáticas.

523 Na restinga, a precipitação é um fator determinante na nidificação de *C.*
524 *tarsata*, que ficou concentrada na estação seca, principalmente em áreas abertas. Por
525 outro lado, em vegetação arbustiva e secundária a nidificação foi favorecida por
526 condições climáticas específicas, além de ser maior no sub-bosque. No entanto, são
527 necessários mais estudos sobre o comportamento de nidificação e exploração de
528 recursos florais na restinga.

529 Portanto, compreender a composição da comunidade de espécies de abelhas
530 e vespas solitárias é crucial para o manejo e conservação dessas espécies que
531 fornecem serviços ambientais essenciais ao funcionamento dos ecossistemas.

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